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4 **Low and annually variable migratory connectivity in a long distance migrant:**

5 **Whinchats *Saxicola rubetra* may show a bet-hedging strategy**

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7 EMMA BLACKBURN^{1,2}, MALCOLM BURGESS³, BENEDICTUS FREEMAN^{2,6}, ALICE RISELY⁴, ARIN

8 IZANG², SAM IVANDE^{1,2}, CHRIS HEWSON⁵ & WILL CRESSWELL^{1*,2}

9

10 ¹*Centre for Biological Diversity, University of St Andrews, Harold Mitchell Building, St Andrews, Fife*

11 *KY16 9TH, UK*

12 ²*AP Leventis Ornithological Research Institute, Jos, Nigeria*

13 ³*RSPB Centre for Conservation Science, The Lodge, Sandy, Beds SG19 2DL, UK*

14 ⁴*Centre for Integrative Ecology, Deakin University, Geelong, VIC 3216, Australia*

15 ⁵*British Trust for Ornithology, The Nunnery, Thetford, IP24 2PU UK*

16 ⁶*Fauna & Flora International (FFI), 2 Harmon's Compound, Oldest Congo Town, Monrovia, Liberia*

17 *Correspondence author: wrlc@st-and.ac.uk

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19 **Keywords:** climate change, long-distance migrant, migrant declines, migratory spread, Palearctic,

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22 Running head: Low and variable connectivity in Whinchats

23 The spatial scale of non-breeding areas used by long distance migrant animals can vary from specific,
24 relatively small non-breeding areas for each independent breeding population (high connectivity) to a
25 distribution over a large non-breeding area with mixing of breeding populations (low connectivity).
26 Measuring variation in the degree of connectivity and how it arises is crucial to predict how migratory
27 animals can respond to global habitat and climate change because low connectivity is likely an
28 adaptation to environmental uncertainty. Here, we test whether use of non-breeding areas in a long
29 distance migrant may be stochastic by measuring the degree of connectivity, and whether it is
30 annually variable. 29 wintering Whinchats tagged with geolocators over two years within 40 km² in
31 central Nigeria were found breeding over 2.549 million km² (26% of the land area of Europe), without
32 an asymptote being approached in the relationship between area and sample size. Ranges differed in
33 size between years by 1.506 million km² and only 15% of the total breeding range across both years
34 overlapped (8% overlap between years when only first year birds were considered), well above the
35 range size difference and below the proportion of overlap that would be predicted from two equivalent
36 groups breeding at random locations within the observed range. Mean distance between breeding
37 locations (i.e. migratory spread) differed significantly between years (2013, 604 ± 18 km; 2014, 869 ±
38 33 km). The results showed very low and variable connectivity that was reasonably robust to the
39 errors and assumptions inherent in the use of geolocators, but with the caveat of only two years'
40 ranges to compare, and the sensitivity of range to the breeding locations of a small number of
41 individuals. However, if representative, the results suggest the scope for between-year variation
42 (cohort effects) to determine migrant distribution on a large scale. Furthermore, for species with
43 similar low connectivity, we would predict breeding population trends to reflect average conditions
44 across large non-breeding areas: thus, as large areas of Africa become subject to habitat loss,
45 migrant populations throughout Europe will decline.

46 Understanding how breeding and non-breeding sites are connected (migratory connectivity) is crucial
47 for successfully predicting the response of migratory populations to environmental change (Bauer et
48 al. 2016, Webster et al. 2002), particularly in the light of the widespread decline of migratory animals
49 and increasing habitat loss and degradation (Gilroy et al. 2016, Flockhart et al. 2015, Costa et al.
50 2012, Vickery et al. 2014). Indeed, strategies to address migratory declines are often based on the
51 idea that migrant populations will have connected breeding and non-breeding (or wintering) ranges
52 and that the identification of these ranges is a priority (Runge et al. 2014, Runge et al. 2015, Martin et
53 al. 2007). However, the scale of connectivity within and between populations of migrants and how this
54 scale arises is still poorly known (Bauer et al. 2016).

55 The spatial scale of use of non-breeding areas by long distance migrant animals can vary from high
56 connectivity - specific non-breeding areas for each independent breeding population – to low
57 connectivity - a spread over a large non-breeding area and so mixing across breeding populations
58 (Webster et al. 2002, Newton 2008). High connectivity is thought to arise from deterministic, targeted
59 migration which allows population specialisation, but which reduces the resilience of a breeding
60 population to changes in the distribution of non-breeding habitat, and so should only be selected for
61 when the location and suitability of non-breeding sites are highly predictable and stable (Reilly &
62 Reilly 2009, Cresswell 2014). In contrast, low connectivity is thought to arise from bet-hedging -
63 untargeted migration which allows populations to deal with unpredictable environmental conditions
64 (Reilly & Reilly 2009, Cresswell 2014, see also Botero et al. 2015). Breeding populations which are
65 distributed over a large area during the non-breeding season should have greater potential to
66 encounter suitable habitat over a larger range, even as this shifts due to climate change (Cresswell
67 2014). Furthermore, breeding populations that mix during the non-breeding season are likely to
68 respond synchronously to changing conditions outside of the breeding season, whereas highly
69 connected populations will respond independently (Esler 2000).

70 Long-distance migration of birds between northern Europe and sub-Saharan Africa is a recently
71 evolved adaptive response to dynamic global climatic conditions (Cresswell et al. 2011, Fryxell & Holt
72 2013), suggesting that a bet-hedging strategy of individuals from the same brood migrating to
73 geographically separate non-breeding sites might actually be the norm because of the resilience this
74 strategy grants against dynamic conditions (Botero et al. 2015, Reilly & Reilly 2009). Consequently,
75 migrants are most likely selected to be generalists within their wintering habitat (Ivande & Cresswell

76 2016), with low and variable migratory connectivity (the serial residency hypothesis - Cresswell 2014).
77 A recent analysis of all migratory tracks that were available from tagged birds from 45 ecologically
78 diverse land-bird species (including Passeriformes, Apodiformes, Falconiformes, Cuculiformes and
79 Coraciiformes) in both Europe-Africa and the Americas flyways showed that connectivity at a large
80 scale was the strategy used by most species (Finch et al. 2017). Here, we test whether use of non-
81 breeding areas in a long distance migrant is likely to arise from a bet-hedging strategy by measuring
82 degree of connectivity and whether it is annually variable.

83 Testing theories of migratory connectivity on a sufficiently large scale has recently become possible
84 because of advances in tracking technology such as satellite and GPS tags: geolocators (light-level
85 loggers) now allow animals of relatively small body size to be tracked on global scales for the first
86 time (Bridge et al. 2013). We measured and tested migratory connectivity in the Whinchat *Saxicola*
87 *rubetra*, a common but declining Afro-Palearctic long distance migrant bird (Müller et al. 2005,
88 Britschgi et al. 2006). Uniquely, we measured connectivity: (i) from a location on the wintering ground
89 to the breeding grounds, which allows much greater precision of locations (because of more variable
90 daylength at higher latitudes, and because birds are highly likely to be stationary during the breeding
91 period, Lisovski et al. 2012), (ii) in two separate years, (iii) with relatively large sample sizes (for
92 geolocator studies, Bridge et al. 2013), (iv) in a system where we know there are no effects of tagging
93 (Blackburn et al. 2016), and (v) where there is very high non-breeding site fidelity (Blackburn &
94 Cresswell 2016b). We tested the following two predictions that arise directly if non-breeding site
95 selection has an important bet-hedging or stochastic component for first year birds, followed by site
96 fidelity for adults to the sites which allowed their non-breeding survival (Cresswell 2014):

- 97 1. There will be an extensive range (on the scale of >1000km) of breeding locations sampled from
98 individuals sampled at a single wintering location (Fig. 1).
- 99 2. There will be differences in breeding ranges between years when comparing samples from the
100 same wintering location (Fig. 2). This may be particularly pronounced when comparing cohorts (i.e.
101 first year birds only) because, in each year we would expect birds from spatially closer breeding
102 populations to be affected in a similar way by conditions on migration (and so tending to end up in
103 similar non-breeding areas), but for these conditions to vary between years (particularly because of
104 large-scale annually variable timing of breeding or possibly variation in departure post-breeding) so
105 that these non-breeding areas might shift in location annually.

106 We tested these predictions by mapping annual ranges and comparing how these differed in terms of
107 size, overlap and mean distance between individuals, with the expectation that annual ranges will be
108 large but variable between years.

109 **METHODS**

110 The study took place between February 2013 and November 2013 (Year 1) and February 2014 until
111 April 2015 (Year 2) during the dry season (early September to late April) on the Jos Plateau in the
112 guinea savannah zone of central Nigeria, West Africa ($N09^{\circ}53'$, $E08^{\circ}59'$, approximately 1250 m
113 altitude). Some colour-ringed only Whinchats were captured outside of these months (i.e. earlier in
114 the wintering period or were colour-ringed birds that had returned from previous winters) to evaluate
115 whether the geolocators affected survival (see Blackburn et al. 2016). Whinchats were captured within
116 an area of approximately 5 x 8 km; full site details are described in Blackburn and Cresswell (2016b).
117 Capture areas were principally open scrubland with varying degrees of habitat degradation from
118 human habitation, arable farming and livestock grazing, the latter increasing in intensity over the dry
119 season (see Hulme & Cresswell 2012, Blackburn & Cresswell 2015). The study area represents
120 typical wintering habitat for this species in the region (open savannah) and has high densities of
121 Whinchats.

122 Whinchats were caught with spring traps and mist nets in late February and March in 2013 or 2014
123 (birds in 2014 were on average tagged 10.3 ± 3.3 SE days earlier than in 2013, $F_{1,27} = -3.2$, $P =$
124 0.004). Birds were aged and sexed (Jenni & Winkler 1994), ringed with unique combinations of
125 colour-rings, and fitted with a geolocator. In Year 1, we deployed 49 and in Year 2 we deployed 131
126 geolocators fitted using leg-loop 'Rappole-Tipton' (also called backpack) harnesses. Full details of tag
127 and harness design are given in Blackburn et al. (2016), but importantly for this study, there was
128 variation between years in the length of light stalks of the tags. In 2013 all tags had a 10mm length
129 light stalk, whereas in 2014 tags had either no light stalks, 5mm or 10mm length light stalk (see
130 below). Tags weighed on average 0.63 g (0.01SE), representing 4.1 % of average body mass. There
131 was no overall significant reduction in between-year resighting rate (our proxy for survival, Blackburn
132 & Cresswell 2016b) comparing tagged and untagged birds in either year (Blackburn et al. 2016).

133 Attempts were made to recapture any returning tagged bird resighted in the following winter. Upon
134 recapture, geolocators were removed by cutting the harness and birds were released unharmed after
135 briefly assessing body condition (see Blackburn et al. 2016). Sample sizes of breeding locations used

136 in this paper are less than the number of individuals that returned with geolocators because 18/39
137 returning birds in 2014 could not be recaptured (because many had become extremely wary of spring-
138 traps and mist-nets), two individuals had lost their loggers because of harness failure, and several
139 loggers in 2013 suffered battery failure before the birds reached the breeding ground. Overall we
140 include all possible data from 29 individuals (12 in 2013 and 17 in 2014 – any individual was tracked
141 only in one year); data came from eight females and 21 males (ratio was not significantly different by
142 year, Fisher's exact test $P = 0.41$), and 13 adults (see Supplementary Material Figure S2 for the
143 locations of these individuals), 13 first years and three that could not be aged confidently (ratio of
144 known age birds was not significantly different by year, Fisher's exact test $P = 0.43$). In 2014, of the
145 recaptured birds, 4 had tags with no light stalk (see Supplementary Material Figure S2 for the
146 locations of these individuals), 8 with 5 mm light stalks and 5 with 10 mm light stalks.

147 Analyses

148 Raw data were downloaded, viewed and preliminarily cleaned using the BASTrack software suite
149 (British Antarctic Survey, Cambridge, UK; see Fox 2010 for an overview of the following processes).
150 We adjusted for clock drift, assuming that any drift was linear. We used the Transedit2 software that is
151 part of the BASTrack software to view raw data as light curves over time. We used a threshold value
152 of 2 to define sunrise, which is close to civil twilight. False twilight events due to shading from weather
153 or vegetation were identified and removed with the ‘minimum dark period’ filter (we used 4 hours),
154 which removes any impossible sunrise and sunset events (for a review and exploration of the effects
155 of environmental factors on geolocator data, see Lisovski et al. 2012). Data were then visually
156 inspected to ensure that only one sunrise and sunset occurred within any 24-hour period.

157 Further analyses were then carried out using R 2.13.1 (R Development Core Team 2014). We used
158 the LoessFilter in the R package ‘GeoLight’ (Lisovski & Hahn 2012) to validate twilight events and
159 identify those influenced by shading events at dawn and dusk: a polynomial regression of twilight
160 events identified residuals that were greater than three times the interquartile range and are therefore
161 likely to be erroneous (Lisovski & Hahn 2012). Identified outliers were checked within the original data
162 and were retained because large movements resulting from migration were sometimes incorrectly
163 identified as outliers, and outliers were reasonably normally distributed in the winter and summer
164 periods analysed (and so their effects cancelled out when calculating mean locations).

Because conditions away from the wintering grounds are unknown, we carried out two calibrations to determine the correct sun elevation angle value (henceforth ‘SEA’, which influences the determined latitude) for winter (SEAw) and summer (SEAs). For SEAw, we used the LocatorAid software from the BASTrack software suite, which uses known residency times and wintering location to calculate the corresponding wintering ground SEA value. We attempted to find the correct summer angle using the Hill-Ekstrom (H-E) calibration method (Hill & Braun 2001, Ekstrom 2004, see Tottrup et al. 2012) in which we selected the SEA value that gave the least amount of variation in latitude. If the calibration was not successful, we used the mean value for all of the loggers for which the calibration had been successful, calculated for each year separately (see Supplementary Material Figure S1 for a mapped description of which birds were affected; 4 in 2013 and 12 in 2014, and the locations associated with each SEA value used). Because the mean SEA value may have differed between years (e.g. because of inter-annual differences in cloud cover), we also simply used the same average SEA value for all birds regardless of year in a further analysis but with little change to the results (see below). In further sensitivity analyses (see below), we also varied the SEA value for these ‘mean value’ birds over the range of observed values from the Hill-Ekstrom calibration for that year to demonstrate that this uncertainty (and the assumption of using mean values for those birds where the calibration did not give clear results) made little difference to the results (see below). Once a corresponding SEAw and SEAs value was known for each logger, noon and midnight locations for a two week period when variance in locations was minimal (and all Whinchats should have been stationary on their breeding grounds in any case) were derived from sunrise and sunset times using the ‘coord’ function in the Geolight package (Lisovski & Hahn 2012). Breeding location was further confirmed from visual inspection of latitude and longitude changes with date and through identification of stationary periods using Geolight. We used the mean of all summer locations (calculated using SEAs values) for the first two weeks in June for all birds except two which showed evidence of some small movement in June where we used the last two weeks in May, and one bird that showed evidence of major movement in June, to a new stable location in July (perhaps after breeding failure) where we used the first two weeks in July (‘Best’ Analysis 1). All two week ‘most’ stationary locations identified were at very similar locations to the mean of all locations from mid-May to mid-July for all but the three birds which moved during this period, and for these three birds, the ‘second’ location was then stable for several weeks afterwards probably indicating breeding. Using these ‘best’ most stationary periods, or simply all locations from each bird in June gives very similar results for all analyses (See Supplementary

196 Material Tables S1 and S2, compare results for Analysis 1 & 4); other periods between mid-May to
197 the end of July gave similar results (unpubl. data).

198 Maps were plotted using the raster, rgdal, rgeos and mapproj libraries in R. Minimum convex polygon
199 areas, overlaps of these polygons and distances (considering the curvature of the earth) were
200 calculated using the alphahull and geosphere libraries in R. Areas were calculated for both years
201 combined and for 2013 and 2014 separately; the range overlap between the two years as a
202 percentage of the overall range (PHR) was also calculated. More sophisticated analyses using
203 kernels generated from the density of points were not appropriate because we had only a single
204 breeding location for each bird (albeit with variation due to geolocator errors in precision and SEA
205 assumptions) and so sample sizes of overall points split by year were too small for meaningful
206 calculations of probability density functions.

207 Ranges are sensitive to sample size with range increasing with sample size until an asymptote is
208 reached when individuals from nearly all parts of the range are part of the sample. We tested the
209 degree to which the ranges which we measured approached an asymptote, i.e. whether the range
210 calculated from our small samples likely approached the true range. The effect of sample size on the
211 range recorded overall and in each year separately was tested by randomly sub-sampling the
212 breeding locations for all birds, selecting 4 points up to $N - 1$ points overall and for the individual
213 years, 1000 times and then calculating the range for each sub-sample. The relationship between
214 mean range of the 1000 sub-samples and number of tagged birds was then tested with a GLM to
215 determine the line of best fit, comparing linear and quadratic fits to determine if an asymptote in range
216 with sample size had been reached.

