



## 26 **Highlights**

- 27 • Net global river Ba flux to the ocean estimated to be  $\sim 10$  to  $20 \text{ Gmol yr}^{-1}$
- 28 • Estuarine Ba release likely accounts for  $\sim 20$ - $75\%$  of this flux
- 29 •  $\delta^{138/134}\text{Ba}$  of net riverine inputs to ocean modified by estuarine Ba desorption
- 30 • Accounting for this process helps to balance the marine Ba isotope budget
- 31 • Adsorption to river particles fractionates Ba isotopes by  $0.2$  to  $0.3 \text{ ‰}$

32

## 33 **Abstract**

34 Barium (Ba) isotope variations offer the potential to trace environmental processes,  
35 including long-term changes in river discharge and marine export production. Riverine  
36 inputs are an important source of dissolved Ba to the ocean, which we estimate to be  $\sim 10$   
37 to  $20 \text{ Gmol yr}^{-1}$ . A large fraction ( $\sim 20$  to  $75\%$ ) of this net riverine dissolved Ba flux to the  
38 ocean is derived from estuarine processes, in particular the release of exchangeable Ba  
39 from riverine suspended particles due to exchange with major cations in seawater. Despite  
40 the importance of this process for controlling the input of dissolved Ba to the ocean, its  
41 impact on the  $\delta^{138/134}\text{Ba}$  of net riverine dissolved Ba fluxes remains unknown. To address  
42 this observational gap, Ba concentration and isotope data from across the estuarine mixing  
43 zones of the Amazon (Brazil), Fly (Papua New Guinea) and Johor (Malaysia) Rivers are  
44 presented. Desorption from suspended riverine particles releases Ba with  $\delta^{138/134}\text{Ba}$   $0.2$  to  
45  $0.3 \text{ ‰}$  lower than corresponding river dissolved loads, modifying the isotope composition  
46 of net riverine dissolved Ba fluxes that reach the ocean. This offset likely represents an  
47 isotope fractionation accompanying the adsorption of Ba by particulate phases within  
48 river catchments, which can explain the systematic enrichment of heavier Ba isotopes in  
49 river dissolved loads relative to weathering lithologies. River dissolved loads are also  
50 systematically offset to higher  $\delta^{138/134}\text{Ba}$  than the main oceanic Ba sink: burial of  $\text{BaSO}_4$

51 in marine sediment. This represents an apparent imbalance in the modern marine Ba  
52 isotope budget. Our results suggest that accounting for modification of the  $\delta^{138/134}\text{Ba}$  of  
53 net riverine Ba fluxes to the ocean by estuarine processes is likely to play a key role  
54 towards balancing the modern marine Ba isotope budget.

55

## 56 **1. Introduction**

57

58 The marine barium (Ba) cycle forms the basis of numerous approaches to trace  
59 environmental processes, including river discharge to the ocean (e.g. Fleitmann et al.,  
60 2007, Gebregiorgis et al., 2016) and past changes in marine export production (e.g.  
61 Paytan and Griffiths, 2007). Rivers are considered to be the dominant source of dissolved  
62 Ba to the ocean (Wolgemuth & Broecker, 1970, Paytan and Kastner, 1996), with estuarine  
63 processes playing a key role in controlling net riverine dissolved Ba fluxes to coastal  
64 waters (e.g. Coffey et al., 1997). A fraction of the Ba leached from mineral phases during  
65 chemical weathering is adsorbed onto the surfaces of secondary minerals within river  
66 catchments (Nesbitt et al., 1980, Price et al., 1991, Zhang et al., 2007, Gong et al., 2019,  
67 Gou et al., 2020). Upon the transport of this material through estuaries, this adsorbed Ba  
68 is released back into solution due to exchange with major cations in seawater (e.g. Hanor  
69 and Chan, 1977). This significantly enhances the dissolved Ba flux to coastal waters  
70 within most estuaries and creates interesting links between chemical weathering and  
71 physical erosion for controlling net riverine dissolved Ba fluxes to the ocean (e.g.  
72 McCulloch et al., 2003).

73 Stable Ba isotope variations are a new tool for studying the global biogeochemical  
74 cycle of this element. Initial observations indicate that river dissolved loads are enriched  
75 in heavier Ba isotopes relative to the major oceanic Ba sink, burial of  $\text{BaSO}_4$  in marine

76 sediment (Crockford et al., 2019, Gou and Deng, 2019, Gou et al., 2020, Cao et al., 2020,  
77 Tieman et al., 2020). Marine Ba inputs from submarine groundwater discharge (Mayfield  
78 et al., 2021) and hydrothermal vents (Hsieh et al., 2021) are also enriched in heavier  
79 isotopes relative to this sink, representing an apparent imbalance in the modern marine Ba  
80 isotope budget (Horner and Crockford, 2021). Available data also show that river  
81 dissolved loads are generally enriched in heavier Ba isotopes relative to rocks (Nan et al.,  
82 2018, Gou and Deng, 2019, Gou et al., 2020, Cao et al., 2020, Tieman et al., 2020,  
83 Charbonnier et al. 2020). This difference is suggested to reflect the preferential adsorption  
84 of lighter Ba isotopes to mineral surfaces within soils (Gong et al., 2019, 2020) and  
85 riverine suspended sediment loads (Gou et al., 2020). Other processes such as Ba uptake  
86 into terrestrial biomass (Bullen and Chadwick, 2016) or secondary minerals (Gou et al.,  
87 2020), or isotope fractionation during leaching of Ba from primary minerals (Gong et al.,  
88 2020), could also play important roles in controlling the isotope composition of riverine  
89 Ba fluxes. These initial observations highlight the potential of Ba isotopes to trace these  
90 processes (e.g. Charbonnier et al., 2020), but improved understanding of the controls on  
91 riverine Ba isotope fluxes are still needed.

92         The cycling of Ba isotopes in estuaries represents a key observational gap in our  
93 understanding of the global Ba isotope cycle. Given the postulated isotope fractionation  
94 accompanying Ba adsorption to mineral phases within river catchments (Gou et al., 2020,  
95 Gong et al., 2019, 2020), the release of this Ba in estuaries due to cation exchange (e.g.  
96 Hanor and Chan, 1977) is likely to influence the isotope composition of net riverine Ba  
97 inputs to the ocean. Studying the cycling of Ba isotopes in estuaries is therefore required  
98 to help balance the modern marine Ba isotope budget and may provide insights into the  
99 role of adsorption for controlling the Ba isotope composition of riverine dissolved loads.  
100 To this end, this study presents dissolved Ba concentration and isotope data for transects

101 across the estuarine mixing zones of the Amazon, Fly and Johor Rivers. These estuaries  
102 were selected based on the availability of water samples obtained during previous  
103 sampling campaigns (Swarzenski et al., 2004, Chen et al., 2015). Unfortunately  
104 particulate samples were not available for these sample sets, nor were time-series samples  
105 to assess seasonal variations in net riverine Ba fluxes. Nevertheless, these sample sets are  
106 well suited to address the aims of this study, which are to gain first-order insights into role  
107 of particulate Ba release in estuaries for controlling the net riverine flux of Ba isotopes to  
108 the ocean.

109

## 110 **2. Study area and samples**

111

112 The Amazon River provides the largest freshwater ( $6300 \text{ km}^3 \text{ yr}^{-1}$ ) and sediment  
113 ( $1200 \text{ Mt yr}^{-1}$ ) inputs to the ocean of any river, representing  $\sim 18\%$  and  $6\%$  of the global  
114 fluxes respectively (Milliman & Farnsworth, 2011). Ten surface water samples spanning a  
115 salinity range of 0.32 to 35.84 were analyzed for Ba concentration and isotope  
116 composition (Fig. 1a). These samples were collected along a transect following the  
117 northward transport of the Amazon freshwater plume, from the Amazon mouth and along  
118 the continental shelf towards Barbados, in March, 1996 as described by Swarzenski et al.  
119 (2004). Samples were filtered ( $0.45 \mu\text{m}$ ) prior to acidification by addition of distilled HCl.

120 The Fly River discharges into the Gulf of Papua, New Guinea with mean annual  
121 average freshwater and sediment discharge rates of  $180 \text{ km}^3 \text{ yr}^{-1}$  and  $110 \text{ Mt yr}^{-1}$ ,  
122 respectively (Milliman and Farnsworth, 2011). It has a very high sediment yield ( $1.5 \text{ Mt}$   
123  $\text{km}^{-2} \text{ yr}^{-1}$ ) and high average suspended sediment concentrations of  $611 \text{ mg L}^{-1}$  (Milliman  
124 and Farnsworth, 2011). Six surface water samples spanning a salinity range of 2 to 35  
125 were measured for Ba concentration and isotope composition (Fig. 1b). These samples

126 were collected across the estuarine mixing zone of the Fly River in September 1997, as  
127 described by Swarzenski et al. (2004). Samples were filtered (0.45  $\mu\text{m}$ ) prior to  
128 acidification by addition of distilled HCl.

129 The Johor River discharges into the Johor and Singapore Straits with average  
130 freshwater and sediment discharge rates of 2.4  $\text{km}^3 \text{yr}^{-1}$  and 1.2 to 2.9  $\text{Mt yr}^{-1}$  respectively  
131 (Milliman and Farnsworth, 2011, van Maren et al., 2014). Eight surface water samples  
132 spanning a salinity range of 0 to 27 along the Johor River estuary were measured for Ba  
133 concentration and isotopes composition (Fig. 1cd). Water samples were collected using  
134 trace-metal cleaned bottles, transported in dark, cool conditions back to the laboratory  
135 where they were filtered (0.45  $\mu\text{m}$  Nuclepore polycarbonate filters) and acidified  
136 following trace metal clean procedures (Chen et al., 2015).

137

### 138 **3. Analytical techniques**

139

140 The Ba concentrations and isotope compositions of waters were determined using  
141 thermal ionization mass spectrometry (TIMS; TRITON instrument, Thermo Scientific),  
142 and a previously described double spike technique to correct for instrumental mass bias  
143 (Hsieh & Henderson, 2017). Sample aliquots of 50 ml (Amazon and Fly samples) and 2  
144 to 30 ml (Johor samples) were accurately weighed and equilibrated with a known quantity  
145 of a  $^{135}\text{Ba}$ - $^{137}\text{Ba}$  double spike solution. Three of the spiked Johor samples (salinities 0, 8  
146 and 9), and one of the Amazon samples (salinity 0.32) were evaporated to dryness and  
147 refluxed in 1 ml 15M  $\text{HNO}_3$  to digest any organic material. The mass of total dissolved  
148 material in the remainder of the spiked sample aliquots required the pre-concentration of  
149 Ba by co-precipitation with  $\text{CaCO}_3$  (Foster et al., 2004). To achieve this, samples with  
150 salinities  $<20$  were first partially evaporated to increase their salinities to about 30, while

151 samples with salinities >20 were subject co-precipitation directly, induced by the addition  
152 of 1 ml 0.9 M Na<sub>2</sub>CO<sub>3</sub> solution per 16 ml of sample. The Na<sub>2</sub>CO<sub>3</sub> solution was pre-  
153 cleaned by inducing CaCO<sub>3</sub> precipitation as previously described (Hsieh and Henderson,  
154 2017). The CaCO<sub>3</sub> precipitates were separated by centrifugation, and washed in 15 ml  
155 MilliQ water. The dried sample residues and the CaCO<sub>3</sub> precipitates were dissolved in 1  
156 ml 3M HCl and processed through a cation exchange procedure (Supplementary  
157 Information). The eluted Ba fractions were dried and any organics that were leached from  
158 the cation exchange resin were oxidized prior to TIMS analyses by addition and  
159 evaporation of 15 M HNO<sub>3</sub> and 9.8 M H<sub>2</sub>O<sub>2</sub>. The procedural blanks were determined to  
160 be 0.05 – 3.72 ng of Ba (n = 4, total range), representing <1.4% of the Ba processed in the  
161 samples.