217 We tested whether there was any significant difference in the size of the range and degree of overlap
218 of range between the two years by creating comparison data sets of the range size and overlap that
219 would occur if two groups of random points were selected from the total breeding range observed in
220 the two years. (i) 29 geographic points were selected randomly from the combined observed range
221 plotted for all individuals across both years; the first 12 points were placed in group 1 corresponding
222 to the first year's sample and the final 17 points in group 2 corresponding to the second year's
223 sample. Minimum convex polygons for each group and the combined groups were calculated using
224 the alphahull library in R; the area in each year and the difference in area between years, overall area
225 and the overlap in area between the two groups as a percentage of this total area for both groups was

226 then calculated. The procedure was repeated 1000 times to determine the 95% confidence intervals
227 of area difference and overlap for randomly selected points (i.e. with random difference in location
228 between years). (ii) We repeated (i) with 13 points randomly selected (7 in group 1 and 6 in group 2)
229 to create a similar random comparison data set considering only the first year birds.

230 We also measured the mean distance between breeding locations in each year (equivalent to
231 migratory spread as calculated in Finch et al. 2017). A matrix of all possible pairs of distances was
232 created and the distance between locations was calculated using the distHaversine function in R, and
233 then the mean and standard error of these annual mean distances was calculated. We tested whether
234 there was any significant difference in mean spread of locations between years using a t test with
235 degrees of freedom conservatively set to the number of original locations (i.e. $N = 29$ for all birds and
236 $N = 13$ for first year birds).

237 We tested for any systematic biases in latitude or longitude calculated with respect to variation in sex,
238 age, stalk length, SEA value, and whether this SEA value was estimated using the H-E calibration or
239 an average, in a GLM controlling for year. We also included interactions with year in this model to test
240 whether any effects of these potential biases acted differently in the two years of the study. There
241 were no significant effects (systematic biases that may have accounted for the results) in any
242 variables or interactions apart from light stalk length (See Supplementary Material Table S3 & S4). A
243 significant effect of light stalk length on latitude (but not longitude) was found so that more southerly
244 latitudes were recorded from tags without light stalks, dependent on the SEA value considered
245 (details in Table 2). Correspondingly we adjusted all latitudes for the light stalk effect so that if a tag
246 did not have a light stalk we added the parameter estimate for light stalk (between +3.8 to +5.0
247 degrees of latitude – see Table 2) to the estimate of latitude: note this had the effect of reducing any
248 differences between years, because 2014 locations were generally to the south of 2013 locations (see
249 Figures in Supplementary Material and particularly Figure S2 which identifies those loggers that had
250 no light stalks). Results using stalk length as a three-way factor, or as a two-way factor of light stalk
251 presence or absence, were similar, and models which considered stalk length as a three-way factor
252 were more than 2 AIC units worse than identical models with stalk length as a two-way factor. We
253 therefore only considered the presence and absence of a light stalk when considering the potential
254 confounding effects of tag design in detail. There was no significant variation in latitude with age or
255 sex in any model and models were substantially ($\Delta\text{AIC} > 2$) improved by the removal of these terms

which allowed the full data set to be used (i.e. by including those birds that lacked age data; Supplementary Material Table S3 & S4). Final adjustment of latitude was therefore calculated from a model including year and light stalk presence or absence (Supplementary Material Table S3 & S4). The appropriate adjustment (i.e. depending on SEA value) was used for each analysis and its associated range of SEAs values (Table 2). Because of the potential confounding effect of light stalk between the years that we identified, all analyses were carried out on both the raw observed location data set (Table 1) and also the data set with locations adjusted for the absence of a light stalk (Table 2).

Because of uncertainty introduced by not knowing breeding location SEA value, we also analysed the data under a set of scenarios with different assumptions for the SEA value (Supplementary Material Tables S1 & S2). These analyses were: 1. 'Best' with summer SEAs values determined using the Hill-Ekstrom calibration, or the mean value from the calibrations for that year for birds where the calibration was not conclusive, for the 2 week breeding season period with lowest variation in latitude; 2. Winter SEA_w value for the 2 week breeding season period with the lowest variation in latitude; 3. Average SEA value - the mean location calculated for each bird for sun elevation angles 2 to 6 at 0.5 increments (i.e. 9 mean locations), then averaged across these 9 locations, applied to all birds regardless of year; 4. As analysis 1, but for *all* locations in June; 5. As analysis 1, but with the maximum or minimum SEA value observed for that year used for those birds where the Hill-Ekstrom calibration was not conclusive, so that the range for each year was minimised (i.e. an analysis to examine the minimum range possible under the uncertainty of the SEAs values); 6. As analysis 1, but with the maximum or minimum SEA value observed for that year used for those birds where the Hill-Ekstrom calibration was not conclusive, so that overlap for each year was maximised (i.e. an analysis to examine the maximum overlap possible under the uncertainty of the SEAs values).

In short, extensive sensitivity analyses (Tables S1 & S2 and illustrated in Figures S3 – S20 in Supplementary Material) - regardless of SEA value and light stalk adjustment assumptions – showed very similar results. The results presented here use the data set which we considered to be the most accurate (Analysis 1. 'best' data set), but these results are representative of other scenarios which vary the assumptions of sun elevation angle; these are also detailed in Supplementary Material Tables S1 & S2.

285 Best models were identified on the basis of AIC, and models differing in AIC of less than 2 were
286 considered as equally valid (Burnham and Anderson 2002). Model fits were evaluated from diagnostic
287 model plots, and models were presented if assumptions were reasonably met (Crawley 2007). Mean
288 values are presented with one standard error (SE) in all cases; R^2 values were adjusted in all cases.

289 **Results**

290 *Overall range extent*

291 Whinchats tagged within 40 km² of central Nigeria had a breeding range extent of up to 2.549 million
292 km² (range 2.351 – 2.549 million km²; note here and subsequently, range is calculated across the
293 different SEA values, see Supplementary Material Tables S1 & S2), or 26% of the land area of
294 Europe (Table 1, Figure 3). The latitudinal range of the breeding grounds was approximately 16.3
295 degrees from Serbia to the southern border of Finland (a north to south distance of approximately
296 1,700 km) and the longitudinal range was 35.9 degrees from Poland to the Ural Mountains in Russia
297 (an east to west distance of approximately 2,300 km; Fig. 3). When values were adjusted for the
298 effect of light stalk absence, breeding range extent was 1.681 million km² (1.681 – 1.835 million km²),
299 or 17% of the land area of Europe (Table 2).

300 The total breeding range observed overall did not reach an asymptote with sample size. A straight line
301 relationship between range and sample size for all birds unadjusted for light stalk absence gave a
302 reasonable model fit (Fig. 4a). When latitudes were adjusted for the absence of light stalks, a
303 quadratic fit was much better for the relationship between range and sample size, but although the
304 relationship was becoming less steep it did not approach an asymptote (Fig. 4b).

305 *Annual variation in range extent*

306 Whinchats had larger ranges in 2014 compared to 2013. The breeding area of all Whinchats tagged
307 in central Nigeria, unadjusted for light stalk absence, was much lower in 2013 (0.651 million km²,
308 range 0.503 – 0.744) than that of the range in 2014 (2.157 million km², range 1.971 – 2.230; Fig. 5).
309 The differences in unadjusted area between the two years were greater than expected by chance for
310 all SEA analyses for all birds (Table 1, Supplementary Material Table S1). The breeding area of all
311 Whinchats tagged in central Nigeria in 2013 was much smaller than that of the range adjusted for light
312 stalk absence in 2014 (1.650 million km², range 1.650 – 1.814); the differences in adjusted area
313 between the two years were greater than expected by chance for all SEA analyses (Table 2,

314 Supplementary Material Table S2). For first year birds only, the breeding area unadjusted for light
315 stalk absence in 2013 (0.470 million km², range 0.429 – 0.518) was much smaller than that of the
316 range in 2014 (1.199 million km², range 1.145 – 1.199; Fig. 3). The differences in area between the
317 two years were greater than expected by chance for all SEA analyses of first year birds (Table 1,
318 Supplementary Material Table S1). For first year birds only, adjusted for light stalk presence in 2014,
319 the breeding area in 2013 was much smaller than that of the 2014 area (1.232 million km², range
320 0.829 – 1.232). The differences in area between the two years were greater than expected by chance
321 for all SEA analyses of first year birds, apart from when using the summer SEA for all locations in
322 June (Table 2, Supplementary Material Table S2).

323 Whinchat ranges in the two years generally overlapped relatively little, although this was less clear in
324 first year birds. The breeding range of all Whinchats showed relatively little overlap between years:
325 only 15% (range 8.7% - 27.1%) of the total range for both years was overlap, with all overlaps being
326 lower than that expected by chance (Table 1, Supplementary Material Table S1). For latitudes
327 adjusted for light stalk absence, overlap was 37% (range 31% - 42%) with a trend ($P = 0.03$ to $P =$
328 0.07) for overlaps to be lower than that expected by chance (Table 2, Supplementary Material Table
329 S2). For first year birds only, there was also a trend for relatively low overlap in the breeding ranges
330 between years: 8% (range 2.7% - 18.9%) of the total range for both years overlapped, with overlaps
331 only being lower than that expected by chance when using the summer SEA for all locations in June,
332 and marginally significant for the 'best' analysis (Table 1, Supplementary Material Table S1). For
333 latitudes adjusted for light stalk absence for first year birds, however, there was reasonable overlap in
334 the breeding ranges between years: 33% (range 3.3% - 39%) of the total range for both years
335 overlapped, with all overlaps only being lower than that expected by chance when using the summer
336 SEA for all locations in June (Table 2, Supplementary Material Table S2).

337 The mean distance between breeding locations was greater in 2014. For all birds, the mean distance
338 was 604 km in 2013 (range 591 – 631) and was 869 km in 2014 (range 839 – 903) and the difference
339 was highly significant for all SEA analyses (Table 1, Supplementary Material Table S1). For all birds,
340 the mean distance adjusted for light stalk in 2014 was 820 km (range 817 – 828) and the difference
341 was highly significant for all SEA analyses (Table 2, Supplementary Material Table S2). For first year
342 birds, the mean distance was 598 km in 2013 (range 567 – 631) and was 1059 km in 2014 (range
343 1013 – 1060) and the difference was highly significant for all SEA analyses (Table 1, Supplementary

344 Material Table S1). The mean distance adjusted for light stalk in 2014, for first year birds, was 1042
345 km (range 950 – 1042) and the difference was significant for all SEA analyses (Table 2,
346 Supplementary Material Table S2).

347 The total breeding range observed in each year separately did not reach an asymptote with sample
348 size. For all birds, the relationship between range and sample size was dependent on sample size,
349 but with different slopes dependent on year (interaction of year* sample size $t_{1,19} = 11.5$, $P < 0.0001$).
350 There was therefore a significantly greater spread in breeding locations for 2014 (0.141 ± 0.007)
351 compared to 2013 (0.061 ± 0.006), i.e. a much higher rate of increase in range per increase in sample
352 size (Fig. 4c). The relationships between range and sample size were linear (overall model $R^2 = 0.99$).
353 The same result was found using 2014 range adjusted for the effect of light stalk absence. For all
354 birds adjusted for the effect of light stalk, the relationship between range and sample size was
355 dependent on sample size, but with different slopes dependent on year (interaction of year*sample
356 size $t_{1,19} = 11.7$, $P < 0.0001$): there was therefore a significantly greater spread in breeding locations
357 for 2014 (0.109 ± 0.004) compared to 2013 (0.061 ± 0.004): i.e. a much higher rate of increase in
358 range per increase in sample size (Fig. 4d). The relationships between range and sample size were
359 linear (overall model $R^2 = 0.99$).

360 **DISCUSSION**

361 We made two predictions: a large range and an annually variable range. Both were supported by our
362 results. Our results probably show very low connectivity, i.e. connectivity only at a very large scale:
363 Whinchats from a single wintering location were spread out over a continental scale in Europe
364 (approximately 40% of the maximum east to west width of Europe). The reverse, that individuals from
365 the same breeding location in Europe will be spread out over a large non-breeding area in Africa, is
366 logically very likely to apply (Fig. 1), especially in a species with such high non-breeding site fidelity
367 (Blackburn & Cresswell 2016b), breeding ground site fidelity (Bastian 1992) and a non-breeding
368 distribution across the whole of Africa, from Senegal to Tanzania. Whinchats from any local breeding
369 population in the east of Europe are therefore likely to spread out over a similar continental scale in
370 Africa (33% of the maximum east to west width of Africa at typical wintering latitude, 8 degrees).
371 Furthermore, we suggest that the distribution of a wintering population of Whinchats in Africa may
372 also be annually variable, with potential shifts in average occurrence between years of the order of
373 hundreds of kilometres.

374 Our results, at least in terms of the large scale of the ranges and variation we have described, seem
375 reasonable. Although locations from geolocators have errors, these errors within individual stationary
376 periods were reasonably normally distributed, as probably were errors between individuals through
377 incorrectly assessing SEA value (e.g. some individuals will have lower and some higher SEAs values
378 than they should, but these will likely be random with respect to individual location on a large scale).
379 There were, for example, no correlations between SEA value used in the 'best' analysis and latitude
380 or longitude (model of SEAs ~ latitude + longitude; latitude 0.04 ± 0.03 , $t = 1.2$, $P = 0.24$, longitude -
381 0.002 ± 0.014 , $t = -0.2$, $P = 0.86$; $F_{2,26} = 0.7$, $R^2 = 0$). Furthermore, our analyses (analyses 2 & 3, see
382 Supplementary Material Tables S1 & S2) that simply used an average value of SEA for all birds will
383 have increased the randomness of the errors, and both gave very similar results to analyses which
384 used the best available information to more accurately estimate SEA value. Consequently, range
385 areas and range differences between years should be real, even if imprecise in location.
386 Fundamentally, the scale of geolocator errors (Fudickar et al. 2012) is also relatively small compared
387 to the scale of the ranges identified in this study. Our analyses, which used different assumptions
388 regarding sun elevation angle values, breeding stationary period and effect of absence of light stalks,
389 showed broadly similar results at the large scale of this study: despite variation in the individual
390 breeding locations in each analysis, 2013 birds had a smaller spread and range than 2014 birds and
391 there was reasonable evidence for a difference in range between years when comparing overlap.

392 Although the H-E calibration should have accounted for differences in shading between birds,
393 locations and years, the effect of light stalk presence / absence on latitude demonstrated that this
394 calibration was biased by shading (in this case, most likely through from flight feathers covering the
395 light sensor). Consequently, the latitudes of breeding locations in each year could have been biased
396 by the degree & variability of weather shading in the period for which data was used in the calibration,
397 especially when this was based on a short period (2 weeks). This could potentially have caused
398 spurious shifts in range between years. The difference in SEA between years after the H-E calibration
399 was small (2014, -0.55 ± 0.58 , $t_{1,9} = -1.0$, $P = 0.37$) and would have shifted 2014 locations by about
400 0.6 degrees latitude to the south away from 2013 locations, thus the bias, if it was operating, would
401 have increased any differences in range and overlap between years for the 'best' analysis (Analysis
402 1). It could not, however, completely explain the differences in spread of individuals between years
403 because this also has a longitudinal component (albeit due to a few birds). In any case, any such bias
404 was completely removed when using the same average SEA value for all birds regardless of year.

405 We still found variation when controlling for cohort effects, despite smaller sample sizes. The larger
406 population considered contained both first year and adult birds from previous years, each individual
407 year having a likely different cohort range effect. Any comparison between two consecutive years that
408 includes adult birds will therefore be conservative because, as more cohorts are included, the range
409 in any one year will regress to the mean range of the population. This will act to blur any differences
410 between years. However, Whinchat annual survival rate for this population is of the order of 54%
411 (Blackburn & Cresswell 2016a), so even when considering adults, it was likely that over 75% of the
412 population sampled in each year consisted of only two cohorts. Unfortunately, any convincing
413 analyses of only adult birds – which should show less variation (this was the case, but not included
414 here) - was precluded by only four adults being part of the 2013 sample versus nine in the 2014
415 sample (see Supplementary Material Figure S2), so making any confirmation of the null hypothesis
416 likely simply because of low power, regardless of any actual differences or not between years.

417 Although the results of the analyses are consistent regardless of assumptions, they rely on small
418 sample sizes. Any range or migratory spread calculation will be dependent on its relatively few
419 peripheral points, and the sample size for inter-annual comparisons is only two years. The overlap
420 result is also difficult to interpret, being confounded by variation in size as well as a possible shift in
421 location, although the study shows clear evidence of annual variation in measures that will affect
422 overlap (i.e. migratory spread, and the relationship between range and sample size). The study
423 clearly needs to be repeated, ideally tracking only first-year Whinchats from other areas in Africa in
424 consecutive years to determine inter-annual variability in range, with sufficient breeding locations
425 obtained to analyse range using probability density functions. Nevertheless, our results suggest that
426 low and variable connectivity can occur in Whinchats. Our range estimates did not approach an
427 asymptote with a sample size of 29 tagged birds and there were differences in range between years
428 sufficient that conclusions about where Whinchats wintering in central Nigeria may have bred, or their
429 degree of connectivity, would have differed on an international scale, if the study had been only
430 conducted in a single year, or with a different sample size.

431 Our results show migratory spread consistent with the pattern just emerging from other studies. Long
432 distance migrant birds have high migratory spread regardless of phylogeny and flyway; the predicted
433 average maximum distance between wintering individuals from the same breeding population, across
434 45 species, is of the order more than 3,000 km (Finch et al. 2017). It should be noted, however, that

435 all species lie on a continuum of connectivity and there are examples of particularly low connectivity
436 species (e.g. Pallid Harriers *Circus macrourus* Terraube et al. 2012) and particularly high connectivity
437 species (e.g. Nightingales *Luscinia megarhynchos* Hahn et al. 2013); Whinchats would appear to lie
438 closer to the extreme end of low migratory connectivity. Inter-annual shifts in range are less well
439 documented or tested, even in the few multi-year datasets, but Liechti et al. (2015) reported that Barn
440 Swallows *Hirundo rustica* wintered 400km to the north in one year compared to the population mean
441 for the previous year, and Swainson's Thrush *Catharus ustulatus* (Delmore and Irwin, 2014) and
442 Semi-collared Flycatcher *Ficedula semitorquata* (Briedis et al., 2016) populations also probably show
443 inter-annual variability in mean wintering location.

444 Although the low and variable connectivity probably demonstrated here could arise through a bet-
445 hedging strategy because first year non-breeding site selection is to a degree stochastic, other
446 mechanisms may have caused the pattern. As detailed in Figure 1, if migrants with wide breeding
447 ranges in Europe have relatively small wintering ranges in Africa (e.g. Eleonora's Falcon *Falco*
448 *eleonora* Gschwend et al. 2008, Kassara et al. 2012), effectively concentrating in a small non-
449 breeding area, and they are then sampled in that small area, they will show apparently very low
450 connectivity. This is unlikely to be the case for Whinchat, however, due to its very large non-breeding
451 distribution in Africa. Variation in breeding success on a continental scale may also cause differences
452 in breeding ranges when sampled in different years from Africa. For example, northerly breeding
453 Whinchats in 2012 may have produced more first years than southerly Whinchats, and vice-versa in
454 2013. Annual variation in settlement patterns for first breeding, particularly if a species has low or
455 environmentally variable natal site fidelity (Studds et al. 2008), could also lead to the patterns
456 observed here.

457 With the caveat that alternative explanations are possible, the results of this study are consistent with
458 the predictions of the serial residency hypothesis, where low and variable connectivity emerges from
459 juveniles having deterministic migratory directions varying around a mean 'approximate' migration
460 direction and then subsequent non-breeding locations further varying due to annually variable events
461 (e.g. breeding phenology and weather patterns during migration), followed by high wintering site
462 fidelity of surviving adults (Cresswell 2014). Such a system makes sense in terms of the evolution of
463 migration, which is an adaptation to large scale climate change and consequent shifting of suitable
464 habitat (Cresswell et al. 2011, Fryxell & Holt 2013). As required resources or environments shift in

465 location, at least some offspring from a brood will find survivable routes, thus a bet-hedging strategy –
466 where some offspring migrate in slightly different directions and/or different times (Reilly & Reilly
467 2009) - is adaptive when the location of suitable habitat is unknown and unpredictable (Cresswell
468 2014). As location of suitable conditions becomes more predictable, this bet-hedging strategy
469 becomes less adaptive because a proportion of offspring will always miss the target and arrive in
470 unsuitable habitat. Although adjustments within the wintering range can be made by further small or
471 even large scale migratory movements, there may be costs because migration itself can represent a
472 period of high mortality (Sillett & Holmes 2002, Strandberg et al. 2010).

473 A bet-hedging strategy is, however, unlikely to be adaptive if the availability of habitat declines,
474 because an increasing proportion of offspring (or the population) will miss the target. Thus the
475 connectivity optimum for climate change (bet-hedging leading to a wide non-breeding distribution) is
476 the opposite of the connectivity optimum for habitat loss (with specific site or area selection leading to
477 a narrow non-breeding distribution): migratory species cannot adapt to both. This creates an adaptive
478 conflict in the face of both shifting habitat due to climate change and disappearing habitat because of
479 rapidly increasing human populations in Africa. The many migrant species that apparently make a first
480 naïve bet-hedging migration (i.e. excluding those relatively few species that use social learning and
481 that migrate in mixed age groups) will have a wide non-breeding distribution and an annually
482 increasing proportion of their populations arriving in unsuitable habitat on the wintering ground, thus
483 reducing survival. The wide non-breeding distribution ensures that *any* reduction in habitat availability
484 over a very wide area of the wintering grounds affects *all* populations over a very wide breeding area
485 (Sutherland & Dolman 1994, Taylor & Norris 2010, Cresswell 2014). This general characteristic may
486 then significantly contribute to the decline of so many populations of migratory species (see Vickery et
487 al. 2014), regardless of their location on the breeding grounds and their ecology.

488 A number of predictions arise from this framework with respect to population trends and distribution.
489 First, population trends will be a function of a population's distribution or migratory spread on the non-
490 breeding ground: at some point the proportion of first years that miss the target must reduce juvenile
491 survival below the level of productivity/recruitment, although this will vary dependent on the degree of
492 non-breeding habitat/resource specialisation (i.e. how big the target is, or how costly it is to miss it).
493 Species with low migratory spread will have favourable population trends where they winter in
494 climatically stable areas and are likely to be habitat specialists; selection may act to reduce genetic

495 variability in migration direction and/or to promote social learning of routes. Species with high
496 migratory spread will have favourable population trends (e.g. see Gilroy et al. 2016), particularly if
497 they winter in climatically variable areas and are likely to be habitat generalists (e.g. Hulme &
498 Cresswell 2012, Blackburn & Cresswell 2015); selection may therefore act to increase variability in
499 migration direction. Consequently, species with intermediate migratory spreads and generalist habitat
500 requirements might have more favourable population trends. Second, appreciating that there are
501 cohort effects will greatly affect our understanding of distribution. Range should shift rapidly on the
502 wintering ground so that range maps built from sight records over decades may not accurately reflect
503 current range, or indeed the actual range in any year. Furthermore, models to predict the habitat
504 requirements of migrants at a large scale will be subject to errors both from the false negatives from
505 stochastic range shifts as well as the ‘false’ positives of juvenile cohorts in unsuitable areas with much
506 lower survival.

507 Testing such predictions is essential if we are to respond to the extremely rapid decline of migratory
508 species across global flyways. Whether we adopt species-specific and local site specific land sparing
509 solutions, rather than widespread land sharing solutions likely depends on degree of connectivity and
510 how it affects survival.

511

512 The study was carried out in Nigeria where no licences are required for the procedures used.
513 Nevertheless this study was carried out under the ethical guidelines of the AP Leventis Ornithological
514 Research Institute Scientific Committee (APLORI is the only ornithological research institute in
515 Nigeria) based on the Association for the Study of Animal Behaviour guidelines and those of the
516 British Trust for Ornithology’s ringing scheme. All personnel involved in fieldwork – either catching,
517 colour-ringing or tagging birds had BTO ringing licences. MB had been previously licensed to fit
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521 Ornithological Research Institute.

522

523 REFERENCES

- 524 Bastian, H.-V. 1992. Breeding and natal dispersal of Whinchats. *Ringing & Migration*, **13**: 13-19.
- 525 Bauer, S., Lisovski, S. & Hahn, S. 2016. Timing is crucial for consequences of migratory connectivity.
526 *Oikos*, **125**: 605-612.
- 527 Blackburn, E., Burgess, M., Freeman, B., Risely, A., Izang, A., Ivande, S., Hewson, C. & Cresswell,
528 W. 2016. An experimental evaluation of the effects of geolocator design and attachment
529 method on between-year survival on Whinchats *Saxicola rubetra*, a long distance migrant.
530 *Journal of Avian Biology*, **47**: 530-539.
- 531 Blackburn, E. & Cresswell, W. 2015. Fine-scale habitat use during the non-breeding season suggests
532 that winter habitat does not limit breeding populations of a declining long-distance Palearctic
533 migrant. *Journal of Avian Biology*, **46**: 622-633.
- 534 Blackburn, E. & Cresswell, W. 2016a. High overwinter and annual survival for a declining Palearctic
535 migrant: evidence that wintering conditions may not limit migrant populations *Ibis*, **158**: 92-
536 105.
- 537 Blackburn, E. & Cresswell, W. 2016b. High winter site fidelity in a long-distance migrant: implications
538 for wintering ecology and survival estimates. *Journal of Ornithology*, **157**: 93-108.
- 539 Botero, C. A., Weissing, F. J., Wright, J. & Rubenstein, D. R. 2015. Evolutionary tipping points in the
540 capacity to adapt to environmental change. *Proceedings of the National Academy of
541 Sciences*, **112**: 184-189.
- 542 Bridge, E. S., Kelly, J. F., Contina, A., Gabrielson, R. M., MacCurdy, R. B. & Winkler, D. W. 2013.
543 Advances in tracking small migratory birds: a technical review of light-level geolocation.
544 *Journal of Field Ornithology*, **84**: 121-137.
- 545 Britschgi, A., Spaar, R. & Arlettaz, R. 2006. Impact of grassland farming intensification on the
546 breeding ecology of an indicator insectivorous passerine, the Whinchat *Saxicola rubetra*:
547 Lessons for overall Alpine meadowland management. *Biological Conservation*, **130**: 193-205.

- 548 Costa, D. P., Breed, G. A. & Robinson, P. W. 2012. New Insights into Pelagic Migrations: Implications
549 for Ecology and Conservation. In *Annual Review of Ecology, Evolution, and Systematics*, Vol
550 43. (ed. D. J. Futuyma), **Vol. 43**, pp. 73-96.
- 551 Crawley, M. J. (2007) *The R Book*, Chichester: John Wiley & Sons Ltd.
- 552 Cresswell, K. A., Satterthwaite, W. H. & Sword, G. A. 2011. Understanding the evolution of migration
553 through empirical examples. In *Animal migration: a synthesis*. (eds. E. J. Milner-Gulland, J. M.
554 Fryxell & A. R. E. Sinclair), pp. 1-16. Oxford, U.K.: Oxford University Press.
- 555 Cresswell, W. 2014. Migratory connectivity of Palaearctic-African migratory birds and their responses
556 to environmental change: the serial residency hypothesis. *Ibis*, **156**: 493-510.
- 557 Ekstrom, P. A. 2004. An advance in the geolocation by light. In *Memoirs of the National Institute of
558 Polar Research, Special Issue*. (ed. Y. Naito), pp. 210-226. Tokyo: National Institute of Polar
559 Research.
- 560 Esler, D. 2000. Applying metapopulation theory to conservation of migratory birds. *Conservation
561 Biology*, **14**: 366-372.
- 562 Finch, T., Butler, S., Franco, A. & Cresswell, W. 2017. Low migratory connectivity is common in long-
563 distance migrant birds. *Journal of Animal Ecology*: n/a-n/a.
- 564 Flockhart, D. T. T., Pichancourt, J. B., Norris, D. R. & Martin, T. G. 2015. Unravelling the annual cycle
565 in a migratory animal: breeding-season habitat loss drives population declines of monarch
566 butterflies. *Journal of Animal Ecology*, **84**: 155-165.
- 567 Fox, J. 2010. Geolocator manual v8 (March 2010).
568 http://www.antarctica.ac.uk/engineering/geo_downloads/Geolocator_manual_v8.pdf.
- 569 Fryxell, J. M. & Holt, R. D. 2013. Environmental change and the evolution of migration. *Ecology*, **94**:
570 1274-1279.
- 571 Fudickar, A. M., Wikelski, M. & Partecke, J. 2012. Tracking migratory songbirds: accuracy of light-
572 level loggers (geolocators) in forest habitats. *Methods in Ecology and Evolution*, **3**: 47-52.

- 573 Gilroy, J. J., Gill, J. A., Butchart, S. H. M., Jones, V. R. & Franco, A. M. A. 2016. Migratory diversity
574 predicts population declines in birds. *Ecology Letters*, **19**: 308-317.
- 575 Greenwood, P. J. & Harvey, P. H. 1982. The natal and breeding dispersal of birds. *Annual Review of*
576 *Ecology and Systematics*, **13**: 1-21.
- 577 Gschweng, M., Kalko, E. K. V., Querner, U., Fiedler, W. & Berthold, P. 2008. All across Africa: highly
578 individual migration routes of Eleonora's falcon. *Proceedings of the Royal Society B-Biological*
579 *Sciences*, **275**: 2887-2896.
- 580 Hahn, S., Amrhein, V., Zehtindijev, P. & Liechti, F. 2013. Strong migratory connectivity and seasonally
581 shifting isotopic niches in geographically separated populations of a long-distance migrating
582 songbird. *Oecologia*, **173**: 1217-1225.
- 583 Hill, R. D. & Braun, M. J. 2001. Geolocation by light level - the next step: Latitude. In *Electronic*
584 *tagging and tracking in marine fisheries*. (eds. J. R. Silbert & J. L. Nielsen), pp. 315-330.
585 Dordrecht: Kluwer Academic publishers.
- 586 Hulme, M. F. & Cresswell, W. 2012. Density and behaviour of Whinchats *Saxicola rubetra* on African
587 farmland suggest that winter habitat conditions do not limit European breeding populations.
588 *Ibis*, **154**: 680-692.
- 589 Ivande, S. T. & Cresswell, W. 2016. Temperate migrants and resident bird species in Afro-tropical
590 savannahs show similar levels of ecological generalism. *Ibis*, **158**: 496-505.
- 591 Jenni, L. & Winkler, R. (1994) *Moult and ageing of European Passerines*, London: Academic Press.
- 592 Kassara, C., Fric, J., Gschweng, M. & Sfenthourakis, S. 2012. Complementing the puzzle of
593 Eleonora's Falcon (*Falco eleonorae*) migration: new evidence from an eastern colony in the
594 Aegean Sea. *Journal of Ornithology*, **153**: 839-848.
- 595 Lisovski, S. & Hahn, S. 2012. GeoLight-processing and analysing light-based geolocator data in R.
596 *Methods in Ecology and Evolution*, **3**: 1055-1059.

- 597 Lisovski, S., Hewson, C. M., Klaassen, R. H. G., Korner-Nievergelt, F., Kristensen, M. W. & Hahn, S.
598 2012. Geolocation by light: accuracy and precision affected by environmental factors.
599 *Methods in Ecology and Evolution*, **3**: 603-612.
- 600 Martin, T. G., Chades, I., Arcese, P., Marra, P. P., Possingham, H. P. & Norris, D. R. 2007. Optimal
601 Conservation of Migratory Species. *Plos One*, **2**.
- 602 Müller, M., Spaar, R., Schifferli, L. & Jenni, L. 2005. Effects of changes in farming of subalpine
603 meadows on a grassland bird, the whinchat (*Saxicola rubetra*). *Journal of Ornithology*, **146**:
604 14-23.
- 605 Newton, I. (2008) *The migration ecology of birds*, Oxford: Academic Press.
- 606 R Development Core Team. 2014. R: A Language and Environment for Statistical Computing. Vienna,
607 Austria: R Foundation for Statistical Computing.
- 608 Reilly, J. R. & Reilly, R. J. 2009. Bet-hedging and the orientation of juvenile passerines in fall
609 migration. *Journal of Animal Ecology*, **78**: 990-1001.
- 610 Runge, C. A., Martini, T. G., Possingham, H. P., Willis, S. G. & Fuller, R. A. 2014. Conserving mobile
611 species. *Frontiers in Ecology and the Environment*, **12**: 395-402.
- 612 Runge, C. A., Watson, J. E. M., Butchart, S. H. M., Hanson, J. O., Possingham, H. P. & Fuller, R. A.
613 2015. Protected areas and global conservation of migratory birds. *Science*, **350**: 1255-1258.
- 614 Sillett, T. S. & Holmes, R. T. 2002. Variation in survivorship of a migratory songbird throughout its
615 annual cycle. *Journal of Animal Ecology*, **71**: 296-308.
- 616 Strandberg, R., Klaassen, R. H. G., Hake, M. & Alerstam, T. 2010. How hazardous is the Sahara
617 Desert crossing for migratory birds? Indications from satellite tracking of raptors. *Biology
618 Letters*, **6**: 297-300.
- 619 Studds, C. E., Kyser, T. K. & Marra, P. P. 2008. Natal dispersal driven by environmental conditions
620 interacting across the annual cycle of a migratory songbird. *Proceedings of the National
621 Academy of Sciences of the United States of America*, **105**: 2929-2933.