162 Barium isotope measurements using TIMS featured a double Re filament  
163 assembly following previously described protocols (Hsieh & Henderson, 2017,  
164 Bridgestock et al., 2019). Briefly, during each analysis, ion beams at atomic masses 134  
165 (Ba), 135 (Ba), 136 (Ba), 137 (Ba), 138 (Ba), 139 (La) and 140 (Ce) were monitored  
166 simultaneously using Faraday cups equipped with 10<sup>11</sup> Ω resistors. Barium isotope  
167 compositions are expressed as δ<sup>138/134</sup>Ba values (eqn. 1) relative to the standard reference  
168 material NIST 3104a:

169

$$170 \delta^{138/134}\text{Ba } \text{‰} = \left( \frac{{}^{138}\text{Ba}/{}^{134}\text{Ba}_{\text{sample}}}{{}^{138}\text{Ba}/{}^{134}\text{Ba}_{\text{NIST3104a}}} - 1 \right) \times 1000 \quad (1)$$

171

172 Repeat measurements of NIST 3104a over a period of 12 months, run at ion beam  
173 intensities of 6 to 10 V on <sup>138</sup>Ba, comparable to the majority of sample analyses, yielded a  
174 reproducibility of ± 0.03 ‰ (2SD, n = 12). This is taken to represent the level of  
175 uncertainty for Ba isotope analyses for the majority of samples. Analyses of one sample

176 from the Amazon plume (at salinity 34.58) featured lower ion beam intensities, with  
177 uncertainty is taken as the internal reproducibility ( $\pm 0.04$  ‰ 2SE) of the individual  
178 analysis. To validate the accuracy of the isotope analyses, a secondary inter-laboratory  
179 BaSO<sub>4</sub> standard, NBS-127, was analyzed during every measurement session, and this  
180 yielded  $\delta^{138/134}\text{Ba} = -0.29 \pm 0.02$  ‰ (mean  $\pm$  2SD, n = 14), in agreement with published  
181 values ( $\delta^{138/134}\text{Ba} = -0.27 \pm 0.02$  ‰; Horner et al., 2017, and  $-0.29 \pm 0.01$  ‰; Crockford et  
182 al., 2019).

183 Barium concentrations were calculated from the isotopic analyses by isotope  
184 dilution. For the Amazon and Johor River estuary samples, Ba concentrations were also  
185 determined by inductively coupled plasma mass spectrometry and agree within 12 % of  
186 those determined by TIMS (Table 1) (Supplementary Information). Barium  
187 concentrations for the Fly River samples have been previously published (Swarzenski et  
188 al., 2004). These values agree within 8 to 31% of those determined in this study from the  
189 TIMS analyses, with the exception of the sample at salinity 4 for which the Ba  
190 concentration presented in this study is 54% higher. The reason for this disagreement is  
191 unknown. We consider the Ba concentrations obtained by isotope dilution from the TIMS  
192 analyses to be more accurate, so where available these results are presented and discussed  
193 in the following.

194

#### 195 **4. Results**

196

197 The distribution of Ba concentrations in the Amazon freshwater plume as a  
198 function of salinity are in agreement with previously published results (Boyle, 1976) (Fig.  
199 2, Table 1). The lowest salinity sample (0.32), which is taken to reflect that of Amazon  
200 River water, features a Ba concentration of  $124 \text{ nmol kg}^{-1}$  consistent with published



201 values (129 to 150 nmol kg<sup>-1</sup>; Boyle, 1976, Gaillardet et al., 1997, Cao et al., 2020).  
202 Barium concentrations increase with salinity to a maximum of 243 nmol kg<sup>-1</sup> at salinity  
203 5.6, before decreasing linearly to 37 nmol kg<sup>-1</sup> at salinity 35.84 (Boyle, 1976, this study).  
204 Amazon River water features a  $\delta^{138/134}\text{Ba}$  of 0.15 ‰, which agrees within analytical  
205 uncertainty with that determined by Cao et al. (2020), of  $0.14 \pm 0.02$  ‰. At salinity 16.31,  
206  $\delta^{138/134}\text{Ba}$  decreases to 0.04 ‰, before increasing to 0.54 ‰ at salinity 35.4, which is in  
207 agreement with published data for North Atlantic surface waters ( $\delta^{138/134}\text{Ba} = 0.46$  to 0.57  
208 ‰; Bates et al., 2017, Hsieh & Henderson, 2017, Hemsing et al., 2018). These samples  
209 between salinities 16.31 and 35.84 display a linear correlation between  $\delta^{138/134}\text{Ba}$  and  
210  $1/\text{Ba}$  concentrations ( $r^2 = 0.96$ ; Fig. 2c).

211 In the Fly River estuary, Ba concentrations increase from 143 nmol kg<sup>-1</sup> at salinity  
212 0 to a maximum of 509 nmol kg<sup>-1</sup> at salinity 4 (Fig. 3, Table 1; Swarzenski et al., 2004,  
213 this study). Between salinity 4 and 35, Ba concentrations display a negative linear  
214 correlation with salinity ( $r^2 = 0.92$ ), decreasing to 63 nmol kg<sup>-1</sup> at salinity 35. The sample  
215 at salinity 2 features a  $\delta^{138/134}\text{Ba}$  value of 0.19 ‰, decreasing to relatively constant  
216  $\delta^{138/134}\text{Ba}$  values of about 0.1 ‰ between salinities 4 and 25, before increasing to 0.33 ‰  
217 at salinity 35. Samples between salinity 4 and 25 display a linear correlation between  
218  $\delta^{138/134}\text{Ba}$  and  $1/\text{Ba}$  concentrations ( $r^2 = 0.88$ ; Fig. 3c). No Ba isotope data are available  
219 for the river water (salinity 0) endmember of this estuary.

220 In the Johor River estuary, Ba concentrations increase from 352 nmol kg<sup>-1</sup> at  
221 salinity 0, to 838 nmol kg<sup>-1</sup> at salinity 8 (Fig.4, Table 1). Between salinities 8 and 27, Ba  
222 concentrations are negatively correlated with salinity ( $r^2 = 0.91$ ), with the exception of the  
223 sample from a separate tributary, which lies below this relationship and is excluded from  
224 the regression. Likewise the  $\delta^{138/134}\text{Ba}$  values decrease from 0.19 ‰ at salinity 0 to -0.02  
225 ‰ at the Ba concentration maximum (salinity 8), before systematically increasing to 0.34

226 ‰ at salinity 27. Samples between salinities 8 and 27 display a linear correlation between  
227  $\delta^{138/134}\text{Ba}$  and 1/Ba concentration ( $r^2 = 0.75$ ; Fig. 4c).

228

## 229 **5. Discussion**

230

### 231 *5.1 Estuarine processes modify the $\delta^{138/134}\text{Ba}$ of net riverine fluxes to the ocean*

232

233         The determined dissolved Ba concentration distributions in the Amazon, Fly and  
234 Johor river estuaries (Fig. 2, 3 and 4) are typical of those observed in the majority of  
235 previously studied estuaries (Boyle, 1976, Hanor and Chan, 1977, Edmond et al., 1978, Li  
236 and Chan, 1979, Carroll et al., 1993, Swarzenski et al., 2004, Coffey et al., 1997, Guay  
237 and Falkner, 1998, Stecher and Kogut, 1999, Nozaki et al., 2001, Joung and Shiller, 2014,  
238 Samanta and Dalai, 2016, Hong et al., 2018). Dissolved Ba concentrations increase to  
239 maximum values at mid salinities (~4 to 10) due to the addition of dissolved Ba by  
240 estuarine processes. This is predominantly due to particulate Ba desorption (e.g. Hanor  
241 and Chan, 1977), although in some estuaries benthic inputs may be important (Moore et  
242 al, 1997, Colbert & McManus, 2005, Joung & Shiller, 2014, Hong et al., 2018). In higher  
243 salinity regions of estuaries, linear relationships between dissolved Ba and salinity are  
244 indicative of conservative mixing with seawater, as observed in the results presented here.  
245 Notably, the higher salinity regions of the studied estuaries are also characterized by  
246 linear relationships between  $\delta^{138/134}\text{Ba}$  and 1/Ba concentration, which are consistent with  
247 conservative mixing with respect to  $\delta^{138/134}\text{Ba}$ . Extrapolation of these mixing relationships  
248 to salinity 0 provides an estimate of ‘effective’ river endmember values (Fig. 2, 3 and 4)  
249 (Supplementary Information), which is the theoretical Ba concentration or  $\delta^{138/134}\text{Ba}$  value  
250 of a parcel of river water after modification by estuarine processes. The quantity and

251 isotope composition of Ba added to the dissolved flux due to estuarine processes can be  
252 estimated through comparison of the effective river Ba concentration and  $\delta^{138/134}\text{Ba}$  value  
253 to those measured for the riverine dissolved load (eqn. 2 and 3).

254

$$255 \quad [\text{Ba}]_{\text{added}} = [\text{Ba}]_{\text{effective}} - [\text{Ba}]_{\text{river}} \quad (\text{eqn. 2})$$

256

$$257 \quad \delta^{138/134}\text{Ba}_{\text{added}} = 1/f_{\text{added}} \times [\delta^{138/134}\text{Ba}_{\text{effective}} - (\delta^{138/134}\text{Ba}_{\text{river}} \times f_{\text{river}})] \quad (\text{eqn. 3})$$

258

259 With:

$$260 \quad f_{\text{added}} = [\text{Ba}]_{\text{added}}/[\text{Ba}]_{\text{effective}} \quad (\text{eqn. 4})$$

$$261 \quad f_{\text{river}} = [\text{Ba}]_{\text{river}}/[\text{Ba}]_{\text{effective}} \quad (\text{eqn. 5})$$

262

263 The subscripts ‘effective’, ‘river’ and ‘added’ refer to the Ba concentration ([Ba]) and  
264  $\delta^{138/134}\text{Ba}$  values of the effective river endmember, riverine dissolved load and the Ba  
265 added by estuarine processes (per kg of river water) respectively. For the Fly River  
266 estuary, measurement of the  $\delta^{138/134}\text{Ba}$  of the riverine dissolved load is not available.  
267 Instead  $\delta^{138/134}\text{Ba}_{\text{added}}$  and  $\delta^{138/134}\text{Ba}_{\text{river}}$  are estimated using mass balance calculations  
268 based on the data for the sample at salinity 2 and the effective river endmember  
269 (Supplementary Information).

270 Extrapolation of the conservative mixing data for each river are used to calculate  
271 estimated effective river Ba concentrations and  $\delta^{138/134}\text{Ba}$  values, along with the estimated  
272 amount and  $\delta^{138/134}\text{Ba}$  value of dissolved Ba added by processes in the Amazon, Fly and  
273 Johor river estuaries (Table 2). These calculations reveal that estuarine processes lower  
274 the  $\delta^{138/134}\text{Ba}$  of net riverine fluxes to the ocean (i.e. effective river endmembers) by 0.13  
275 to 0.23 ‰ compared to those of riverine dissolved loads (Fig. 2, 3 and 4), due to the

276 addition of Ba featuring  $\delta^{138/134}\text{Ba}$  between 0.20 to 0.32 ‰ lower than the corresponding  
277 dissolved loads.

278         Next the potential importance of desorption from suspended riverine particles  
279 versus benthic inputs for controlling the addition of isotopically light Ba to the dissolved  
280 phase in the Amazon, Fly and Johor river estuaries are assessed. The Amazon and Fly  
281 rivers, respectively, have average suspended sediment concentrations of 191 mg L<sup>-1</sup> and  
282 611 mg L<sup>-1</sup> (Milliman and Farnsworth, 2011), with concentrations up to 1500 mg L<sup>-1</sup> and  
283 1000 mg L<sup>-1</sup> measured in the 3 to 8 salinity zones of these estuaries at the time of sample  
284 collection (Swarzenski et al., 2004). At these suspended sediment concentrations,  
285 exchangeable particulate Ba concentrations of 0.1 to 0.8 nmol mg<sup>-1</sup> and of 0.3 to 0.5 nmol  
286 mg<sup>-1</sup> would be required to supply the amount of Ba added in the Amazon and Fly River  
287 estuaries, respectively (Table 2). These values are within the range of exchangeable Ba  
288 concentrations (0.12 and 1.7 nmol mg<sup>-1</sup> of sediment) determined for suspended particles  
289 from other rivers (Yellow, Humber, Hooghly, and Hudson; Li and Chan, 1979, Coffey et  
290 al., 1997, Samanta and Dalai, 2016, Gou et al., 2020). Particulate Ba desorption can  
291 therefore feasibly account for the observed Ba addition in the Amazon and Fly River  
292 estuaries.