- 622 Sutherland, W. J. & Dolman, P. M. 1994. Combining behaviour and population dynamics with
623 applications for predicting consequences of habitat loss. *Proceedings of the Royal Society of*
624 *London Series B*, **255**: 133-138.
- 625 Taylor, C. M. & Norris, D. R. 2010. Population dynamics in migratory networks. *Theoretical Ecology*,
626 **3**: 65-73.
- 627 Terraube, J., Mogeot, F., Cornulier, T., Verma, A., Gavrilov, A. & Arroyo, B. 2012. Broad wintering
628 range and intercontinental migratory divide within a core population of the near-threatened
629 pallid harrier. *Diversity and Distributions*, **18**: 401-409.
- 630 Tottrup, A. P., Klaassen, R. H. G., Strandberg, R., Thorup, K., Kristensen, M. W., Jorgensen, P. S.,
631 Fox, J., Afanasyev, V., Rahbek, C. & Alerstam, T. 2012. The annual cycle of a trans-
632 equatorial Eurasian-African passerine migrant: different spatio-temporal strategies for autumn
633 and spring migration. *Proceedings of the Royal Society B-Biological Sciences*, **279**: 1008-
634 1016.
- 635 Vickery, J. A., Ewing, S. R., Smith, K. W., Pain, D. J., Bairlein, F., Skorpilova, J. & Gregory, R. D.
636 2014. The decline of Afro-Palaearctic migrants and an assessment of potential causes. *Ibis*,
637 **156**: 1-22.
- 638 Webster, M. S., Marra, P. P., Haig, S. M., Bensch, S. & Holmes, R. T. 2002. Links between worlds:
639 unraveling migratory connectivity. *Trends in Ecology & Evolution*, **17**: 76-83.
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- 641

642 **Table 1.** Observed area and overlap (million square kilometres) of breeding locations compared to a
 643 random sample (lower and upper 95% CI), and mean distance (\pm 1 SE km) between breeding
 644 locations for all birds (a) and first year birds (b) with ‘best’ summer sun elevation angle values (SEAs).
 645 See Supplementary Material Table S1 for a comparison of these values when using different sun
 646 elevation angle values: all are reasonably similar regardless of assumptions of sun elevation angle
 647 value.

648

	Overall	2013	2014	Difference 2014 - 2013	Lower CI value for the random sample	95% CI value for the random sample	Upper 95% CI value for the random sample	P value
a. All birds								
Area	$N = 29$	$N = 12$	$N = 17$					
Overlap	2.549	0.651	2.157	1.506	-0.612	1.225	0.01	
Mean distance			604 ± 18	869 ± 33	$t_{28} = 5.2$			0.0001
b. First year birds								
	$N = 13$	$N = 7$	$N = 6$					
Area	1.700	0.470	1.199	0.729	-0.857	0.589	0.02	
Overlap		7.9			3.8	56.8	0.06	
Mean distance			598 ± 40	1059 ± 72	$t_{12} = 4.1$			0.0010

649

650 **Table 2.** Observed area and overlap (million square kilometres) of breeding locations compared to a
 651 random sample (lower and upper 95% CI), and mean distance (\pm 1 SE km) between breeding
 652 locations for all birds (a.) and first year birds (b.) with ‘best’ summer sun elevation angle values
 653 (SEAs), latitudes adjusted for the effect of no light stalks on four birds in 2014. Latitude adjustments
 654 for each analysis are given in italics. Sample sizes as in Table 1. See Supplementary Material Table
 655 S2 for a comparison when using different sun elevation angle values: all are reasonably similar
 656 regardless of assumptions of sun elevation angle value.

657

	Overall	2013	2014	Difference 2014 - 2013	Lower 95% CI value for the random sample	Upper 95% CI value for the random sample	P value
a. All birds							
Area	1.681	0.651	1.650	0.999	-0.440	0.814	0.01
Overlap	36.5				35.2	77.8	0.06
Mean distance		604 ± 18	820 ± 29	$t_{28} = 4.5$			0.0001
b. First year birds							
Area	1.275	0.470	1.232	0.762	-0.651	0.486	0.01
Overlap	33.2				4.2	60.3	0.51
Mean distance		598 ± 40	1042 ± 63	$t_{12} = 4.3$			0.001

658

659

660 **Figure legends**

661 **Figure 1.** How sampling at any one point on the non-breeding ground can result in sampling birds
662 over a large breeding range (i.e. low connectivity arises from wide migratory dispersal). This model
663 assumes high natal (see Greenwood & Harvey 1982) and non-breeding (see Cresswell 2014) site
664 fidelity. The model also assumes that the non-breeding range of a migrant is at a very large scale (as
665 it is with the Whinchat and most other Afro-Palearctic migrants which have non-breeding ranges that
666 encompass much of Africa), where migrants may have small, concentrated, non-breeding areas thus
667 tagging from these areas will also result in sampling birds over a large breeding range.

668 **Figure 2.** How sampling at any one point on the wintering ground in different years can result in
669 sampling birds with different breeding ranges (i.e. inter-annual variation in connectivity arises through
670 stochastic cohort effects such as, for example wind conditions during migration varying between
671 years). This model assumes high natal site fidelity (see Greenwood & Harvey 1982) and high non-
672 breeding site fidelity if a first year bird has found a non-breeding site that allowed survival (see the
673 serial residency hypothesis Cresswell 2014).

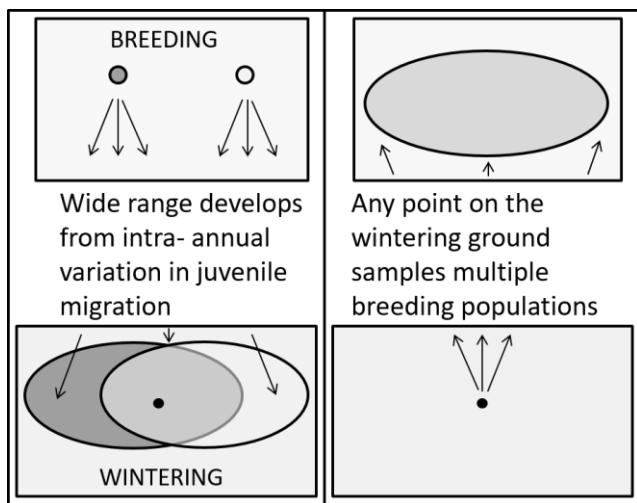
674 **Figure 3.** Breeding locations and range for 29 Whinchats tagged at Jos Nigeria (A) in 2013 (black
675 circular points, white fill) and in 2014 (white triangular points and darker grey fill). Each point is plotted
676 at the mean of all latitudes and longitudes recorded over the least variable 2 week stationary period
677 during mid-May – mid-July; +/- 2 Standard Errors from the mean location are also plotted but only the
678 highly variable locations are clearly visible at this scale. All birds plotted at the ‘best’ estimate
679 locations, unadjusted for light stalk absence (see Table 1).

680 **Figure 4.** How range depended on sample size. Mean range plotted for 1000 random subsamples of
681 points, each of different sample size. a. All birds and years pooled, unadjusted for light stalk absence.
682 A linear fit is plotted with 95% CI plotted as dotted lines ($\text{range} = (0.097 \pm 0.003 \times \text{sample size}) + (-$
683 $0.03 \pm 0.04)$, $F_{1,12} = 37.2$, $P < 0.0001$; $R^2 = 0.99$; $\Delta\text{AIC} = + 0.2$, compared to a quadratic model
684 including the square of sample size which gives a near identical plot). b. All birds and years pooled,
685 adjusted for light stalk absence. Line of best fit is quadratic, 95% CI plotted as dotted lines ($\text{range} =$
686 $(0.094 \pm 0.005 \times \text{sample size}) + (-0.0011 \pm 0.0002 \times \text{sample size}^2) + (-0.04 \pm 0.03)$, $F_{1,11} = 1010$, $P <$
687 0.0001 ; $R^2 = 0.99$; $\Delta\text{AIC} = - 19.8$ better than a linear model without the square of sample size). c. All
688 birds, years plotted separately, unadjusted for light stalk absence. Lines of best fit are linear (1
689 standard error plotted as dotted lines) and the gradients of the two lines are highly significantly

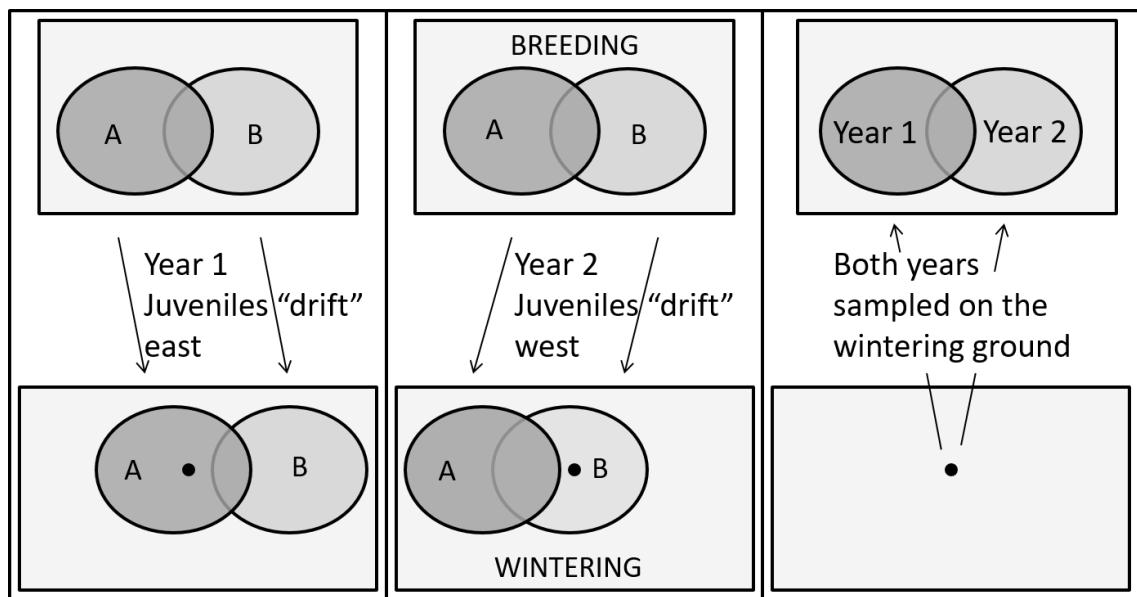
690 different. d. All birds, years plotted separately, adjusted for light stalk absence. Lines of best fit are
691 linear (1 standard error plotted as dotted lines) and the gradients of the two lines are highly
692 significantly different.

693 **Figure 5.** Breeding locations and range for Whinchats tagged in 2013 (black circular points, white fill,
694 N = 12) and in 2014 (white triangular points and darker grey fill, N = 17). Each point is plotted at the
695 mean of all latitudes and longitudes recorded over the least variable 2 week stationary period during
696 mid-May – mid-July (N = 28 locations), +/- 2 Standard Errors from the mean location are also plotted
697 but only the highly variable locations are visible at this scale. a. All birds plotted at the ‘best’ estimate
698 locations, unadjusted for light stalk absence (i.e. an enlarged version of Fig. 3). b. All birds plotted at
699 the ‘best’ estimate locations, adjusted for light stalk absence. c. First year birds plotted at the ‘best’
700 estimate locations, unadjusted for light stalk absence. d. First year birds plotted at the ‘best’ estimate
701 locations, adjusted for light stalk absence. Ranges that arose from varying the assumptions with
702 respect to sun elevation angles (see Table 1 and 2) are broadly similar and are illustrated in
703 Supplementary Material Figures S3 – S20).

704 Figure 1:



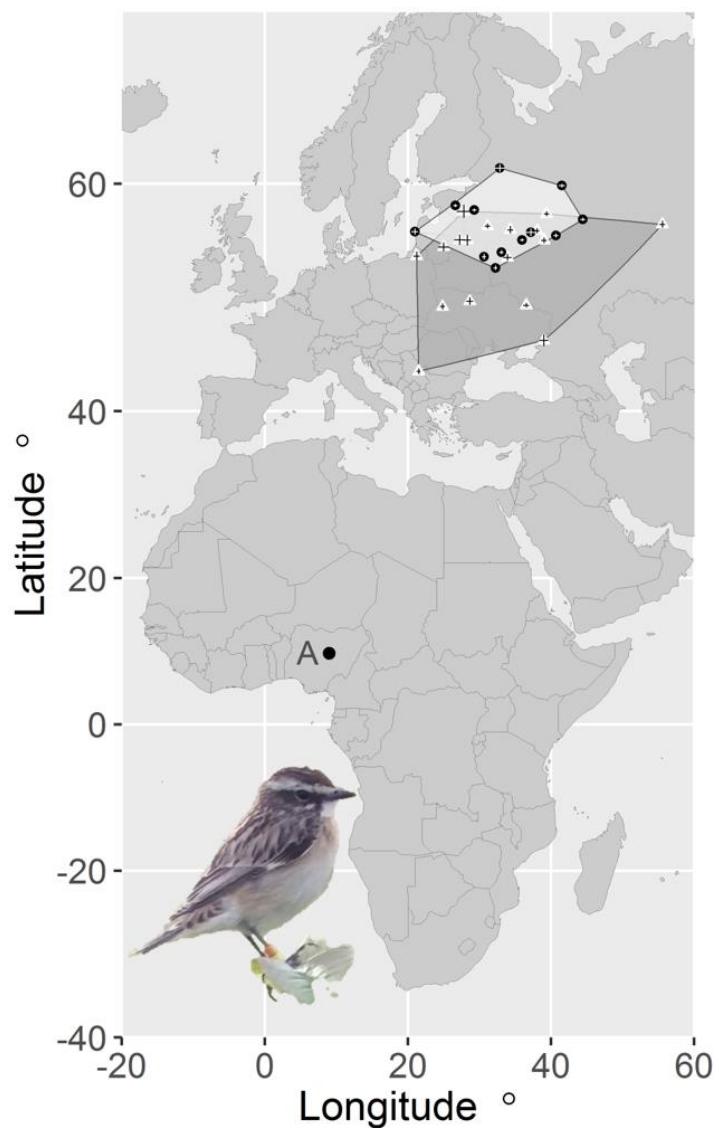
708 Figure 2:



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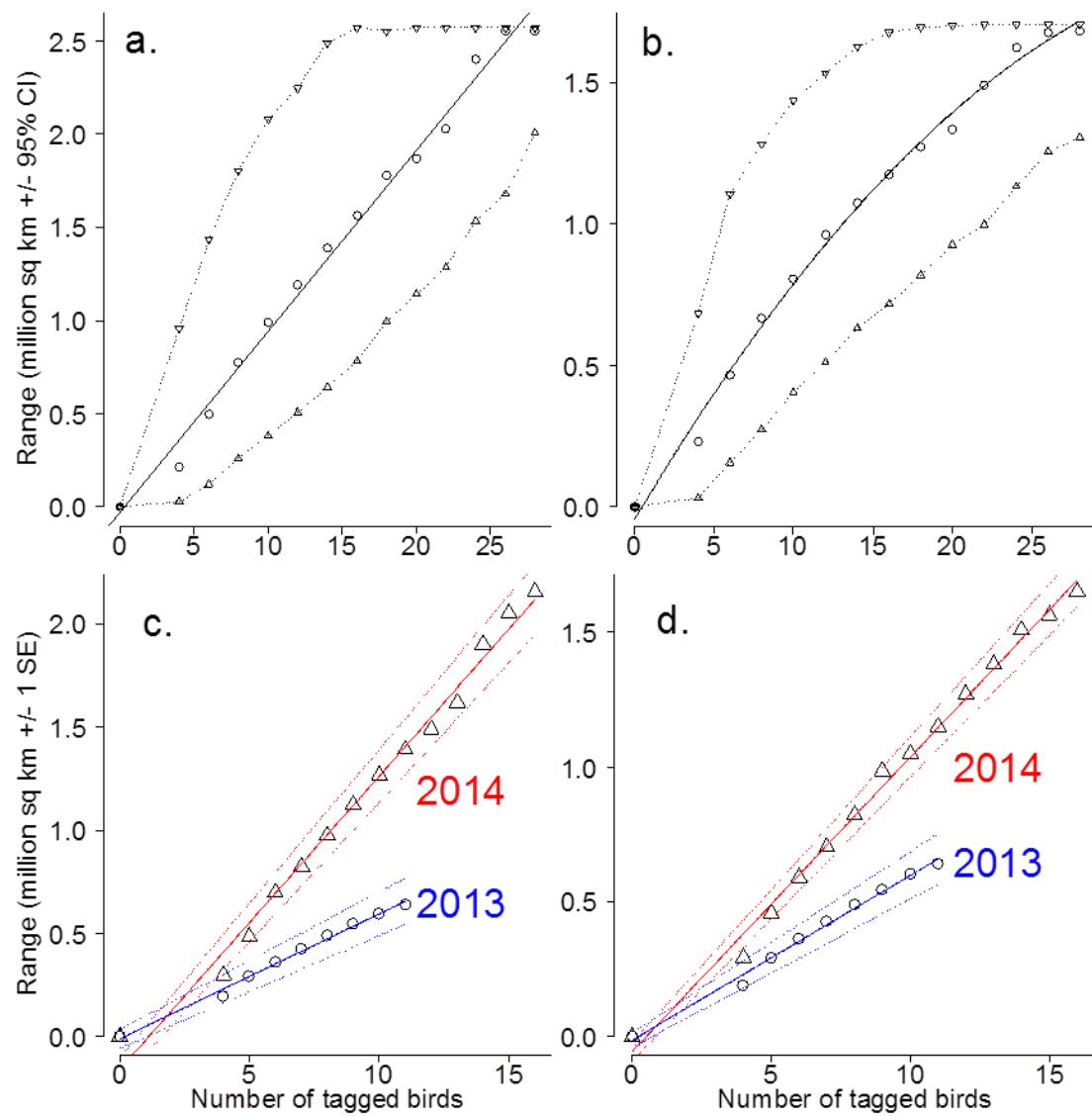
710

711 Figure 3



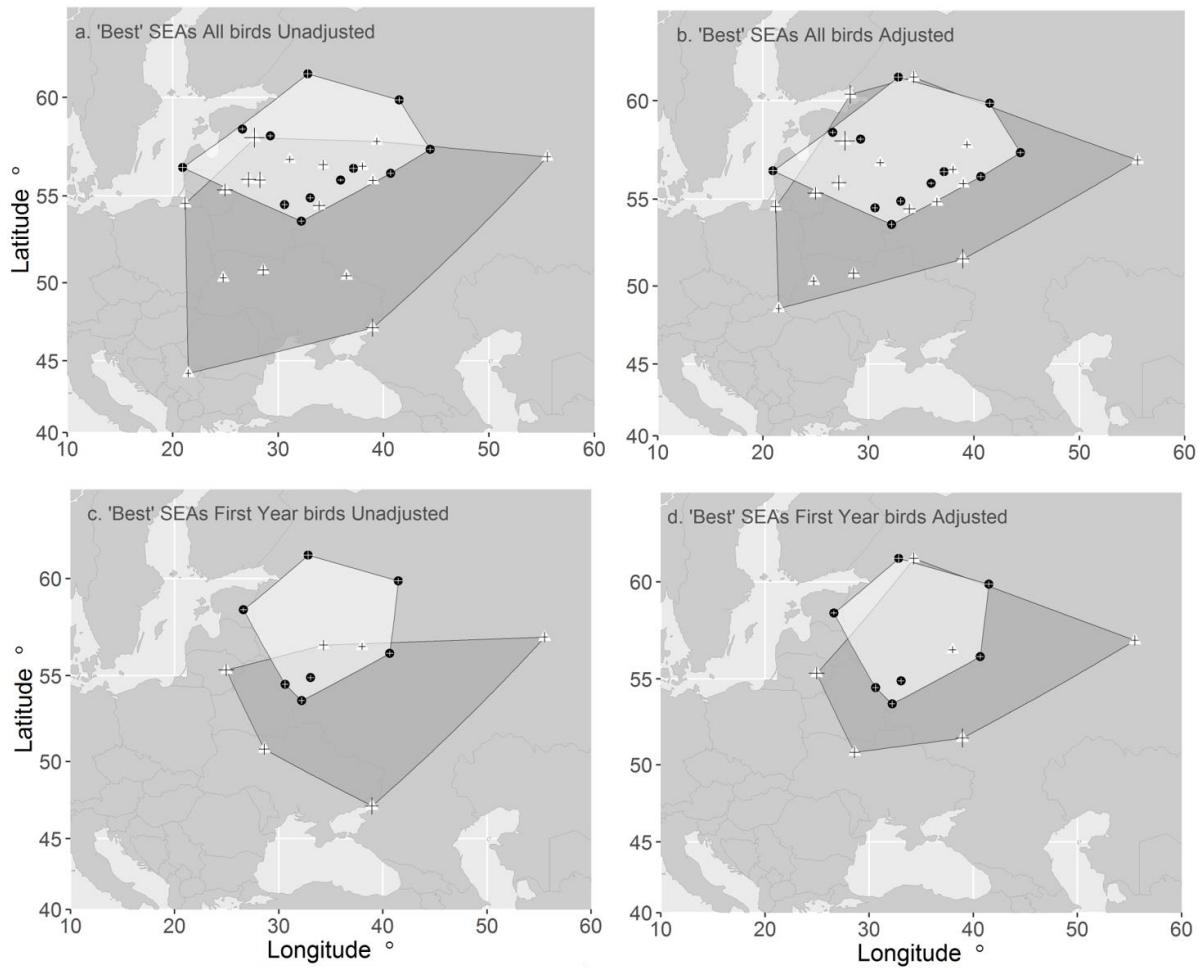
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714 Figure 4



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716 Figure 5
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721 **ONLINE ONLY SUPPLEMENTARY MATERIAL**

722

723 **Low and annually variable migratory connectivity in a long distance migrant:**
724 **Whinchats *Saxicola rubetra* may show a bet-hedging strategy**

725

726 EMMA BLACKBURN^{1,2}, MALCOLM BURGESS³, BENEDICTUS FREEMAN^{2,6}, ALICE RISELY⁴, ARIN
727 IZANG², SAM IVANDE^{1,2}, CHRIS HEWSON⁵ & WILL CRESSWELL^{1*,2}

728

729 ¹*Centre for Biological Diversity, University of St Andrews, Harold Mitchell Building, St Andrews, Fife
730 KY16 9TH, UK*

731 ²*AP Leventis Ornithological Research Institute, Jos, Nigeria*

732 ³*RSPB Centre for Conservation Science, The Lodge, Sandy, Beds SG19 2DL, UK*

733 ⁴*Centre for Integrative Ecology, Deakin University, Geelong, VIC 3216, Australia*

734 ⁵*British Trust for Ornithology, The Nunnery, Thetford, IP24 2PU UK*

735 ⁶*Fauna & Flora International (FFI), 2 Harmon's Compound, Oldest Congo Town, Monrovia, Liberia*

736 *Correspondence author: wrlc@st-and.ac.uk

737

738 **Table S1.** Observed area and overlap (million square kilometres) of breeding locations compared to a
 739 random sample (lower and upper 95% CI: random low and random up respectively in table), and
 740 mean distance (\pm 1 SE km) between breeding locations for all birds (a) and first year birds (b) with
 741 different values of sun elevation angle (SEA).

a. ALL	Overall	2013	2014	Diff	Random low	Random up	P value
1. BEST SUMMER SEA	N = 29	N = 12	N = 17				
Area	2.549	0.651	2.157	1.506	-0.612	1.225	0.01
Overlap	14.7				32.4	76.2	0.01
Mean distance		604 \pm 18	869 \pm 33	$t_{28} = 5.2$			0.0001
2. WINTER SEA							
Area	2.332	0.703	2.230	1.527	-0.586	1.063	0.01
Overlap	27.1				34.3	77.9	0.01
Mean distance		631 \pm 20	859 \pm 34	$t_{28} = 4.2$			0.0002
3. AVERAGE SEA							
Area	2.514	0.744	2.183	1.439	-0.566	1.215	0.02
Overlap	20.9				35.3	77.1	0.01
Mean distance		634 \pm 19	862 \pm 33	$t_{28} = 4.4$			0.0002
4. SUMMER SEA JUNE							
Area	2.519	0.603	2.049	1.446	-0.567	1.147	0.01
Overlap	20.9				35.3	75.2	0.01
Mean distance		593 \pm 18	903 \pm 32	$t_{28} = 6.4$			0.0001
5. MINIMUM RANGE SUMMER SEA							
Area	2.351	0.503	1.971	1.468	-0.568	1.144	0.01
Overlap	8.7				32.1	75.9	0.01
Mean distance		591 \pm 19	839 \pm 33	$t_{28} = 4.8$			0.0001
6. MAXIMUM OVERLAP SUMMER SEA							
Area	2.351	0.640	2.126	1.486	-0.520	1.061	0.01
Overlap	21.9				34.0	76.6	0.01
Mean distance		599 \pm 18	870 \pm 33	$t_{28} = 5.3$			0.0001
b. FIRST YEARS							
1. BEST SUMMER SEA	N = 13	N = 7	N = 6				
Area	1.700	0.470	1.199	0.729	-0.857	0.589	0.02
Overlap	7.9				3.8	56.8	0.06
Mean distance		598 \pm 40	1059 \pm 72	$t_{12} = 4.1$			0.0010
2. WINTER SEA							
Area	1.462	0.487	1.145	0.658	-0.706	0.590	0.04
Overlap	18.9				4.8	58.6	0.24
Mean distance		620 \pm 43	1032 \pm 73	$t_{12} = 3.6$			0.003
3. AVERAGE SEA							
Area	1.689	0.518	1.199	0.681	-0.832	0.622	0.05
Overlap	11.6				3.9	60.2	0.10
Mean distance		640 \pm 41	1060 \pm 72	$t_{12} = 3.7$			0.003
4. SUMMER SEA JUNE							
Area	1.672	0.429	1.153	0.681	-0.78	0.577	0.03
Overlap	2.7				4.6	59.3	0.04
Mean distance		567 \pm 40	1013 \pm 70	$t_{12} = 3.3$			0.007

742 **Table S2.** Observed area and overlap (million square kilometres) of breeding locations compared to a
 743 random sample (lower and upper 95% CI: random low and random up respectively in table), and
 744 mean distance (\pm 1 SE km) between breeding locations for all birds (a.) and first year birds (b.) with
 745 different values of SEA, latitudes adjusted for the effect of no light stalks on four birds in 2014.
 746 Latitude adjustments for each SEA value analysis are given in italics. Sample sizes as in Table 1.

a. ALL	Overall	2013	2014	Diff	Random low	Random up	P value
1. BEST SUMMER SEA <i>Light stalk adjustment for latitude -4.45 ± 1.6, $F_{1,26} = -2.7$, $P = 0.010$</i>							
Area	1.681	0.651	1.650	0.999	-0.440	0.814	0.01
Overlap	36.5				35.2	77.8	0.06
Mean distance		604 ± 18	820 ± 29	$t_{28} = 4.5$			0.0001
2. WINTER SEA <i>Light stalk adjustment for latitude -3.75 ± 1.7, $F_{1,26} = -2.2$, $P = 0.034$</i>							
Area	1.737	0.703	1.727	1.024	-0.523	0.871	0.02
Overlap	41.1				35.2	77.7	0.07
Mean distance		631 ± 20	828 ± 30	$t_{28} = 3.9$			0.0005
3. AVERAGE SEA <i>Light stalk adjustment for latitude -4.31 ± 1.7, $F_{1,26} = -2.6$, $P = 0.014$</i>							
Area	1.754	0.744	1.721	0.977	-0.385	0.869	0.02
Overlap	41.7				34.4	77.7	0.07
Mean distance		634 ± 19	817 ± 29	$t_{28} = 3.8$			0.0007
4. SUMMER SEA JUNE <i>Light stalk adjustment for latitude -5.02 ± 1.5, $F_{1,26} = -3.4$, $P = 0.002$</i>							
Area	1.835	0.603	1.814	1.211	-0.454	0.897	0.01
Overlap	31.1				35.2	76.0	0.03
Mean distance		593 ± 18	824 ± 30	$t_{28} = 4.7$			0.0001
b. FIRST YEARS							
1. BEST SUMMER SEA <i>Adjustment as for 1. above</i>							
Area	1.275	0.470	1.232	0.762	-0.651	0.486	0.01
Overlap	33.2				4.2	60.3	0.51
Mean distance		598 ± 40	1042 ± 63	$t_{12} = 4.3$			0.001
2. WINTER SEA <i>Adjustment as for 2. above</i>							
Area	1.22	0.487	1.176	0.689	-0.605	0.442	0.01
Overlap	39.4				3.6	58.8	0.71
Mean distance		620 ± 43	1028 ± 66	$t_{12} = 3.8$			0.0020
3. AVERAGE SEA <i>Adjustment as for 3. above</i>							
Area	1.327	0.518	1.227	0.709	-0.66	0.449	0.01
Overlap	33.4				2.2	60.1	0.52
Mean distance		640 ± 41	1040 ± 63	$t_{12} = 3.8$			0.002
4. SUMMER SEA JUNE <i>Adjustment as for 4. above</i>							
Area	1.348	0.429	0.829	0.400	-0.652	0.521	0.12
Overlap	3.3				4.1	58.5	0.04
Mean distance		567 ± 40	950 ± 65	$t_{12} = 3.1$			0.01

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749 **Table S3:** Full models (left) to investigate whether there was any bias in latitude due to age, sex,
 750 absence of light stalks, SEA value used (if varied in a model, i.e. not when average SEA value was
 751 used for all birds), and whether the SEA value used was estimated using the H-E calibration or an
 752 average (if this was relevant in a model), controlling for year. Final models (right) with the only
 753 significant term retained (light stalk absence) controlling for year to determine the final adjustment in
 754 latitude due to four tags in 2014 having no light stalks.

	Full Model				Final Model			
	Est.	Std. Error	t value	P value	Est.	Std. Error	t value	P value
BEST SUMMER SEA								
(Intercept) Adult Female 2013	55.7	6.2	9.0	<0.0001	56.9	0.88	64.9	<0.0001
Age first year	-0.88	1.9	-0.5	0.64				
Sex Male	-0.25	2.2	-0.1	0.91				
Year 2014	2.4	12.9	0.2	0.85	-1.6	1.2	-1.3	0.20
With light stalk	-4.3	2.2	-2.0	0.067	-4.5	1.6	-2.7	0.10
Summer SEA	-0.48	1.7	-0.3	0.77				
Average SEA used YES	0.79	2.3	0.3	0.74				
Year 2014 * Summer SEA	1.4	3.2	0.4	0.67				
Year 2014 * Av SEA YES	1.3	3.1	0.4	0.67				
$F_{8,17} = 1.4, P = 0.27, R^2 = 0.11$					$F_{2,26} = 7.1, P = 0.004, R^2 = 0.30$			
WINTER SEA								
(Intercept) Adult Female 2013	41.4	14.4	2.9	0.009	55.3	0.9	60.7	<0.0001
Age first year	0.49	1.6	0.3	0.76				
Sex Male	0.15	2.1	0.1	0.94				
Year 2014	26.5	17.6	1.5	0.15	-0.31	1.3	-0.2	0.81
With light stalk	-5.0	2.3	-2.2	0.040	-3.7	1.7	-2.2	0.034
Winter SEA	-2.7	2.8	-1.0	0.35				
Year 2014 * Winter SEA	5.9	3.8	1.5	0.14				
$F_{6,19} = 1.3, P = 0.30, R^2 = 0.07$					$F_{2,26} = 3.2, P = 0.057, R^2 = 0.14$			
AVERAGE SEA								
(Intercept) Adult Female 2013	56.4	2.7	20.6	<0.0001	56.3	0.9	62.7	<0.0001
Age first year	0.16	1.6	0.1	0.92				
Sex Male	-0.20	2.1	-0.1	0.92				
Year 2014	-1.0	1.6	-0.6	0.55	-0.97	1.3	-0.8	0.45
With light stalk	-4.4	2.1	-2.1	0.044	-4.3	1.7	-2.6	0.015
$F_{4,21} = 2.1, P = 0.11, R^2 = 0.15$					$F_{2,26} = 5.2, P = 0.012, R^2 = 0.23$			
SUMMER SEA JUNE								
(Intercept) Adult Female 2013	52.8	5.4	9.7	<0.0001	56.7	0.8	71.5	<0.0001
Age first year	-0.52	1.6	-0.32	0.75				
Sex Male	2.0	2.0	1.0	0.32				

Year 2014	2.3	11.4	0.2	0.84	-2.6	1.1	-2.3	0.030
With light stalk	-3.9	1.9	-2.1	0.055	-5.0	1.5	-3.4	0.002
Summer SEA	-0.53	1.5	-0.4	0.72				
Average SEA used YES	1.3	2.0	0.6	0.52				
Year 2014 * Summer SEA	1.3	2.9	0.5	0.65				
Year 2014 * Av SEA YES	-0.44	2.7	-0.2	0.88				

$$F_{8,17} = 2.8, P = 0.038, R^2 = 0.36 \quad F_{2,26} = 13.6, P < 0.0001, R^2 = 0.47$$