293         Suspended sediment concentrations in the Johor River are poorly constrained at  
294 ~150 mg L<sup>-1</sup> (average), possibly ranging up to ~1000 mg L<sup>-1</sup> during high discharge events  
295 (van Maren et al., 2014). These suspended sediment concentrations require exchangeable  
296 particulate Ba concentrations of 0.7 to 4.6 nmol mg<sup>-1</sup> to account for the observed estuarine  
297 Ba addition, of 688 nmol kg<sup>-1</sup>. This range overlaps but exceeds exchangeable Ba  
298 concentrations measured for riverine particles (0.12 to 1.7 nmol mg<sup>-1</sup>; Li and Chan, 1979,  
299 Coffey et al., 1997, Samanta and Dalai, 2016, Gou et al., 2020). In this case, significant

300 benthic inputs cannot be ruled out, but are not necessarily required to explain the  
301 magnitude of dissolved Ba addition in the Johor River estuary.

302

### 303 *5.2 Insights into the controls on the $\delta^{138/134}\text{Ba}$ of river dissolved loads*

304

305 Previous studies have noted that river dissolved loads are offset to higher  
306  $\delta^{138/134}\text{Ba}$  values than the weathering lithologies (Gong et al., 2019, Gou et al., 2020,  
307 Charbonnier et al., 2020). Compiled  $\delta^{138/134}\text{Ba}$  data for the dissolved loads of 12 different  
308 rivers are systematically higher than the average  $\delta^{138/134}\text{Ba}$  of rocks further supporting this  
309 observation (Fig. 5) (Miyazaki et al., 2014, Nan et al., 2015, van Zuilen et al., 2016,  
310 Bullen and Chadwick, 2016, Bridgestock et al., 2018, Nielsen et al., 2018, Nan et al.,  
311 2018, Gou et al., 2020, Cao et al., 2020, Nielsen et al., 2020, An et al., 2020, Lin et al.,  
312 2020, Tieman et al., 2020, this study). The available  $\delta^{138/134}\text{Ba}$  for rocks display a  
313 significant total range of -0.63 to 0.37 ‰, however the majority of rocks analyzed so far  
314 feature  $\delta^{138/134}\text{Ba} \approx 0$  ‰ (Fig. 5), as exemplified by the recent estimate of average upper  
315 continental crust  $\delta^{138/134}\text{Ba}$  of  $0.00 \pm 0.04$  ‰ (Nan et al., 2018). Lithological  $\delta^{138/134}\text{Ba}$   
316 variability therefore cannot explain the systematic offset of river dissolved loads to higher  
317  $\delta^{138/134}\text{Ba}$  values, and is unlikely to significantly control river dissolved load  $\delta^{138/134}\text{Ba}$   
318 variability, particularly in larger river catchments where integration towards the average  
319  $\delta^{138/134}\text{Ba}$  of rocks might be expected. The systematic offset of riverine dissolved loads to  
320 higher  $\delta^{138/134}\text{Ba}$  relative to average rock values therefore requires Ba isotope  
321 fractionation during processes operating within the critical zone.

322 Recent studies have argued that preferential adsorption of lighter Ba isotopes  
323 within soils and suspended riverine sediments can explain the offset between the  
324  $\delta^{138/134}\text{Ba}$  of river dissolved loads and those of the weathering lithologies (Gong et al.,

2019, 2020, Gou et al., 2020). The magnitude of this isotope fractionation however remains poorly constrained. Isotope fractionations accompanying additional processes such as the leaching of Ba from primary minerals (Gong et al., 2020), Ba uptake into biomass (Bullen and Chadwick, 2016) and secondary mineral precipitation (Gou et al., 2020) may also influence river dissolved load  $\delta^{138/134}\text{Ba}$ . In particular,  $\delta^{138/134}\text{Ba}$  data for river waters from the Amazon River basin were recently interpreted as being controlled by a combination of isotope fractionations during Ba incorporation into secondary minerals, and uptake into terrestrial biomass (Charbonnier et al., 2020).

Our results indicate that Ba desorbed from riverine suspended particles within estuaries features lower  $\delta^{138/134}\text{Ba}$  than corresponding river dissolved loads (Fig. 2, 3 and 4, Table 2). Desorption of exchangeable particulate Ba from suspended sediment in response to changes in salinity in estuaries is thought to be quantitative (Li et al., 1984, Samanta & Dalai, 2016). The isotope composition calculated for the addition flux ( $\delta^{138/134}\text{Ba} = -0.11 \pm 0.13 \text{‰}$ ,  $+0.02 \pm 0.15 \text{‰}$  and  $-0.01 \pm 0.19 \text{‰}$  for the Amazon, Fly, Johor Rivers respectively; Table 2) is therefore likely to represent that of exchangeable Ba transported on riverine particles. The offset between these compositions and those of the corresponding river dissolved loads likely reflect isotope fractionations during the adsorption of Ba onto particle surfaces, of  $\Delta^{138/134}\text{Ba}_{\text{dissolved-adsorbed}} = -0.26 \pm 0.14 \text{‰}$ ,  $-0.32 \pm 0.28 \text{‰}$  and  $-0.20 \pm 0.19 \text{‰}$  for the Amazon, Fly and Johor rivers respectively (where  $\Delta^{138/134}\text{Ba}_{\text{dissolved-adsorbed}} = \delta^{138/134}\text{Ba}_{\text{river}} - \delta^{138/134}\text{Ba}_{\text{added}}$ ; Table 2). It is interesting to note the consistency of these estimated isotope fractionation factors between these three river systems, although the propagated uncertainties on these values are large. In detail, the magnitudes of isotope fractionations accompanying Ba adsorption in river catchments are likely to depend on numerous factors, including the composition of suspended particulate loads as well as river water pH, ionic strength and major ion compositions, requiring

350 further study through direct measurements of exchangeable particulate Ba isotope  
351 compositions.

352 In contrast to riverine dissolved loads, the effective river  $\delta^{138/134}\text{Ba}$  values for the  
353 Amazon, Fly and Johor are comparable to the upper continental crust/average rock  
354  $\delta^{138/134}\text{Ba}$  (Fig. 6). Hsieh and Henderson (2017) also report an effective river  $\delta^{138/134}\text{Ba}$  of  
355  $0.03 \pm 0.09 \text{ ‰}$  for the combined discharge of the Paraná and Uruguay rivers from the Rio  
356 de la Plata estuary ( $670 \text{ km}^3 \text{ yr}^{-1}$ ; Milliman and Farnsworth, 2011) also in agreement with  
357 the upper continental crust. In other words the combined dissolved and adsorbed Ba pools  
358 transported in these rivers are in isotopic balance with the expected average composition  
359 weathering lithologies in these catchments (Fig. 6). This supports the idea that isotopic  
360 fractionation accompanying Ba adsorption onto particle surfaces is an important control  
361 on riverine dissolved load  $\delta^{138/134}\text{Ba}$  variations (Gong et al., 2019, 2020, Gou et al., 2020).  
362 It implies that additional processes such as Ba uptake into biomass and secondary  
363 minerals, or incongruent weathering of primary minerals do not significantly influence  
364 dissolved load  $\delta^{138/134}\text{Ba}$  of the studied rivers, or that effects of such processes happen to  
365 cancel each other out.

366 A recent study discounted Ba adsorption as an important process for explaining  
367 the systematic offset of river dissolved loads in the Amazon River basin to higher  
368  $\delta^{138/134}\text{Ba}$  weathering lithology, instead favoring isotope fractionations accompanying Ba  
369 uptake into secondary minerals and terrestrial organic matter as the dominant control  
370 (Charbonnier et al., 2021). Our results for the Amazon River estuary however suggest that  
371 Ba adsorption to suspended sediments can explain this offset, at least at the mouth of this  
372 river, at the time of sampling. Coupled measurements of dissolved and adsorbed Ba pools  
373 transported by rivers will be required to further unravel the processes responsible for

374 controlling river dissolved load  $\delta^{138/134}\text{Ba}$ , and hence the development of this novel  
375 isotope system trace critical zone processes.

376

### 377 *5.3 Towards balancing the marine barium isotope budget*

378

#### 379 *5.3.1 Estimating the global riverine Ba flux to the ocean*

380

381 Rivers have previously been considered to be the dominant source of Ba to the  
382 ocean, with additional inputs from submarine groundwater discharge and hydrothermal  
383 vents (Wolgemuth and Broecker, 1970, Paytan and Kastner, 1996, Shaw et al., 1998,  
384 Dickens et al., 2003). Wolgemuth and Broecker (1970) and Gaillardet et al. (2003)  
385 estimate global riverine dissolved Ba fluxes of  $10.5 \text{ Gmol yr}^{-1}$  and  $6.3 \text{ Gmol yr}^{-1}$   
386 respectively. These flux estimates however, are based on only riverine dissolved load data  
387 and do not account for the release of dissolved Ba within estuaries. Using the effective  
388 river Ba concentration reported for the Congo River (Edmond et al., 1978), Paytan and  
389 Kastner (1996) estimated a global riverine dissolved Ba flux to the ocean of  $10.6 \text{ Gmol yr}^{-1}$ .  
390 This latter estimate accounts for estuarine release of dissolved Ba, but relies upon a  
391 single value for the Congo River to be globally representative.

392 To improve on previous global riverine Ba flux estimates, literature data for 47  
393 estuarine dissolved Ba concentration distributions from 27 different rivers have been  
394 compiled (Table 3, Supplementary Information, Electronic Data). From each of these  
395 estuarine dissolved Ba concentration distributions we calculated the Ba concentration of  
396 the river dissolved load ( $[\text{Ba}]_{\text{river}}$ ) and the effective river ( $[\text{Ba}]_{\text{effective}}$ ) endmembers as  
397 described in section 5.1 and the Supplementary Information file. A detailed description



398 and analysis of this compiled dataset is given in the Supplement Information file, and is  
399 summarized in the following.

400 The majority of the estuarine Ba distributions (35 out of the 38 distributions for  
401 which there are constraints on  $[\text{Ba}]_{\text{river}}$ ) yield  $[\text{Ba}]_{\text{effective}}$  that are significantly higher than  
402  $[\text{Ba}]_{\text{river}}$ . This highlights the global importance of estuarine processes for increasing net  
403 riverine Ba fluxes to the ocean. Desorption of Ba from suspended particles is interpreted  
404 to be the dominant process for driving this estuarine Ba addition (e.g. Hanor and Chan,  
405 1977, Edmond et al., 1987, Li and Chan, 1979), however, inputs from submarine  
406 groundwater discharge or diagenetic release from estuarine bottom sediments have also  
407 been invoked to explain some of these estuarine Ba distributions (Frolich et al., 1985,  
408 Moore, 1997, Moore and Shaw, 2008, Joung and Shiller, 2014, Hong et al., 2018). The  
409 magnitude of estuarine Ba addition (calculated following eqn. 2) for the rivers in the  
410 compilation increases with suspended sediment concentration (Fig. 7). Furthermore, the  
411 magnitude of estuarine Ba addition for the majority of the rivers can be explained by  
412 levels of Ba release per mass of sediment that are in range of measured riverine suspended  
413 sediment exchangeable Ba concentrations (0.12 to 1.7 nmol mg<sup>-1</sup>; Li and Chan, 1979,  
414 Coffey et al., 1997, Samanta and Dalai, 2016, Gou et al., 2020) (Fig. 7). This further  
415 supports the global importance Ba desorption from suspended river sediments in  
416 estuaries, and highlights sediment discharge as a key control on net riverine Ba fluxes to  
417 the ocean.