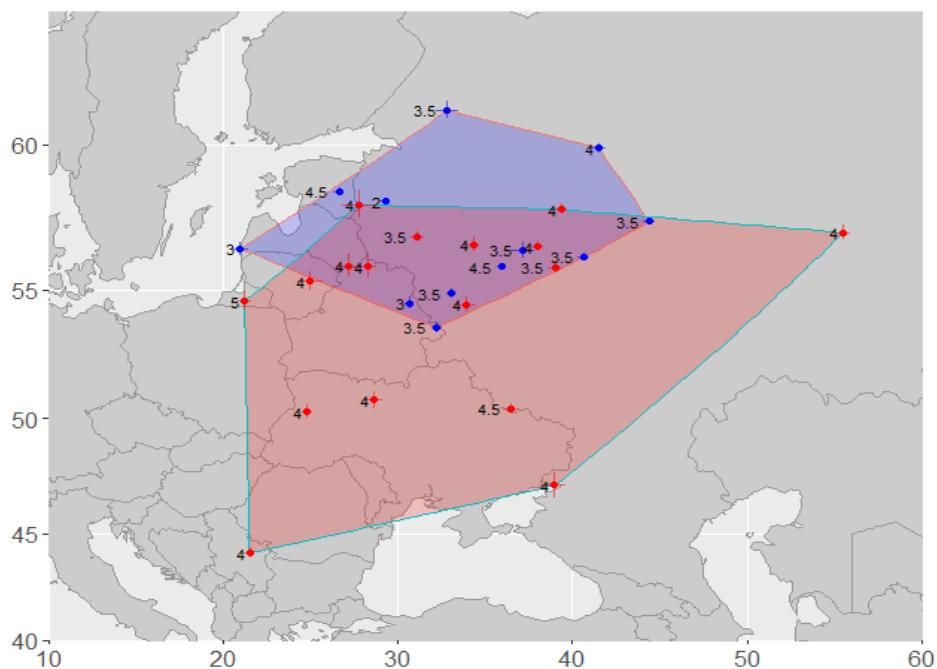
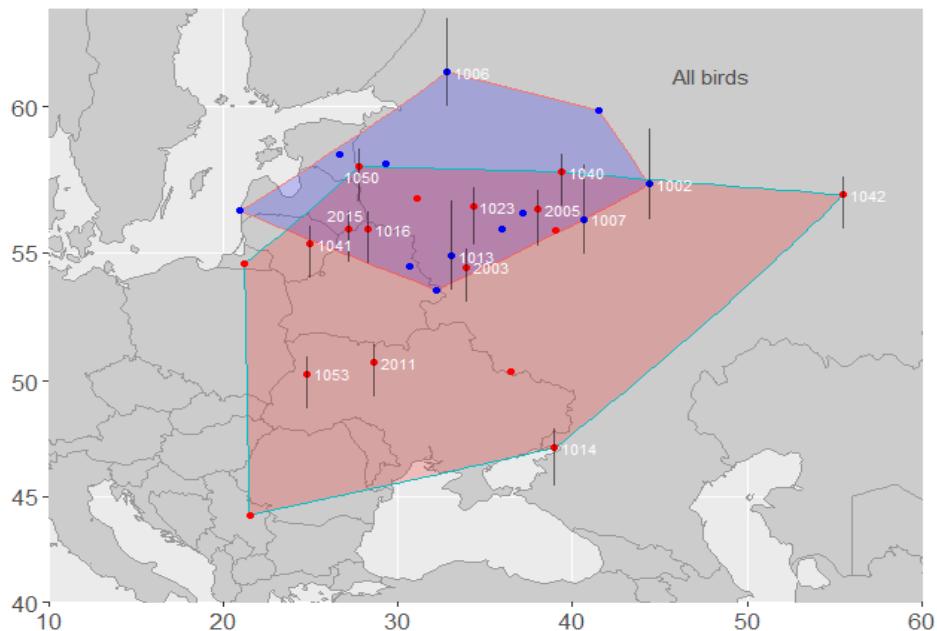
755

756 **Table S4:** Models to investigate whether there was any bias in longitude due to age, sex, absence of
 757 light stalks, SEA value used (if varied in a model, i.e. not when average SEA value was used for all
 758 birds), and whether the SEA value used was estimated using the H-E calibration or an average (if this
 759 was relevant in a model), controlling for year.

BEST SUMMER SEA	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	21.7	14.9	1.5	0.17
Age first year	3.0	4.5	0.7	0.52
Sex Male	0.19	5.4	0.04	0.97
Year 2014	33.2	31.3	1.1	0.30
With light stalk	-0.27	5.3	-0.1	0.96
Summer SEA	-2.1	4.0	-0.5	0.60
Average SEA used YES	6.2	5.6	1.1	0.28
Year 2014 * Summer SEA	8.2	7.9	1.0	0.31
Year 2014 * Av SEA YES	-4.4	7.5	-0.6	0.57
$F_{8,17} = 0.5, P = 0.82, R^2 = 0$				
WINTER SEA	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	48.6	35.9	1.4	0.19
Age first year	4.1	4.1	1.0	0.33
Sex Male	0.25	5.2	0.05	0.96
Year 2014	1.8	44.0	0.04	0.97
With light stalk	-2.9	5.7	-0.5	0.61
Winter SEA	3.7	7.1	0.5	0.61
Year 2014 * Winter SEA	0.82	9.5	0.1	0.93
$F_{6,19} = 0.4, P = 0.85, R^2 = 0$				
AVERAGE SEA	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	30.6	6.7	4.6	0.0002
Age first year	4.5	3.9	1.2	0.26
Sex Male	0.033	5.0	0.01	0.99
Year 2014	0.64	3.9	0.2	0.87
With light stalk	-1.1	5.0	-0.2	0.83
$F_{4,21} = 0.5, P = 0.75, R^2 = 0$				
SUMMER SEA JUNE	Est.	Std. Error	t value	P value
(Intercept) Adult Female 2013	20.9	15.4	1.4	0.19
Age first year	3.7	4.7	0.8	0.44
Sex Male	1.2	5.5	0.2	0.84
Year 2014	33.5	32.3	1.0	0.31
With light stalk	-0.25	5.4	-0.05	0.96
Summer SEA	-2.0	4.1	-0.5	0.64
Average SEA used YES	6.2	5.8	1.1	0.30
Year 2014 * Summer SEA	8.3	8.2	1.0	0.32
Year 2014 * Av SEA YES	-5.1	7.8	-0.7	0.52
$F_{8,17} = 0.5, P = 0.81, R^2 = 0$				

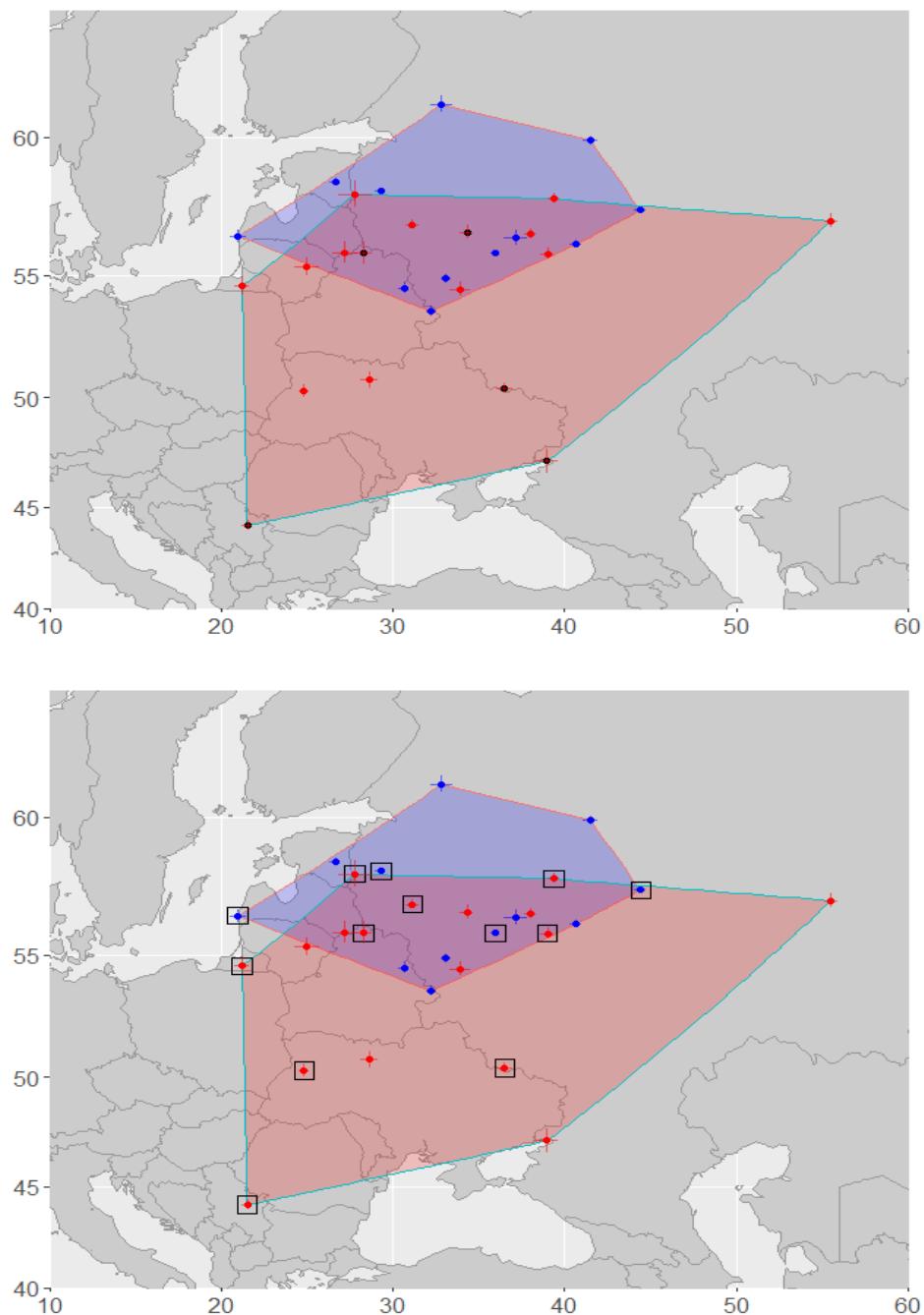
760 **Figure S1:** Top - Loggers where an average summer SEA was used labelled in white, showing the
761 range of SEAs (black vertical lines) for all other loggers where the calibration was successful for that
762 year. Bottom – the SEA (sun elevation angle used for each logger in the “best” Analysis 1. Blue 2013,
763 Red 2014.

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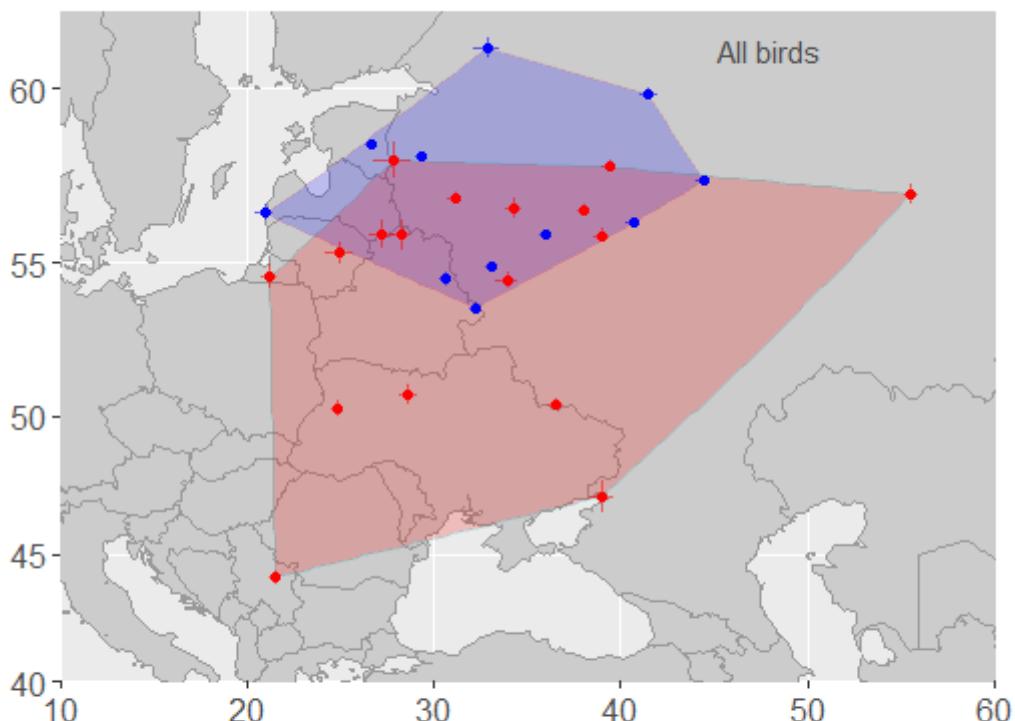
766 **Figure S2:** Top - Loggers in 2014 without a light stalk plotted in black (and so where breeding
767 location latitudes were adjusted northwards according to the models in Table S1). Bottom - Loggers of
768 adult birds plotted within black squares

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771 **Figure S3:** All birds summer unadjusted (*BEST SUMMER SEA* Table S1.a.1). Blue 2013, Red 2014; +/- 2
772 standard errors in latitude and longitude are plotted for each location.

773 2013 0.651; 2014 2.157 million square km (mskm)
774 604 (+/- 18) vs 869 (+/- 33) km ($t_{28} = 5.2$, $P = 0.0001$)
775 Area diff = 1.506 mskm (99th percentile)
776 Random -0.612 to 1.225 mskm
777 Overall 2.549 mskm = 26.2% Europe land area
778 14.7% (99th percentile)
779 Random 32.4% - 76.2%

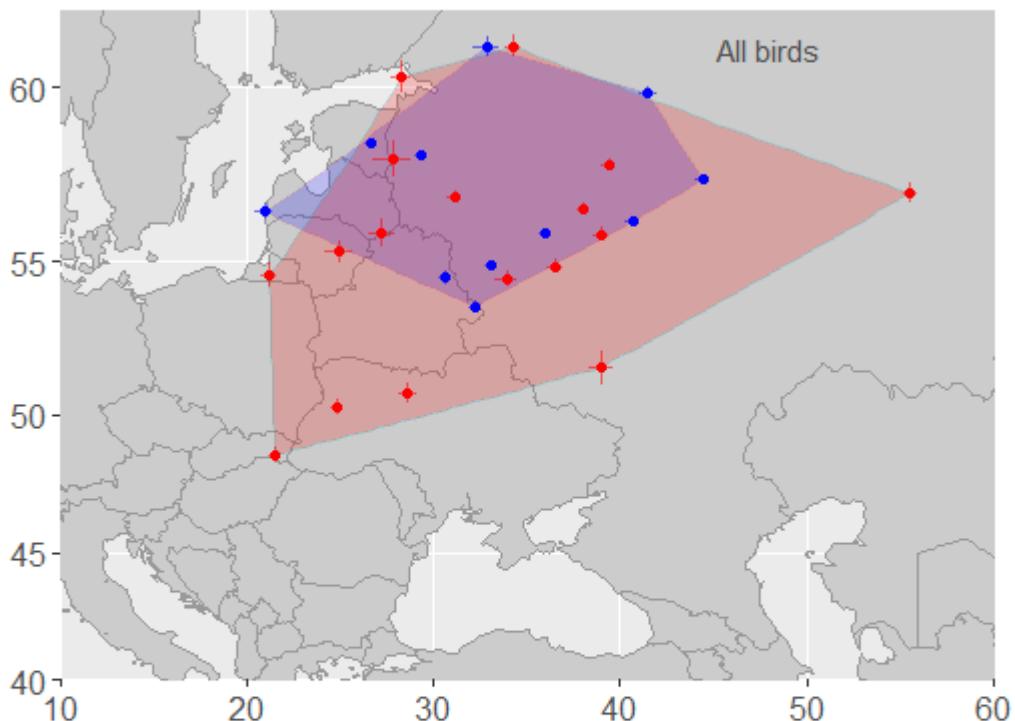


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781

782 **Figure S4:** All birds summer adjusted (*BEST SUMMER SEA* Table S2.a.1). Blue 2013, Red 2014; +/- 2
783 standard errors in latitude and longitude are plotted for each location.

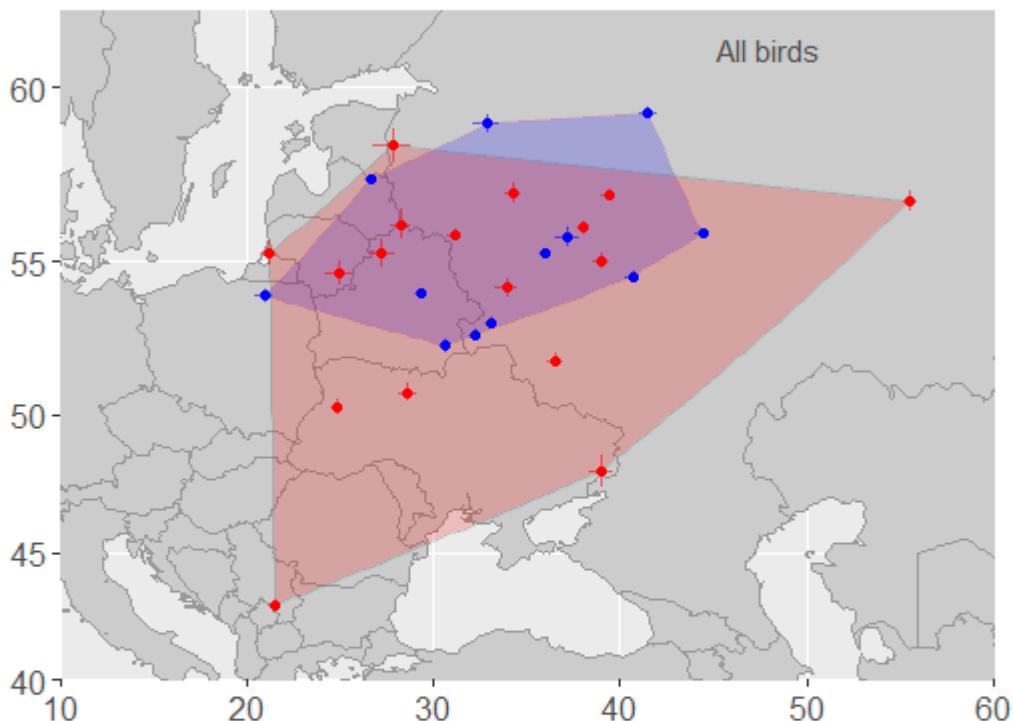
784 2013 0.651; 2014 1.650 million square km
785 604 (+/- 18) vs 820 (+/- 29) km ($t_{28} = 4.5$, $P = 0.0001$)
786 Area diff = 0.999 mskm (99th percentile)
787 Random -0.440 to 0.814 mskm
788 Overall 1.681 mskm = 16.5%
789 37.9% (94th percentile)
790 Random 35.2% - 77.8%



791

792 **Figure S5:** All birds winter unadjusted (*WINTER SEA* Table S1.a.2). Blue 2013, Red 2014; +/- 2
793 standard errors in latitude and longitude are plotted for each location.

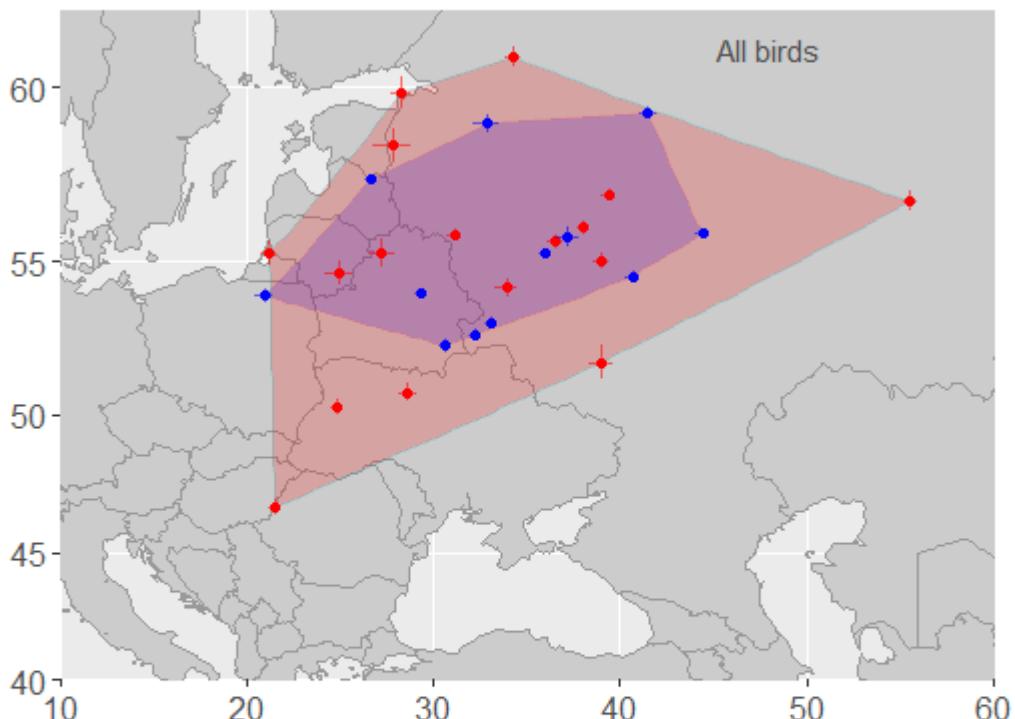
794 2013 0.703; 2014 2.230 million square km
795 631 (+/- 20) vs 859 (+/- 34) km ($t_{28} = 4.2$, $P = 0.0002$)
796 Area diff = 1.527 mskm (99th percentile)
797 Random -0.586 to 1.063 mskm
798 Overall 2.332 mskm = 22.9%
799 27.1% (99th percentile)
800 Random 34.3% - 77.9%



801

802 **Figure S6:** All birds winter adjusted (*WINTER SEA* Table S2.a.2). Blue 2013, Red 2014; +/- 2 standard
803 errors in latitude and longitude are plotted for each location.

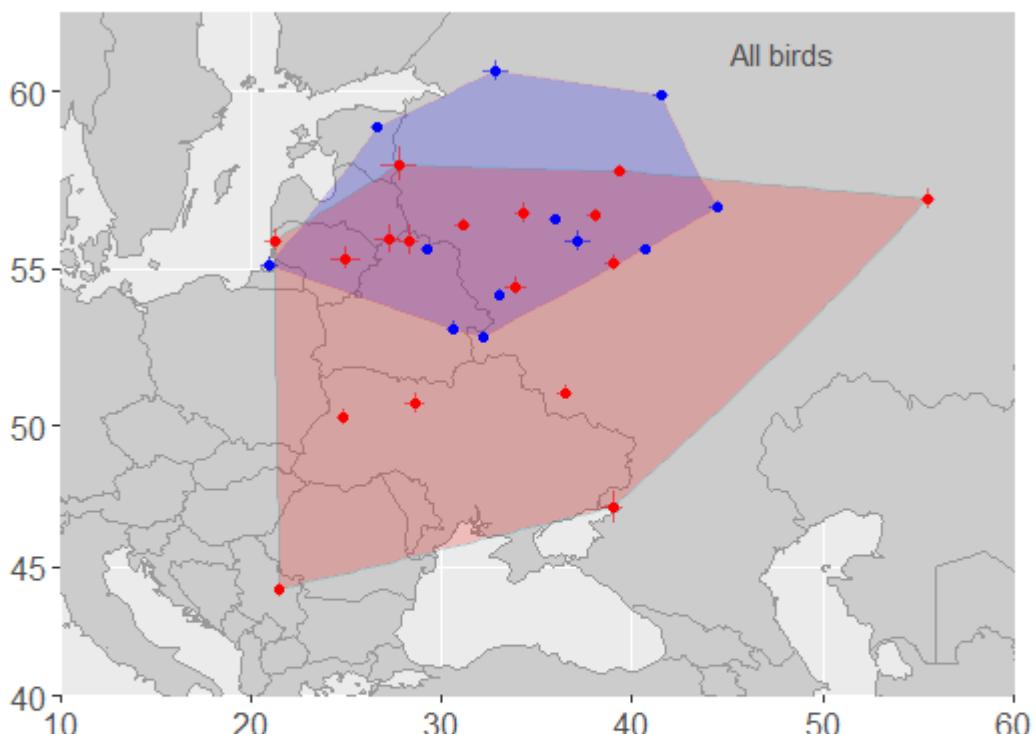
804 2013 0.703; 2014 1.727 million square km
805 631 (+/- 20) vs 828 (+/- 30) km ($t_{28} = 3.9$, $P = 0.0005$)
806 Area diff = 1.024 mskm (98th percentile)
807 Random -0.523 to 0.871
808 Overall 1.737 mskm = 17.0%
809 41.1% (93rd percentile)
810 Random 35.2% - 77.7%



811

812 **Figure S7:** All birds average of SEA angles from 2 to 6, unadjusted (AVERAGE SEA Table S1.a.3). Blue
813 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

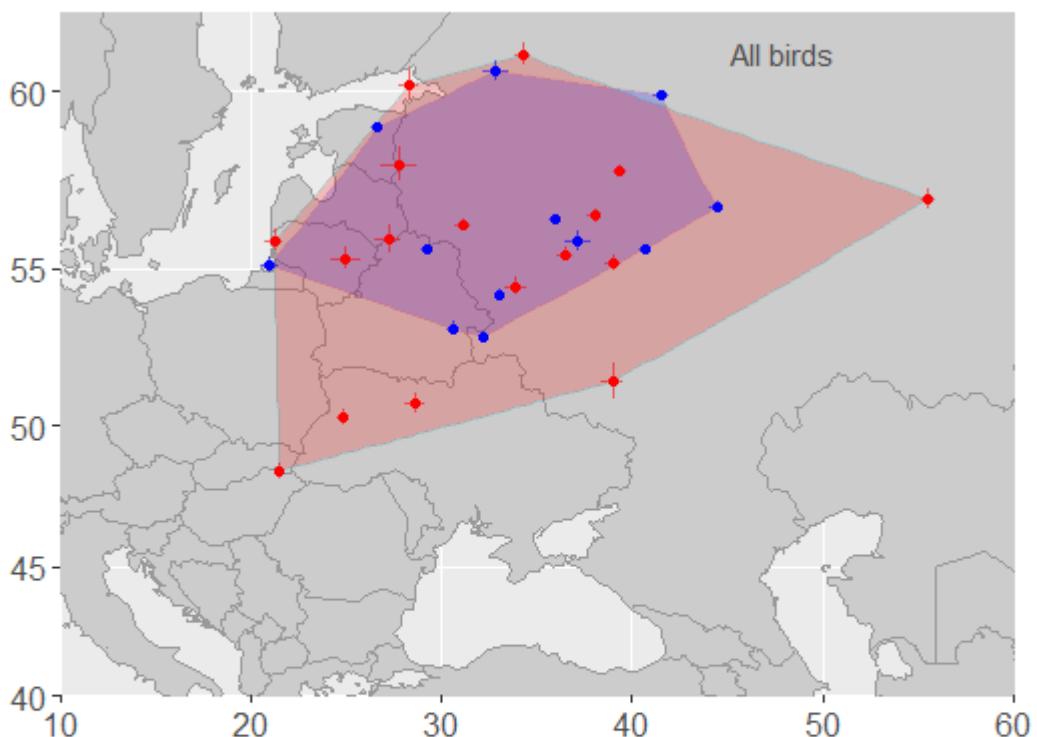
814 2013 0.744; 2014 2.183 million square km
815 634 (+/- 19) vs 862 (+/- 33) km ($t_{28} = 4.4$, $P = 0.0002$)
816 Area diff = 1.439 mskm (98th percentile)
817 Random -0.566 to 1.215 mskm
818 Overall 2.514 mskm = 24.7%
819 20.9% (<99th percentile)
820 Random 35.3% - 77.1%



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823 **Figure S8:** All birds average from 2 to 6, adjusted (*AVERAGE SEA* Table S1.a.3). Blue 2013, Red 2014;
824 +/- 2 standard errors in latitude and longitude are plotted for each location.

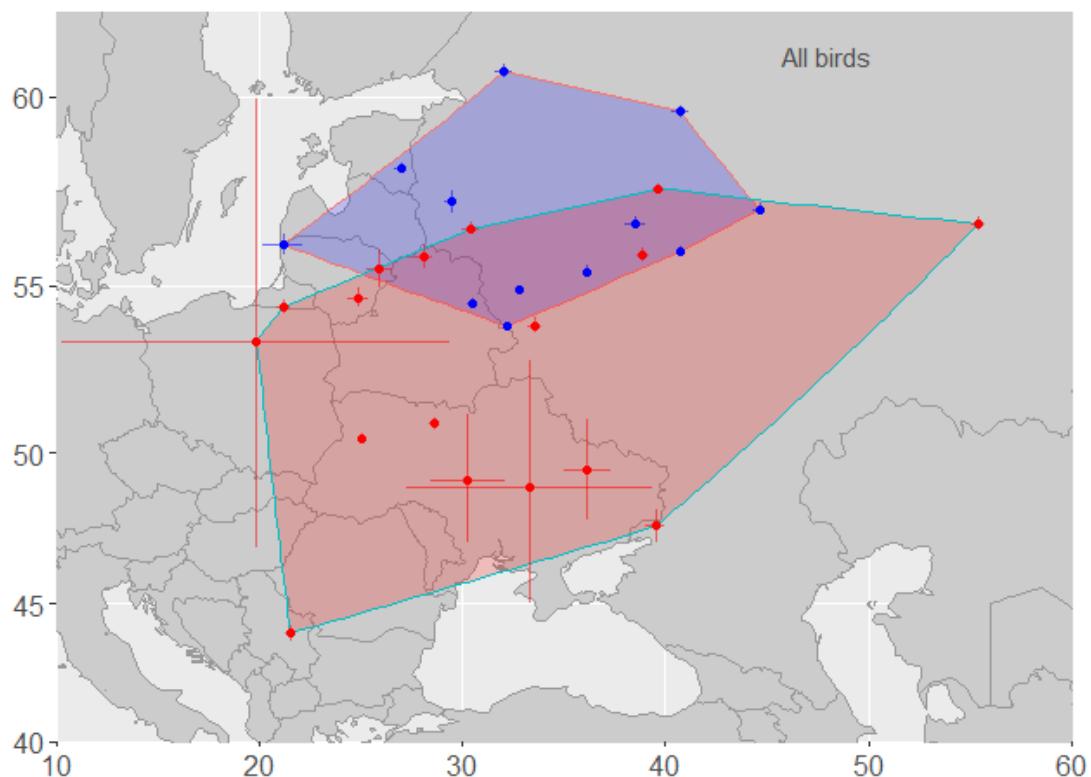
825 2013 0.744; 2014 1.721 million square km
826 634 (+/- 19) vs 817 (+/- 29) km ($t_{28} = 3.8$, $P = 0.0007$)
827 Area diff = 0.977 mskm (98th percentile)
828 Random -0.385 to 0.869 mskm
829 Overall 1.754 mskm = 17.2%
830 41.7% (93rd percentile)
831 Random 34.4% - 77.7%



832

833 **Figure S9:** All birds summer unadjusted; all June locations (*SUMMER SEA JUNE* Table S1.a.4). Blue
834 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

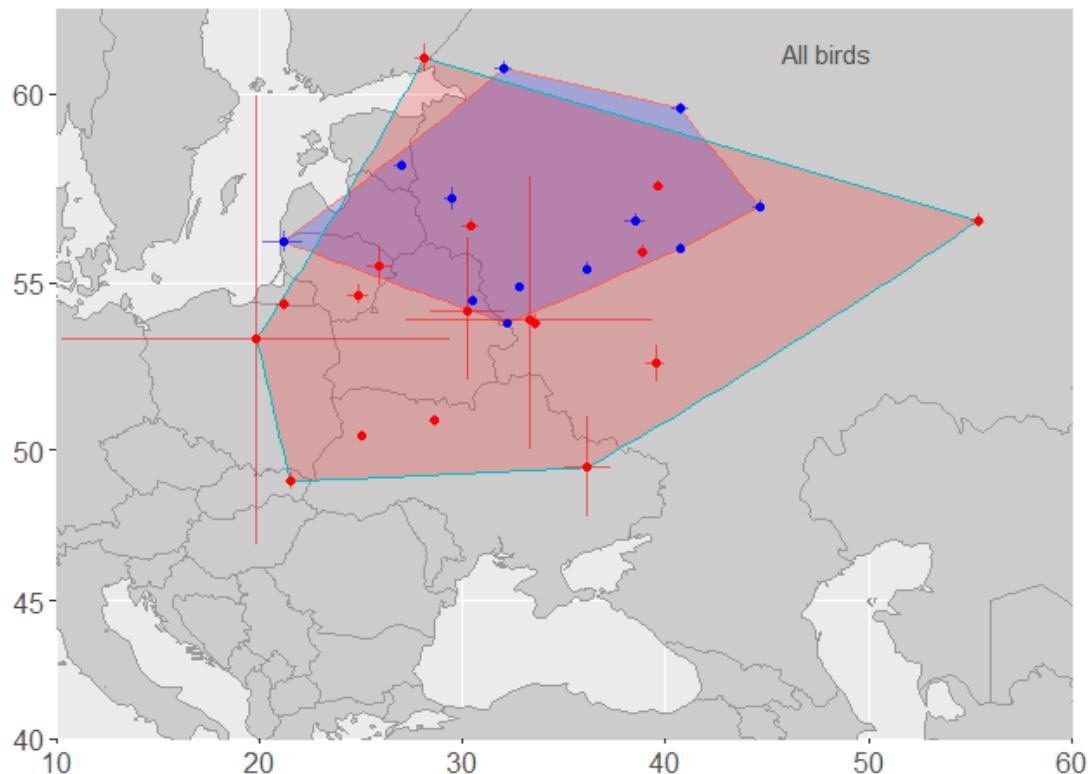
835 2013 0.603; 2014 2.049 million square km
836 593 (+/- 18) vs 903 (+/- 32) km ($t_{28} = 6.4$, $P = 0.0001$)
837 Area diff = 1.446 mskm (99th percentile)
838 Random -0.567 to 1.147mskm
839 Overall 2.519 mskm = 24.7%
840 20.9% (99th percentile)
841 Random 35.3 to 75.2%