418 The combined discharge of the 27 rivers in the literature compilation is 13400 km<sup>3</sup>  
419 yr<sup>-1</sup>, representing 37% of the global freshwater flux to the ocean of 36000 km<sup>3</sup> yr<sup>-1</sup> (Table  
420 3) (Milliman and Farnsworth, 2011). The discharge weighed average  $[\text{Ba}]_{\text{river}}$  and  
421  $[\text{Ba}]_{\text{effective}}$  of the 27 rivers in the literature compilation are 155 and 282 nmol kg<sup>-1</sup>  
422 respectively. Combined with a global riverine freshwater flux to the ocean of 36000 km<sup>3</sup>

423 yr<sup>-1</sup> (Milliman and Farnsworth, 2011), yields a global net riverine Ba input to the ocean of  
424 10.1 Gmol yr<sup>-1</sup> (based on [Ba]<sub>effective</sub>), of which 5.6 Gmol yr<sup>-1</sup> is transported by river  
425 dissolved loads (based on [Ba]<sub>river</sub>). There are two major sources of uncertainty on these  
426 estimates. Firstly, due to temporal variations in [Ba]<sub>river</sub> and [Ba]<sub>effective</sub> for individual  
427 rivers, which are likely on the order of ± 50% based on the rivers for which multiple  
428 estuarine Ba distributions spanning different time periods are available in the compiled  
429 dataset (Supplementary Information). Secondly there is uncertainty in how well these  
430 discharge weighted average [Ba]<sub>river</sub> and [Ba]<sub>effective</sub> represent the 63% of the global  
431 riverine freshwater flux discharged by rivers not included in the compiled literature data.  
432 In particular, the discharge weighted average [Ba]<sub>river</sub> and [Ba]<sub>effective</sub> are strongly biased  
433 towards those of the Amazon River, which accounts for 50% of the total discharge of the  
434 rivers in the compilation. Using the un-weighted average [Ba]<sub>river</sub> and [Ba]<sub>effective</sub> of the  
435 rivers in the data compilation of 296 and 387 nmol kg<sup>-1</sup> respectively, may therefore better  
436 represent globally averaged values. These values yield a global net riverine Ba input to  
437 the ocean of 13.9 Gmol yr<sup>-1</sup> (based on [Ba]<sub>effective</sub>) of which 10.7 Gmol yr<sup>-1</sup> transported by  
438 river dissolved loads (based on [Ba]<sub>river</sub>).

439         A further bias, impacting the both of the flux estimates above, is that the average  
440 suspended sediment concentration of the rivers in the data compilation (of 253 mg L<sup>-1</sup>) is  
441 about half the global average (528 mg L<sup>-1</sup>; Milliman and Farnsworth, 2011). This bias  
442 reflects the importance of small (< 10000 km<sup>2</sup>), mountainous rivers catchments for  
443 supplying up to 45% of the global sediment flux (Milliman and Farnsworth, 2011), which  
444 are poorly represented in the compilation. Given the role of sediment discharge for  
445 driving estuarine Ba release via desorption from suspended sediments (Fig. 7), global  
446 riverine Ba flux estimates based on the compiled data therefore likely underestimate the  
447 true value. The compiled literature data suggest typical estuarine desorption of between

448 0.3 and 0.7 nmol of Ba per mg of suspended river sediment (Fig. 7). Coupling these  
449 values with a global average riverine suspended sediment concentration of 528 mg L<sup>-1</sup>,  
450 and global riverine freshwater discharge of 36000 km<sup>3</sup> yr<sup>-1</sup> (Milliman and Farnsworth,  
451 2011), yields a global estuarine Ba release of 5.7 to 13.3 Gmol yr<sup>-1</sup>. Combined with the  
452 estimates of river dissolved load global Ba fluxes above (5.6 and 10.7 Gmol yr<sup>-1</sup>) yields  
453 estimates of global net riverine Ba fluxes to the ocean of between 11.3 and 24 Gmol yr<sup>-1</sup>.

454         Based on the above discussion we suggest that the global net riverine Ba flux to  
455 the ocean is likely ~10 to 20 Gmol yr<sup>-1</sup>, with between ~20 to 75 % of this flux derived  
456 from estuarine processes. These estimates remain subject to large uncertainty, but  
457 represent a significant advancement on previous estimates. Improving on these flux  
458 estimates requires more data for estuarine Ba distributions, particularly for rivers with  
459 large suspended sediment loads, which likely play an important role in supplying Ba to  
460 the ocean via desorption from suspended riverine sediment (Fig. 7).

461

### 462 *5.3.2. The importance of riverine marine Ba inputs compared to other marine Ba sources*

463

464         In comparison to our global net riverine dissolved Ba flux estimate of ~10 to 20  
465 Gmol yr<sup>-1</sup>, the global dissolved Ba flux by submarine groundwater discharge (SGD) has  
466 recently been estimated to be between 0.4 and 3.6 Gmol yr<sup>-1</sup> (Mayfield et al., 2021). This  
467 estimate however only accounts for the freshwater component of the global marine Ba  
468 input from SGD. Re-circulation of seawater through coastal sediments represents another  
469 important component of SGD chemical fluxes to the ocean (Swarzenski, 2007), which has  
470 been shown to provide Ba inputs that exceed riverine Ba inputs in certain coastal regions  
471 (Shaw et al., 1998, Moore and Shaw, 2008). The global importance of this saline  
472 component of the SGD Ba input to the ocean however remains unconstrained. The Ba in

473 this saline component of SGD is derived from diagenetic reactions in coastal sediment,  
474 rather than from terrestrial chemical weathering. It is therefore unclear as to how much of  
475 this source truly represents an input of ‘new’ Ba to the ocean, rather than the recycling of  
476 Ba recently removed from seawater.

477         Dissolved Ba inputs from hydrothermal vents have previously been estimated at  
478 2.40 to 3.35 Gmol yr<sup>-1</sup> (Paytan and Kastner, 1996, Dickens et al., 2003), based on the  
479 extremely high Ba concentrations of hydrothermal fluids (>10 μmol kg<sup>-1</sup>; Von Damm et  
480 al., 1985). The vast majority of this Ba is however known to precipitate as BaSO<sub>4</sub> during  
481 mixing of hydrothermal fluids with deep ocean waters (e.g. Jamieson et al., 2016),  
482 significantly reducing the magnitude of this input. There is considerable uncertainty  
483 regarding the extent to which this process reduces the hydrothermal Ba input, but it is  
484 likely that ‘effective’ Ba concentration of these hydrothermal inputs are close to deep  
485 ocean concentrations (<200 nmol kg<sup>-1</sup>; Hsieh et al., 2021). Coupled with a hydrothermal  
486 water flux of  $3 \times 10^{13}$  kg yr<sup>-1</sup> (Elderfield and Schultz, 1996) results in hydrothermal Ba  
487 fluxes several orders of magnitude lower than previous estimates (<0.006 Gmol yr<sup>-1</sup>),  
488 which are negligible compared to the estimated global net riverine Ba flux (~10 to 20  
489 Gmol yr<sup>-1</sup>).

490         In summary, the flux estimates discussed above support previous assertions that  
491 rivers are the dominant input of Ba to the ocean. We however caution that all of these flux  
492 estimates have a high degree of uncertainty, and that the importance of Ba inputs from the  
493 recirculated seawater component of SGD (e.g. Moore and Shaw, 2008) needs to be further  
494 developed as it could be an important component of the marine Ba budget.

495

496 *5.3.3. Constraints on the marine Ba isotope budget*

497

498           The main output of Ba from the ocean is through the precipitation of BaSO<sub>4</sub> and its  
499 burial in marine sediment (e.g. Paytan and Kastner, 1996, Dickens et al., 2003). The  
500 currently available  $\delta^{138/134}\text{Ba}$  data for riverine dissolved loads are systematically higher  
501 than BaSO<sub>4</sub> in modern ocean sediment of  $0.04 \pm 0.06$  ‰ (mean  $\pm$  2sd, n = 61; Crockford  
502 et al., 2019) (Fig. 5, 6). Furthermore, recent constraints on the  $\delta^{138/134}\text{Ba}$  of the freshwater  
503 component of submarine groundwater discharge and hydrothermal inputs, of  $0.12 \pm 0.03$   
504 ‰ and  $1.7 \pm 0.07$  ‰, respectively are also systematically higher than this output (Hsieh et  
505 al., 2021, Mayfield et al., 2021). This suggests that either the marine Ba isotope budget is  
506 currently not in steady state, or that there are important additional sources featuring low  
507  $\delta^{138/134}\text{Ba}$  or sinks featuring high  $\delta^{138/134}\text{Ba}$  missing from this budget. In this regard, the  
508 findings presented in this study, that estuarine processes lower the  $\delta^{138/134}\text{Ba}$  of net river  
509 inputs to the ocean represents an important step towards balancing the marine Ba isotope  
510 budget. Specifically the Ba flux weighted ( $0.02 \pm 0.05$  ‰) and un-weighted ( $0.05 \pm 0.05$   
511 ‰) means ( $\pm$ 2sd) of the effective river  $\delta^{138/134}\text{Ba}$  for the Amazon, Rio de la Plata, Fly and  
512 Johor river systems agree with the mean  $\delta^{138/134}\text{Ba}$  of marine BaSO<sub>4</sub> (Fig. 6). These rivers  
513 represent ~20 % of the global freshwater discharge and ~9 to ~19 % of our estimate for  
514 the global net dissolved riverine Ba flux to the ocean (Table 3). It remains unclear as to  
515 how representative these four values are of the  $\delta^{138/134}\text{Ba}$  of global net dissolved riverine  
516 fluxes to the ocean, given the potential for both temporal variability within the studied  
517 river systems and variability between river systems. Nevertheless, it is clear that  
518 accounting for the release of low  $\delta^{138/134}\text{Ba}$  due to desorption from riverine suspended  
519 particles in estuaries, as documented by this study, is important to constrain the modern  
520 marine Ba isotope budget. Thus, river dissolved load  $\delta^{138/134}\text{Ba}$  data do not necessarily  
521 represent the isotope composition of the riverine Ba inputs to the ocean, and more data

522 spanning estuarine mixing zones is required to properly assess the  $\delta^{138/134}\text{Ba}$  of global net  
523 riverine Ba inputs to the ocean.

524

## 525 **6. Conclusions**

526

527 The isotope compositions of net dissolved riverine Ba fluxes to the ocean are  
528 foremost modified by particulate Ba desorption in estuaries. This process contributes Ba  
529 featuring  $\delta^{138/134}\text{Ba}$  that is about 0.2 to 0.3 ‰ lower than corresponding river dissolved  
530 loads, an offset that likely reflects an isotope fractionation during the adsorption of Ba  
531 released during chemical weathering onto mineral surfaces. This supports the idea that  
532 preferential adsorption of lighter Ba isotopes is a key process for controlling  $\delta^{138/134}\text{Ba}$   
533 values of riverine dissolved loads, as previously suggested (Gong et al., 2019, 2020, Gou  
534 et al., 2020).

535 Net dissolved riverine Ba fluxes to the ocean are a globally important marine Ba  
536 source, which we estimate to be  $\sim 10$  to  $20 \text{ Gmol yr}^{-1}$ , a significant component of which is  
537 derived from particulate Ba desorption in estuaries. River dissolved loads are  
538 systematically offset to higher  $\delta^{138/134}\text{Ba}$  than the main output of dissolved Ba from the  
539 ocean, burial of marine  $\text{BaSO}_4$ , contributing to an apparent imbalance in the marine Ba  
540 isotope budget. Our results demonstrate that accounting for estuarine processes is  
541 necessary to properly assess the  $\delta^{138/134}\text{Ba}$  of net riverine Ba fluxes to the ocean, and in  
542 doing so will likely play a key role towards balancing the modern marine Ba isotope  
543 budget.

544

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546

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556

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883 **Table 1**, Dissolved Ba concentration and  $\delta^{138/134}\text{Ba}$  results for samples from the Amazon,  
 884 Fly and Johor River estuaries.

| Latitude<br>°N              | Longitude<br>°E | Collection<br>Date | Salinity | Ba concentration<br>(nmol kg <sup>-1</sup> ) |       | $\delta^{138/134}\text{Ba}$ |
|-----------------------------|-----------------|--------------------|----------|--|-------|-----------------------------|
|                             |                 |                    |          | ICP-MS                                       | TIMS  |                             |
| <i>Amazon River Estuary</i> |                 |                    |          |  |       |                             |
| 1.17417                     | 49.09333        | 18/03/1996         | 0.32     |  | 123.6 | 0.15                        |
| 2.90389                     | 50.14167        | 07/03/1996         | 16.58    | 151.3  | 138.7 | 0.11                        |
| 1.49417                     | 47.90833        | 22/03/1996         | 16.31    | 192.9  | 194.7 | 0.04                        |
| 4.58750                     | 51.27000        | 08/03/1996         | 21.23    | 137.5  | 125.8 | 0.10                        |
| 3.94389                     | 50.60361        | 11/03/1996         | 24.80    | 123.4  | 111.3 | 0.13                        |
| 3.96917                     | 50.57889        | 12/03/1996         | 28.21    | 93.8   | 85.4  | 0.20                        |
| 2.97417                     | 49.11917        | 22/03/1996         | 32.73    | 59.7   | 55.2  | 0.44                        |
| 9.01361                     | 55.18722        | 24/03/1996         | 34.58    | 57.1   | 53.3  | 0.46                        |
| 11.45944                    | 57.88944        | 25/03/1996         | 35.65    | 48.7   | 48.0  | 0.49                        |
| 6.86833                     | 52.94389        | 23/03/1996         | 35.84    | 39.8   | 37.4  | 0.54                        |
| <i>Fly River estuary</i>    |                 |                    |          |  |       |                             |
| -8.392                      | 143.022         |                    | 2        | 237  | 257.7 | 0.19                        |
| -8.383                      | 143.117         |                    | 4        | 237  | 509.2 | 0.13                        |
| -8.392                      | 143.265         |                    | 8        | 388  | 419.7 | 0.14                        |
| -8.359                      | 143.352         |                    | 10       | 368  | 411.8 | 0.12                        |
| -8.373                      | 143.766         |                    | 24       | 162  | 188.7 | 0.11                        |
| -8.781                      | 143.985         |                    | 35       | 43   | 62.5  | 0.33                        |
| <i>Johor River estuary</i>  |                 |                    |          |  |       |                             |
| 1.72662                     | 103.89961       | 03/11/2017         | 0        | 368.6  | 351.5 | 0.19                        |
| 1.69909                     | 103.92558       | 03/11/2017         | 8        | 828.1  | 837.9 | -0.02                       |
| 1.68976                     | 103.94888       | 03/11/2017         | 9        | 444.2  | 417.0 | 0.07                        |
| 1.66540                     | 103.93163       | 03/11/2017         | 15       | 383.0  | 381.6 | 0.09                        |
| 1.63458                     | 103.97203       | 03/11/2017         | 20       | 372.3  | 351.6 | 0.19                        |
| 1.58442                     | 103.98727       | 03/11/2017         | 25       | 252.4  | 260.7 | 0.15                        |
| 1.42067                     | 104.00174       | 17/12/2015         | 27       | 83.7   | 79.2  | 0.34                        |
| 1.42050                     | 104.00240       | 16/04/2016         | 27       | 107.7  | 96.3  | 0.21                        |

885 Barium concentrations determined by inductively coupled plasma mass spectrometry (ICP  
 886 MS) for the Fly River estuary are from Swarzenski et al. (2004). Uncertainty on  $\delta^{138/134}\text{Ba}$   
 887 is  $\pm 0.03$  ‰ based on long-term reproducibility (2sd) of the standard reference material  
 888 NIST 3104a, with the exception of Amazon salinity 34.58 for which an uncertainty of  $\pm$   
 889 0.04 ‰ is assigned, representing the internal reproducibility of this analysis (2se).

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897 **Table 2**, Estimates of the effective river Ba concentrations and  $\delta^{138/134}\text{Ba}$  values, and the  
 898 amount and  $\delta^{138/134}\text{Ba}$  of dissolved Ba added by estuarine processes in the Amazon, Fly  
 899 and Johor river estuaries.

|  | <b>Amazon</b> | <b>Fly</b>   | <b>Johor</b> |
|--|---------------|--------------|--------------|
| <b>[Ba]<sub>river</sub> (nmol kg<sup>-1</sup>)</b>                           | 124           | 143          | 352          |
| <b><math>\delta^{138/134}\text{Ba}_{\text{river}}</math> ‰</b>               | 0.15 ± 0.03   | 0.34 ± 0.19  | 0.19 ± 0.03  |
| <b>[Ba]<sub>effective</sub> (nmol kg<sup>-1</sup>)</b>                       | 268 ± 46      | 512 ± 74     | 1039 ± 318   |
| <b><math>\delta^{138/134}\text{Ba}_{\text{effective}}</math> ‰</b>           | 0.01 ± 0.06   | 0.11 ± 0.06  | 0.06 ± 0.12  |
| <b>[Ba]<sub>added</sub> (nmol kg<sup>-1</sup>)</b>                           | 144 ± 48      | 369 ± 76     | 688 ± 320    |
| <b><math>\delta^{138/134}\text{Ba}_{\text{added}}</math> ‰</b>               | -0.11 ± 0.13  | 0.02 ± 0.15  | -0.01 ± 0.19 |
| <b><math>\Delta^{138/134}\text{Ba}_{\text{dissolved-adsorpted}}</math> ‰</b> | -0.26 ± 0.14  | -0.32 ± 0.28 | -0.20 ± 0.19 |

900 [Ba]<sub>river</sub> and  $\delta^{138/134}\text{Ba}_{\text{river}}$  denote the Ba concentration and  $\delta^{138/134}\text{Ba}$  value of the river  
 901 dissolved load. [Ba]<sub>effective</sub> and  $\delta^{138/134}\text{Ba}_{\text{effective}}$  denote the Ba concentration and  $\delta^{138/134}\text{Ba}$   
 902 value of the effective river endmembers, which represent the theoretical Ba concentration  
 903 and  $\delta^{138/134}\text{Ba}$  value of a parcel of river water after modification by estuarine processes.  
 904 [Ba]<sub>added</sub> and  $\delta^{138/134}\text{Ba}_{\text{added}}$  denote the amount and isotope composition added to a parcel  
 905 of river water by estuarine processes, calculated using equations 2 and 3.  $\Delta^{138/134}\text{Ba}_{\text{dissolved-}}$   
 906  $\text{adsorpted} = \delta^{138/134}\text{Ba}_{\text{river}} - \delta^{138/134}\text{Ba}_{\text{added}}$ . Uncertainties on [Ba]<sub>added</sub>,  $\delta^{138/134}\text{Ba}_{\text{added}}$  and  
 907  $\Delta^{138/134}\text{Ba}_{\text{dissolved-adsorpted}}$  were calculated by propagation of the uncertainties for [Ba]<sub>river</sub>,  
 908  $\delta^{138/134}\text{Ba}_{\text{river}}$ , [Ba]<sub>effective</sub> and  $\delta^{138/134}\text{Ba}_{\text{effective}}$  by Monte Carlo simulation (Supplementary  
 909 Information). Note that for the Fly River,  $\delta^{138/134}\text{Ba}_{\text{river}}$  and  $\delta^{138/134}\text{Ba}_{\text{added}}$  are estimated  
 910 using mass balance calculations based on values for the effective river endmember and  
 911 data for the sample at salinity 2 (Supplementary Information).

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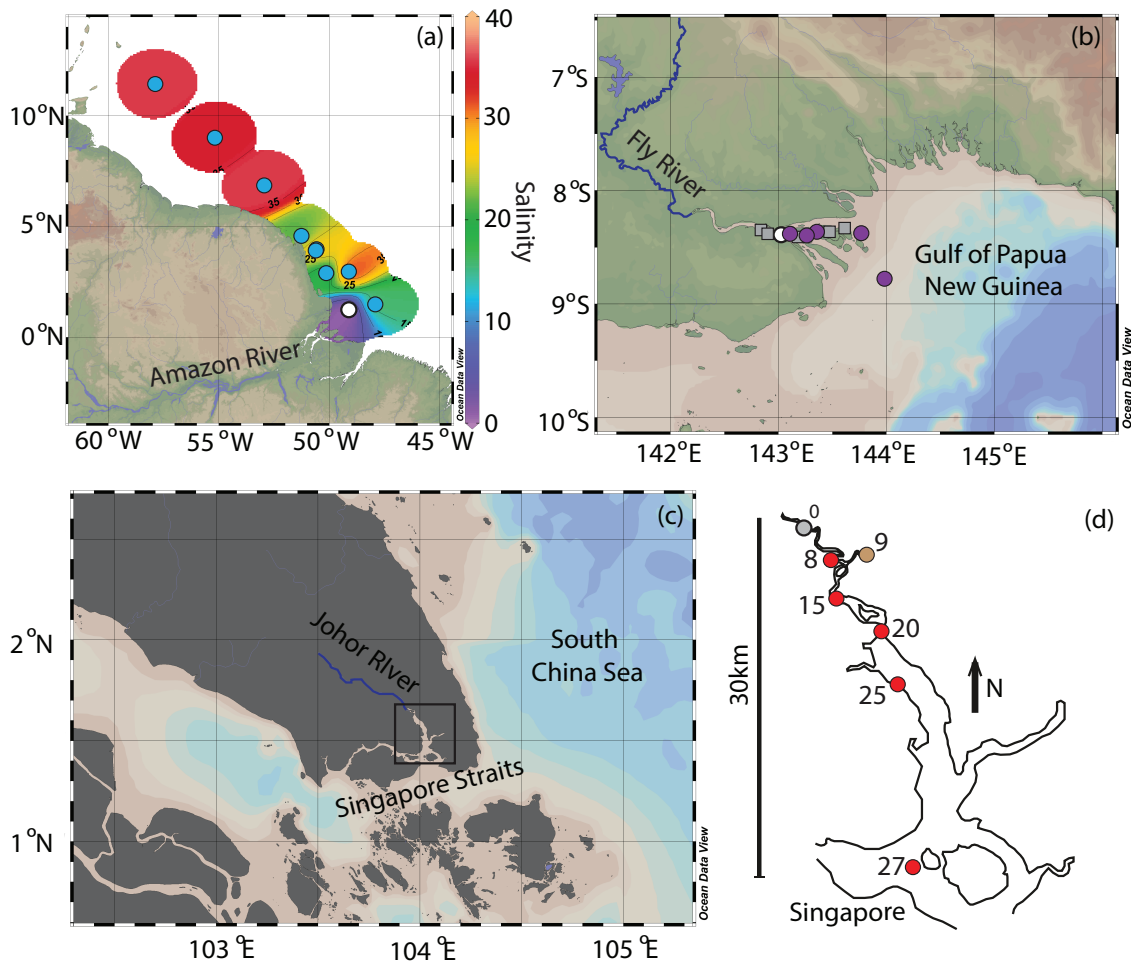
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926 **Table 3**, Literature compilation of Ba concentrations data for river dissolved loads  
 927 ( $[Ba]_{\text{river}}$ ) and effective river endmembers ( $[Ba]_{\text{effective}}$ ) used to estimate the global riverine  
 928 Ba flux to the oceans

| River                             | Water discharge<br>$\text{km}^3 \text{yr}^{-1}$ | Sediment discharge<br>$\text{Mt yr}^{-1}$ | $[Ba]_{\text{river}}$<br>$\text{nmol kg}^{-1}$ | River Ba flux<br>$\text{Gmol yr}^{-1}$ | $[Ba]_{\text{effective}}$<br>$\text{nmol kg}^{-1}$ | Effective Ba flux<br>$\text{Gmol yr}^{-1}$ |
|-----------------------------------|---|---|--|--|--|--|
| Amazon <sup>1</sup>               | 6300  | 1200                                      | 126  | 0.80                                   | 262  | 1.65                                       |
| Congo <sup>2</sup>                | 1300  | 43  | 133  | 0.18                                   | 170  | 0.22                                       |
| Ganges-Bramaputra <sup>3</sup>    | 1270  | 1060                                      | 157  | 0.20                                   | 376  | 0.48                                       |
| Yangtze <sup>4</sup>              | 900   | 470                                       | 231  | 0.21                                   | 520  | 0.47                                       |
| Paraná-Uruguay <sup>5</sup>       | 670   | 100                                       | NA   | NA                                     | 208  | 0.14                                       |
| Yenisei <sup>6</sup>              | 620   | 4.1                                       | 76   | 0.047                                  | 63   | 0.039                                      |
| Lena <sup>6</sup>                 | 520   | 20  | 74   | 0.039                                  | 134  | 0.070                                      |
| Mississippi <sup>7</sup>          | 490   | 210                                       | 467  | 0.23                                   | 607  | 0.30                                       |
| Ob <sup>6</sup>                   | 390   | 16  | 69   | 0.027                                  | 91   | 0.035                                      |
| Mackenzie <sup>8</sup>            | 310   | 100                                       | 393  | 0.12                                   | 499  | 0.16                                       |
| Pearl <sup>9</sup>                | 260   | 25  | 212  | 0.055                                  | 253  | 0.066                                      |
| Fly <sup>10</sup>                 | 180   | 110                                       | 143  | 0.026                                  | 512  | 0.092                                      |
| Indigirka <sup>6</sup>            | 55  | 11  | NA   | NA                                     | 191  | 0.010                                      |
| Chao Phraya <sup>11</sup>         | 30  | 3   | 274  | $8.2 \times 10^{-3}$                   | 416  | 0.012                                      |
| Paraíba do Sul <sup>12</sup>      | 28  | 4   | 225  | $6.3 \times 10^{-3}$                   | 275  | $7.7 \times 10^{-3}$                       |
| Hudson <sup>13</sup>              | 15  | 0.2                                       | 276  | $4.1 \times 10^{-3}$                   | 297  | $4.5 \times 10^{-3}$                       |
| Savannah <sup>14</sup>            | 11  | 1   | 130  | $1.4 \times 10^{-3}$                   | 208  | $2.3 \times 10^{-3}$                       |
| Burdekin <sup>15</sup>            | 10  | 3   | 1175   | 0.012                                  | 650  | $6.5 \times 10^{-3}$                       |
| Delaware <sup>16</sup>            | 10  | 1   | 181  | $1.8 \times 10^{-3}$                   | 229  | $2.3 \times 10^{-3}$                       |
| Pee Dee <sup>17</sup>             | 6.9   | 0.43                                      | NA   | NA                                     | 215  | $1.5 \times 10^{-3}$                       |
| Cape Fear <sup>17</sup>           | 6.5   | 0.29                                      | NA   | NA                                     | 336  | $2.4 \times 10^{-3}$                       |
| Fitzroy <sup>15</sup>             | 5.3   | 3   | 395  | $2.1 \times 10^{-3}$                   | 572  | $3 \times 10^{-3}$                         |
| Humber <sup>18</sup>              | 4.6   | 0.19                                      | 213  | $1 \times 10^{-3}$                     | 465  | $2.1 \times 10^{-3}$                       |
| St Johns <sup>17</sup>            | 2.9   | 0.27                                      | NA   | NA                                     | 133  | $3.2 \times 10^{-4}$                       |
| Johor <sup>19</sup>               | 2.4   | 0.45                                      | 352  | $1 \times 10^{-3}$                     | 1039   | $3 \times 10^{-3}$                         |
| Peace <sup>20</sup>               | 1.5   |   | 50   | $7.5 \times 10^{-5}$                   | 267  | $4 \times 10^{-4}$                         |
| Nueces <sup>21</sup>              | 0.72  | 0.71                                      | 1170   | $8.4 \times 10^{-4}$                   | 1428   | $1 \times 10^{-3}$                         |
| <b>Discharge weighted average</b> |   |   | 155  |  | 282  |  |
| <b>Un-weighted average</b>        |   |   | 287  |  | 387  |  |

929 All water and sediment discharge values are taken from Milliman and Farnsworth (2011),  
 930 with the exception of sediment discharge of the Johor River (van Maren et al., 2004).  
 931  $[Ba]_{\text{river}}$  and  $[Ba]_{\text{effective}}$  are based on estuarine Ba distributions from the following sources:  
 932 <sup>1</sup>This study and Boyle (1976), <sup>2</sup>Edmond et al. (1978), <sup>3</sup>Carroll et al. (1993), <sup>4</sup>Edmond et  
 933 al. (1985), <sup>5</sup>Hsieh and Henderson (2017), <sup>6</sup>Guay and Falkner (1998), <sup>7</sup>Hanor and Chan  
 934 (1977), Shim et al. (2012) and Joung and Shiller (2014), <sup>8</sup>Guay and Falkner (1998) and  
 935 Kipp et al. (2020), <sup>9</sup>Hong et al. (2018), <sup>10</sup>Swarzenski et al. (2004) and this study, <sup>11</sup>Nozaki  
 936 et al. (2001), <sup>12</sup>Souza et al., (2010), <sup>13</sup>Li and Chan, (1979), <sup>14</sup>Shaw et al. (1998) and  
 937 Moore and Shaw (2008), <sup>15</sup>Lewis et al. (2018), <sup>16</sup>Coffey et al. (1997) and Stecher and  
 938 Kogut (1999), <sup>17</sup>Moore and Shaw (2008), <sup>18</sup>Coffey et al., (1997), <sup>19</sup>this study, <sup>20</sup>Frolich et  
 939 al. (1985), <sup>21</sup>Walter and Nims (2015). NA- not available.





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941 **Figure 1**, Locations of samples from the Amazon, Fly and Johor River estuaries. Panel (a)

942 and (b) display locations of samples from the Amazon freshwater plume, and the Fly

943 River estuaries respectively (Swarzenski et al., 2004). In panel (b) the purple and open

944 circles denote samples analyzed in this study for Ba concentrations and isotope

945 compositions, while the grey squares denote samples measured for Ba concentrations only

946 by Swarzenski et al. (2004). Panels (c) and (d) display the location of the Johor River

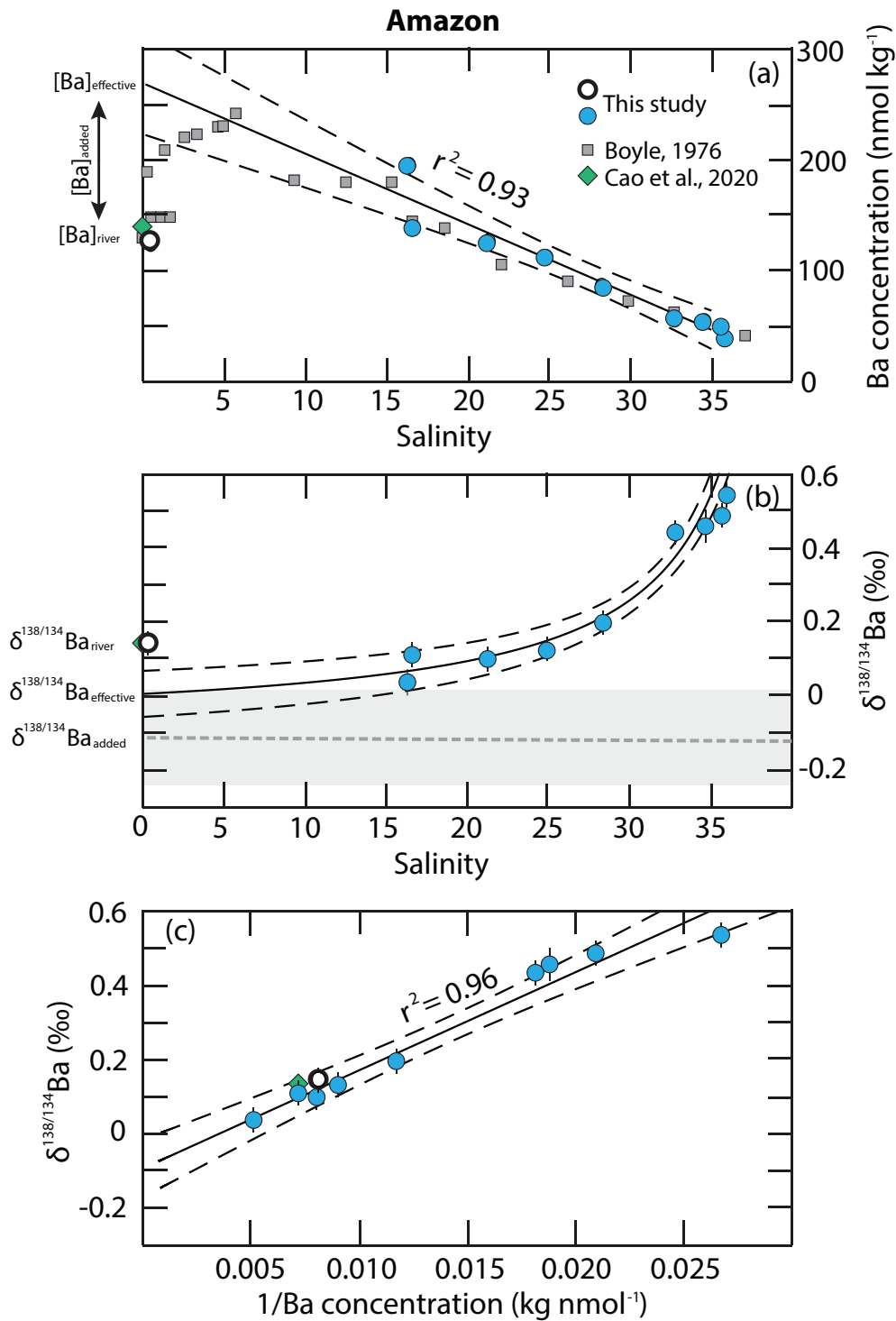
947 estuary, and sampling locations (circles) within the Johor River estuary respectively

948 (Chen et al., 2015). In panel (d), the grey, brown and red circle(s) denote the sampling

949 locations of Johor River water, a separate tributary, and estuarine waters respectively, and

950 the numbers denote the salinities at the different sampling locations.

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952

953 **Figure 2**, Relationships between salinity, Ba concentrations and  $\delta^{138/134}\text{Ba}$  values in the

954 Amazon River estuary. Solid black lines represent conservative mixing relationships,

955 defined by linear regressions between Ba concentration and salinity (panel a) and

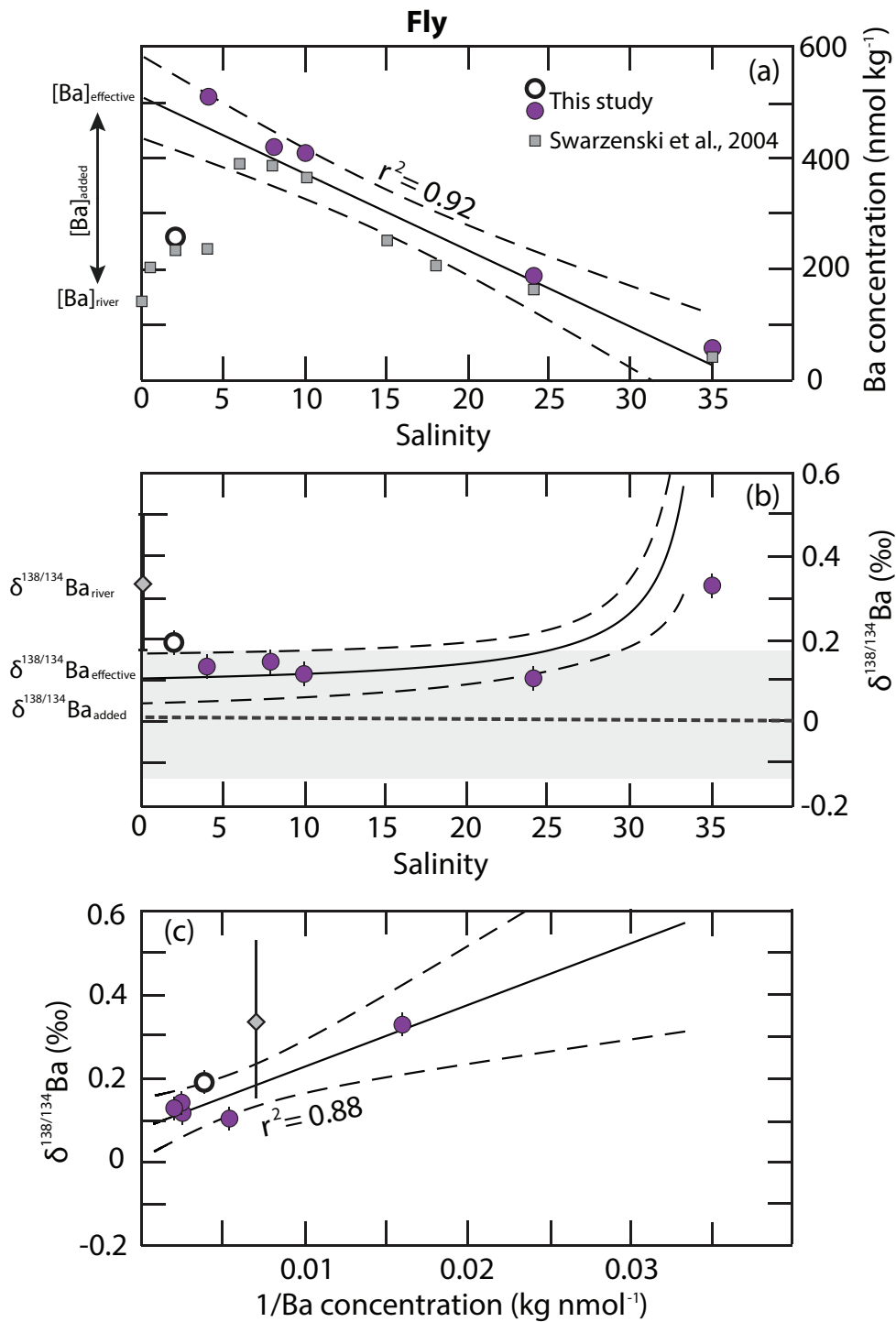
956  $\delta^{138/134}\text{Ba}$  and  $1/\text{Ba}$  concentration (panel c) with dashed lines showing 95% confidence

957 intervals. The mixing relationship in panel (b) is defined by combining the linear

958 regressions in panels (a) and (c). Open circle denotes the Amazon River water  
959 endmember, which is omitted from regressions to define mixing relationships. Literature  
960 data are displayed by grey squares (Boyle, 1976) and the green diamond (Cao et al.,  
961 2020). In panel (b) the horizontal grey dashed lines show the  $\delta^{138/134}\text{Ba}$  value estimated  
962 for the Ba added to the dissolved riverine flux by estuarine processes, with uncertainties  
963 shown by the grey shaded intervals.

964

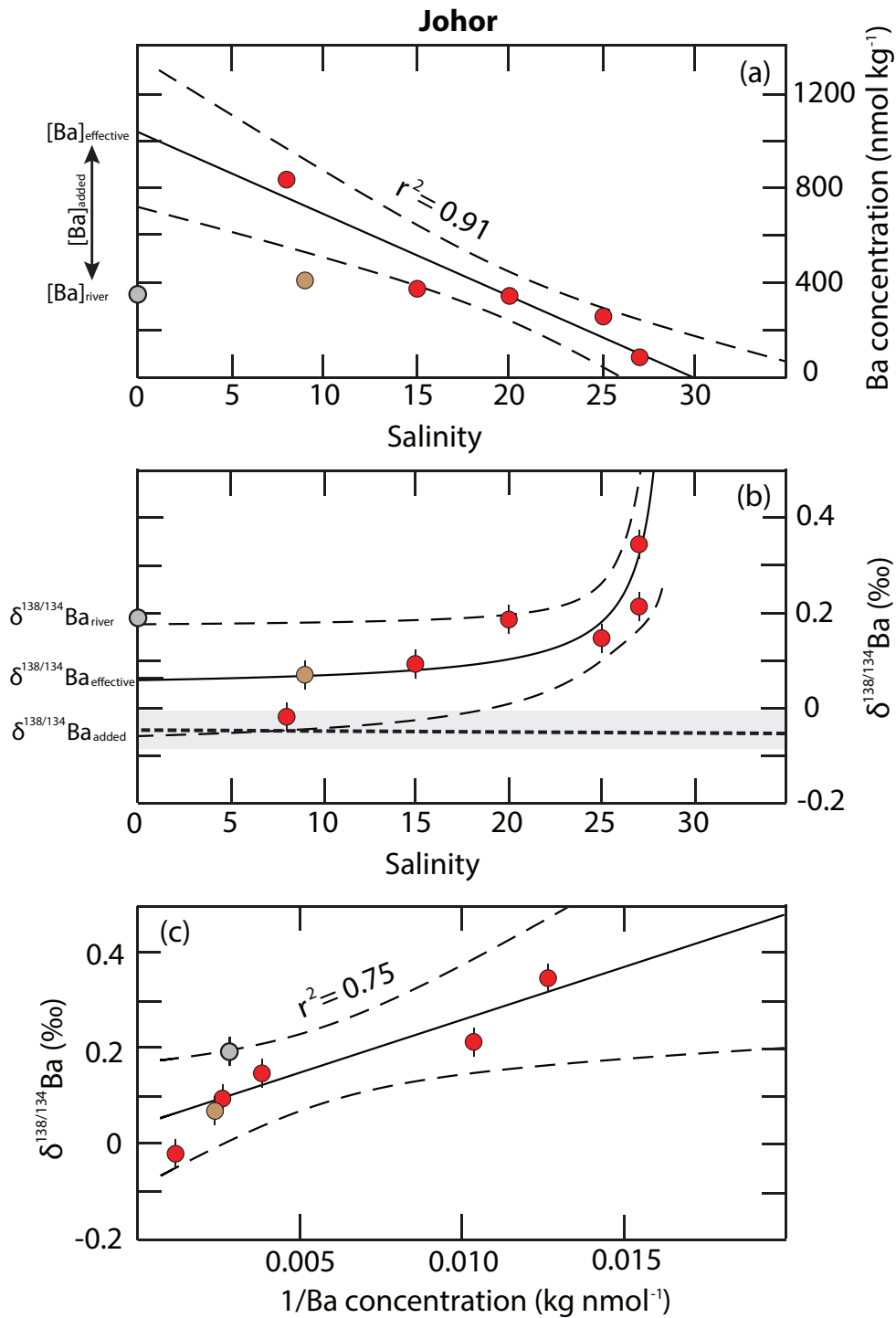
965



966

967 **Figure 3**, Relationships between salinity, Ba concentrations and  $\delta^{138/134}\text{Ba}$  values in the  
 968 Fly River estuary. Solid black lines represent conservative mixing relationships, defined  
 969 by linear regressions between Ba concentration and salinity (panel a) and  $\delta^{138/134}\text{Ba}$  and  
 970  $1/\text{Ba concentration}$  (panel c) with dashed lines showing 95% confidence intervals. The  
 971 mixing relationship in panel (b) is defined by combining the linear regressions in panels

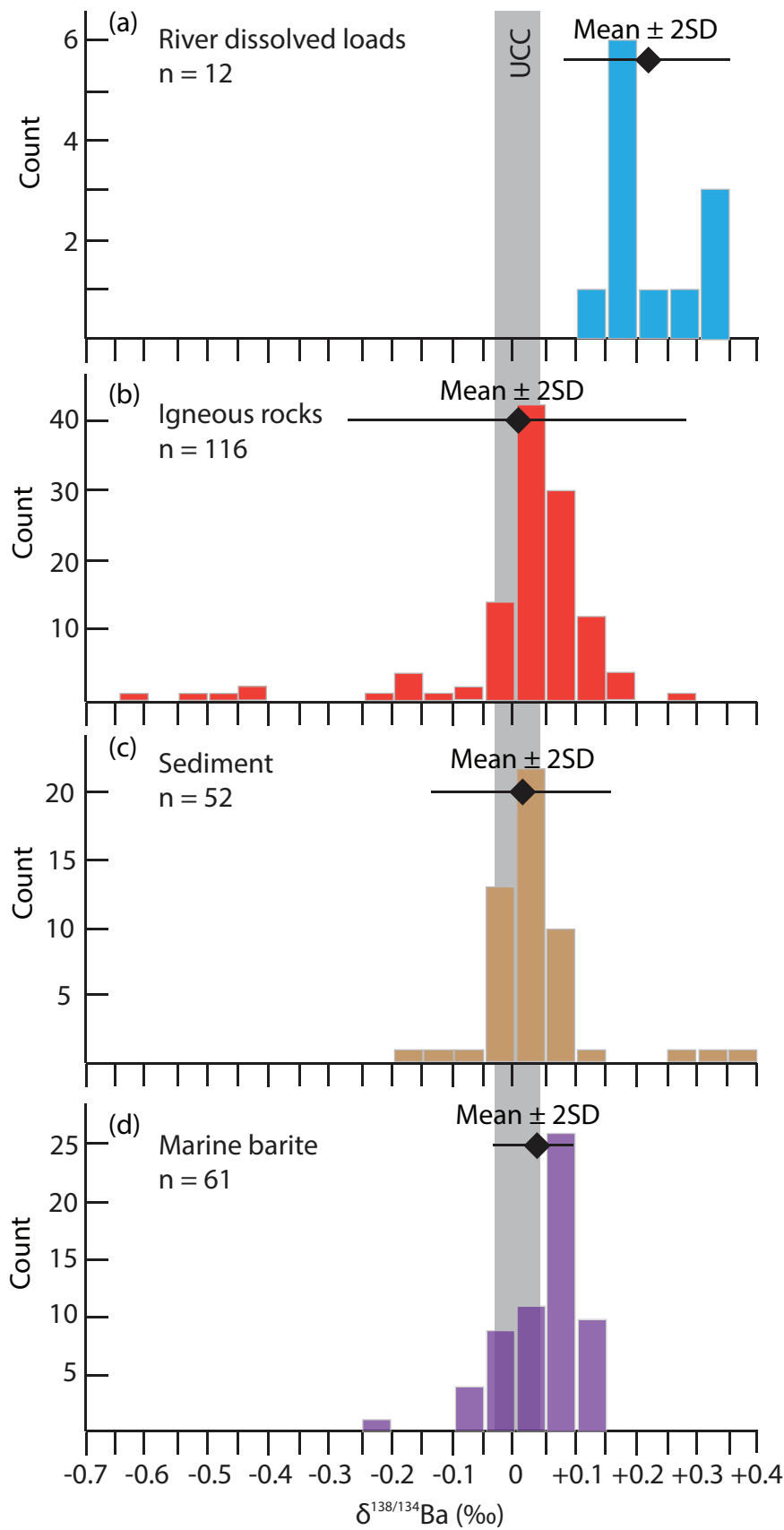
972 (a) and (c). Open circle denotes the sample at salinity 2, which is omitted from regressions  
973 to define mixing relationships. Grey squares display previously published Ba  
974 concentration data for these samples (Swarzenski et al., 2004). The grey diamond denotes  
975 the  $\delta^{138/134}\text{Ba}$  value estimated for the Fly River dissolved load (Supplementary  
976 Information). In panel (b) the horizontal grey dashed lines show the  $\delta^{138/134}\text{Ba}$  value  
977 estimated for the Ba added to the dissolved riverine flux by estuarine processes, with  
978 uncertainties shown by the grey shaded intervals.  
979



980

981 **Figure 4**, Relationships between salinity, Ba concentrations and  $\delta^{138/134}\text{Ba}$  values in the  
 982 Johor River estuary. Solid black lines represent conservative mixing relationships, defined  
 983 by linear regressions between Ba concentration and salinity (panel a) and  $\delta^{138/134}\text{Ba}$  and  
 984  $1/\text{Ba}$  concentration (panel c) with dashed lines showing 95% confidence intervals. The  
 985 mixing relationship in panel (b) is defined by combining the linear regressions in panels

986 (a) and (c). The grey circle and brown circle denote the Johor River endmember and a  
987 sample from the separate tributary to the Johor estuary, which are omitted from  
988 regressions to define the mixing relationships. In panel (b) the horizontal grey dashed  
989 lines show the  $\delta^{138/134}\text{Ba}$  value estimated for the Ba added to the dissolved riverine flux by  
990 estuarine processes, with uncertainties shown by the grey shaded intervals.  
991



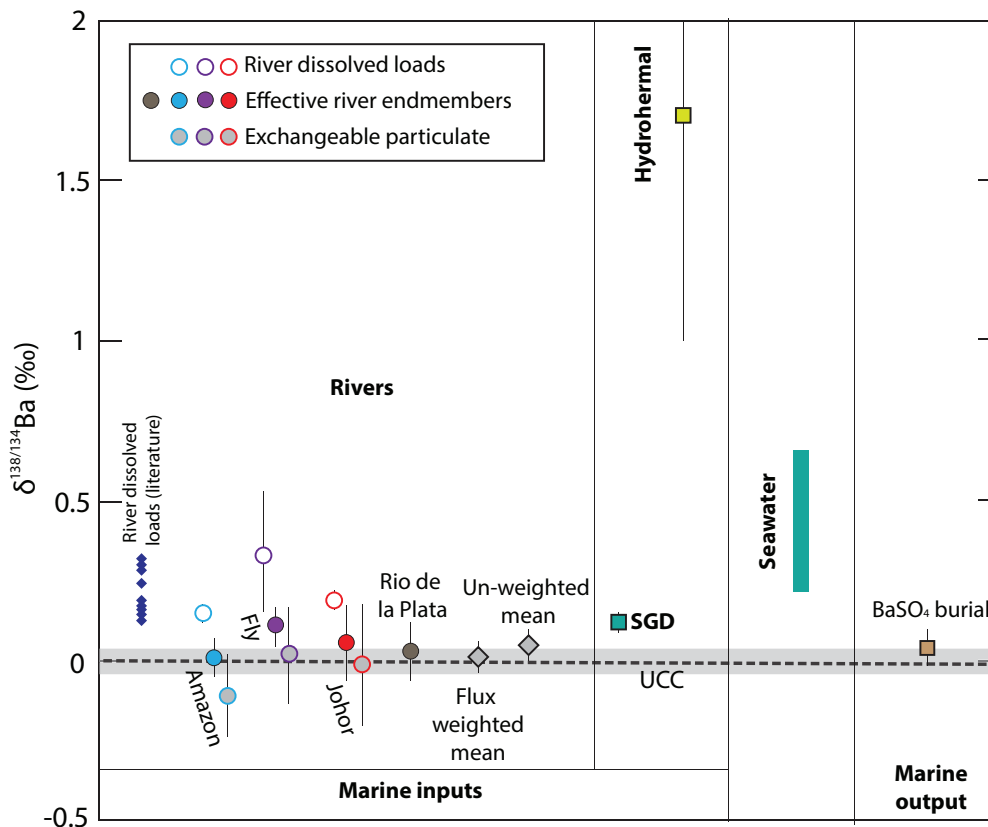
992

993 **Figure 5,** Histograms of the  $\delta^{138/134}\text{Ba}$  of river dissolved loads, igneous and sedimentary

994 rocks, and marine barite. River dissolved load data for the rivers Yangtze, Amazon,

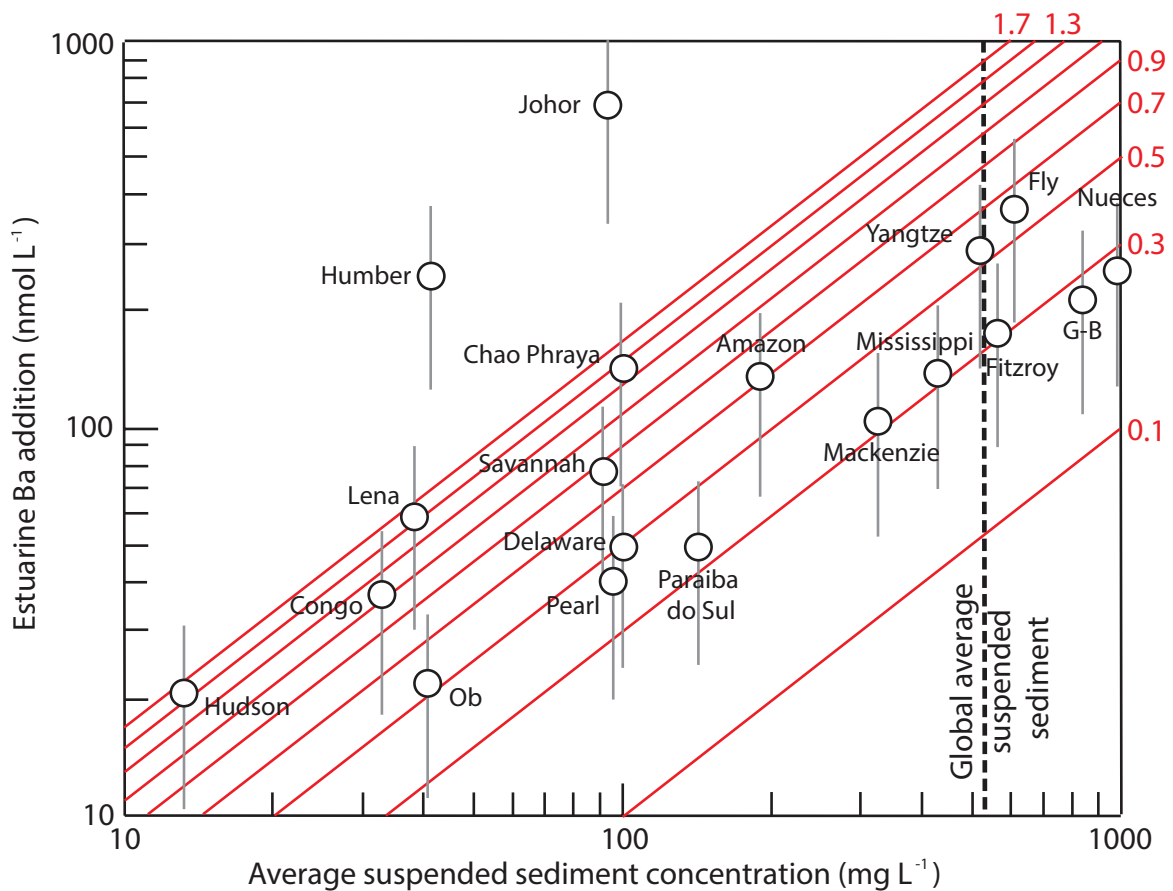


995 Yukon, Pearl, Sepik, Danube, Lena, Colorado, Ohio, Yellow, Fly and Johor are from Gou  
 996 et al. (2020), Cao et al. (2020), Tieman et al. (2020) and this study. For the Yellow River  
 997 timeseries  $\delta^{138/134}\text{Ba}$  measurements are available (Gou et al., 2020), and the Ba flux  
 998 weighted average  $\delta^{138/134}\text{Ba}$  of 0.28 ‰ is used. Igneous and sedimentary rock data are  
 999 from Miyazaki et al. (2014), Nan et al. (2015), van Zuilen et al. (2016), Bullen and  
 1000 Chadwick (2016,) Bridgestock et al. (2018), Nielsen et al. (2018), Nan et al. (2018),  
 1001 Nielsen et al., (2020), An et al. (2020) and Lin et al. (2020). Data for marine  $\text{BaSO}_4$  are  
 1002 from Crockford et al. (2019). The grey bar represents the average composition of the  
 1003 upper continental crust (UCC) given by Nan et al. (2018).  
 1004



1005  
 1006 **Figure 6**, Constraints on the marine Ba isotope budget. Grey diamonds display the Ba  
 1007 flux weighted and un-weighted mean of effective river  $\delta^{138/134}\text{Ba}$  values, representing net  
 1008 riverine inputs to the ocean. The horizontal dashed line and grey shaded area shows the

1009  $\delta^{138/134}\text{Ba}$  of the upper continental crust (UCC; Nan et al., 2018). Blue diamonds display  
 1010 literature data for dissolved loads of the Yangtze, Amazon, Yukon, Pearl, Sepik, Danube,  
 1011 Lena, Colorado, Ohio and Yellow Rivers (Gou et al., 2020, Cao et al., 2020, Tieman et  
 1012 al., 2020). The effective river  $\delta^{138/134}\text{Ba}$  of the Rio de la Plata estuary is from Hsieh and  
 1013 Henderson (2017). Constraints on the average  $\delta^{138/134}\text{Ba}$  of freshwater submarine  
 1014 groundwater (SGD) and hydrothermal inputs (following modification by  $\text{BaSO}_4$   
 1015 precipitation) are from Mayfield et al. (2021) and Hsieh et al. (2021) respectively.  
 1016 Seawater data are from Horner et al. (2015), Bates et al. (2017), Hsieh and Henderson  
 1017 (2017), Bridgestock et al. (2018), Hemsing et al. (2018) and Geymann et al. (2019). The  
 1018 average  $\delta^{138/134}\text{Ba}$  of modern marine barite ( $\pm 2\text{sd}$ ) is from Crockford et al (2019).  
 1019



1020  
 1021 **Figure 7**, Magnitude of estuarine Ba addition ( $[\text{Ba}]_{\text{added}}$ ) versus annual average suspended  
 1022 sediment concentrations for global rivers.  $[\text{Ba}]_{\text{added}}$  is calculated following eqn. 2 from

1023 estuarine Ba distributions compiled from published literature (Table 3, Supplementary  
1024 Information, Electronic Data). For rivers for which multiple estuarine Ba distributions  
1025 exist, averaged  $[Ba]_{\text{river}}$  and  $[Ba]_{\text{effective}}$  are used to calculate  $[Ba]_{\text{added}}$ . A conservative  
1026 relative uncertainty of  $\pm 50\%$  is assigned to  $[Ba]_{\text{added}}$  to reflect the potential scale of  
1027 temporal variations in  $[Ba]_{\text{river}}$  and  $[Ba]_{\text{effective}}$  for individual rivers (Supplementary  
1028 Information). Red contours show Ba release per mass of sediment ( $\text{nmol mg}^{-1}$ ), and span  
1029 the range of measured exchangeable Ba concentrations for riverine suspended sediments  
1030 ( $0.12$  to  $1.7 \text{ nmol mg}^{-1}$ ; Li and Chan, 1979, Coffey et al., 1997, Samanta and Dalai, 2016,  
1031 Gou et al., 2020). Annual average suspended concentrations are from Milliman and  
1032 Farnsworth (2011). The dashed vertical line shows the average global river suspended  
1033 sediment concentration, of  $528 \text{ mg L}^{-1}$  (Milliman and Farnsworth, 2011). G-B denotes the  
1034 Ganges-Brahmaputra river system. Rivers for which constraints on  $[Ba]_{\text{river}}$  are not  
1035 available in the compiled literature estuarine Ba distributions are not included in this  
1036 figure. The rivers Burdekin and Yenisey display negative  $[Ba]_{\text{added}}$  values and are not  
1037 included in this plot, nor is the Peace River due to lack of suspended sediment data.  
1038  
1039