842

843 **Figure S10:** All birds summer adjusted; all June locations (*SUMMER SEA JUNE* Table S2.a.4). Blue
844 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

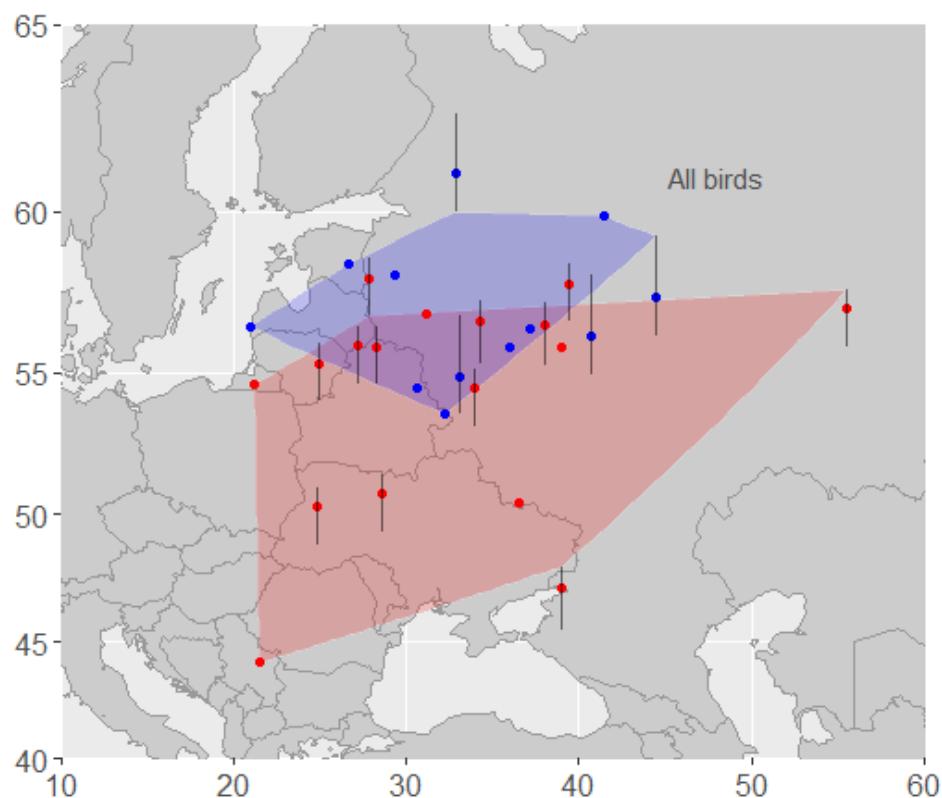
845 2013 0.603; 2014 1.814 million square km
846 593 (+/- 18) vs 824 (+/- 30) km ($t_{28} = 4.7$, $P = 0.0001$)
847 Area diff = 1.211 mskm (99th percentile)
848 Random -0.454 to 0.897 mskm
849 Overall 1.835 mskm = 18.0%
850 31.1% (97th percentile)
851 Random 35.2 to 76.0%



852

853 **Figure S11:** All birds summer SEA with SEA set to minimise total range for birds where H-E
854 calibration did not give an individual value (*MINIMUM RANGE SUMMER SEA* Table S1.a.5). Blue 2013, Red
855 2014; +/- 2 standard errors in latitude and longitude are plotted for each location. The black vertical
856 lines show the range of SEAs observed for all other loggers where the calibration was successful for
857 that year.

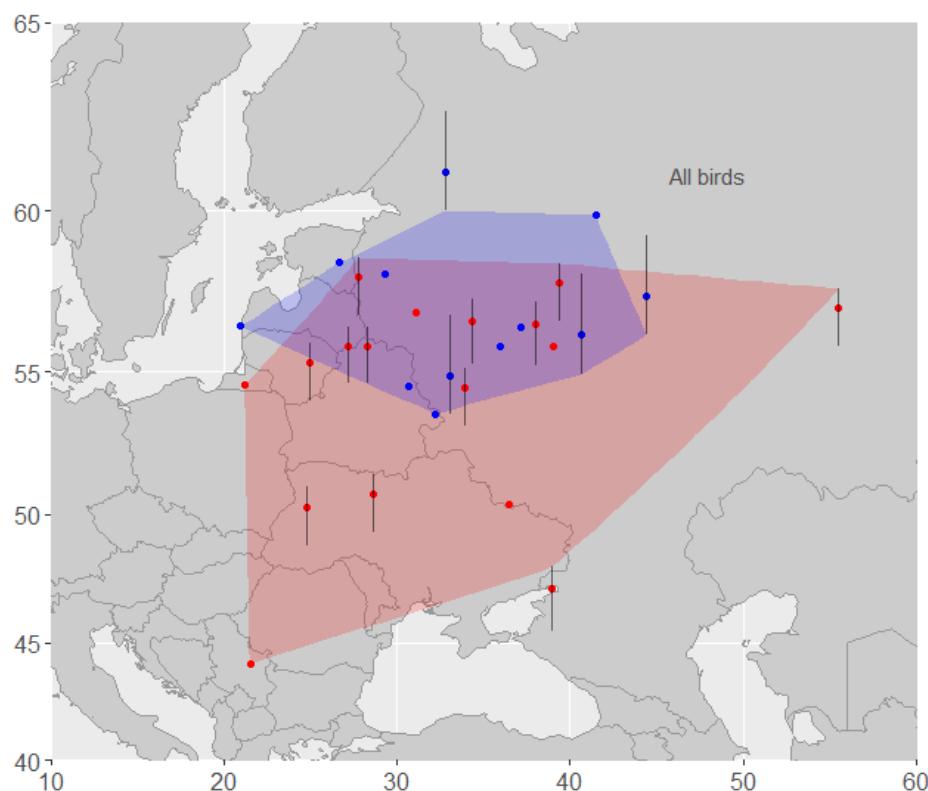
858 2013 0.503; 2014 1.971 million square km
859 591 (+/- 19) vs 839 (+/- 33) km ($t_{28} = 4.8$, $P = 0.0001$)
860 Area diff 1.468 mskm (99th percentile)
861 Random to -0.568 to 1.144 mskm
862 Overall 2.351 mskm = 23.1%
863 8.7% (99th percentile)
864 Random 32.1% to 75.9%



865

866 **Figure S12:** All birds summer SEA with SEA set at maximum to maximise range overlap for birds
867 where H-E calibration did not give an individual value (*MAXIMUM OVERLAP SUMMER SEA* Table S1.a.6).
868 Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location. The
869 black vertical lines show the range of SEAs observed for all other loggers where the calibration was
870 successful for that year.

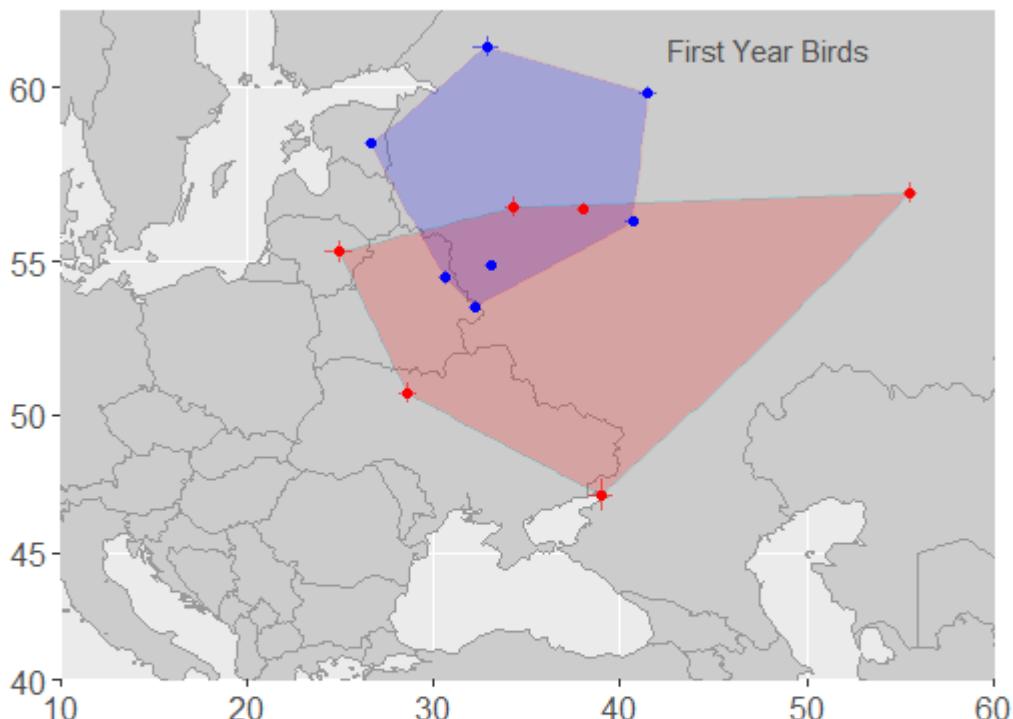
871 2013 0.640; 2014 2.126 million square km
872 599 (+/- 18) vs 870 (+/- 33) km ($t_{28} = 5.3$, $P = 0.0001$)
873 Area diff 1.486 mskm (99th percentile)
874 Random to -0.520 to 1.061 mskm
875 Overall 2.351 mskm = 23.1%
876 21.9% (99th percentile)
877 Random 34.0% to 76.6%



878

879 **Figure S13:** First year birds summer unadjusted (*BEST SUMMER SEA* Table S1.b.1). Blue 2013, Red
880 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

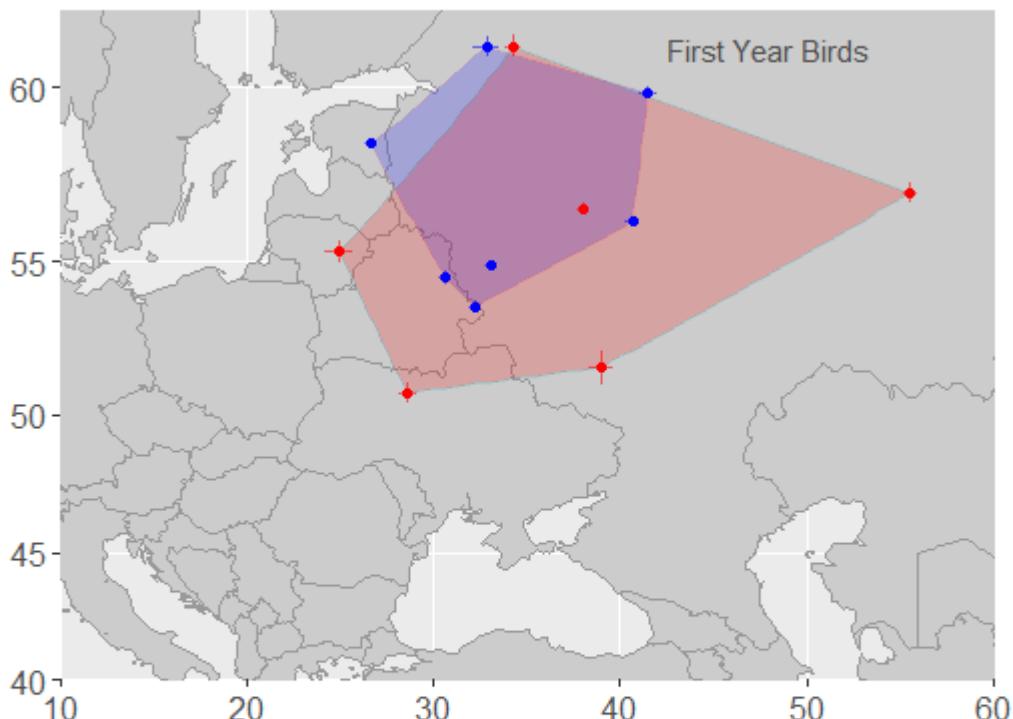
881 2013 0.470; 2014 1.199 million square km
882 598 (+/- 40) vs 1059 (+/- 72) km ($t_{12} = 4.1$, $P = 0.001$)
883 Area diff = 0.729 mskm (98th percentile)
884 Random -0.857 to 0.589 mskm
885 Overall 1.700 mskm = 16.7%
886 7.9% (94th percentile)
887 Random 3.8% - 56.8%



888

889 **Figure S14:** First year birds summer adjusted (*BEST SUMMER SEA* Table S2.b.1). Blue 2013, Red 2014;
890 +/- 2 standard errors in latitude and longitude are plotted for each location.

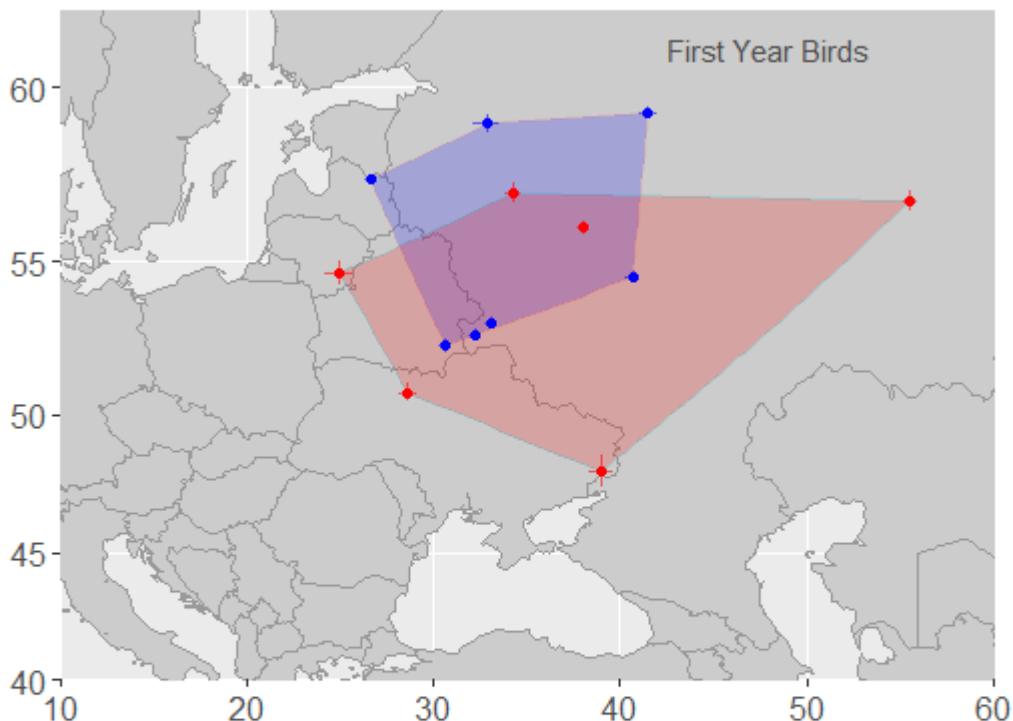
891 2013 0.470; 2014 1.232 million square km
892 598 (+/- 40) vs 1042 (+/- 63) km ($t_{12} = 4.3$, $P = 0.001$)
893 Area diff = 0.762 mskm (99th percentile)
894 Random -0.651 to 0.486 mskm
895 Overall 1.275 mskm = 13.3%
896 33.2% (49th percentile)
897 Random 4.2% - 60.3%



898

899 **Figure S15:** First year birds winter unadjusted (*WINTER SEA* Table S1.b.2). Blue 2013, Red 2014; +/- 2
900 standard errors in latitude and longitude are plotted for each location.

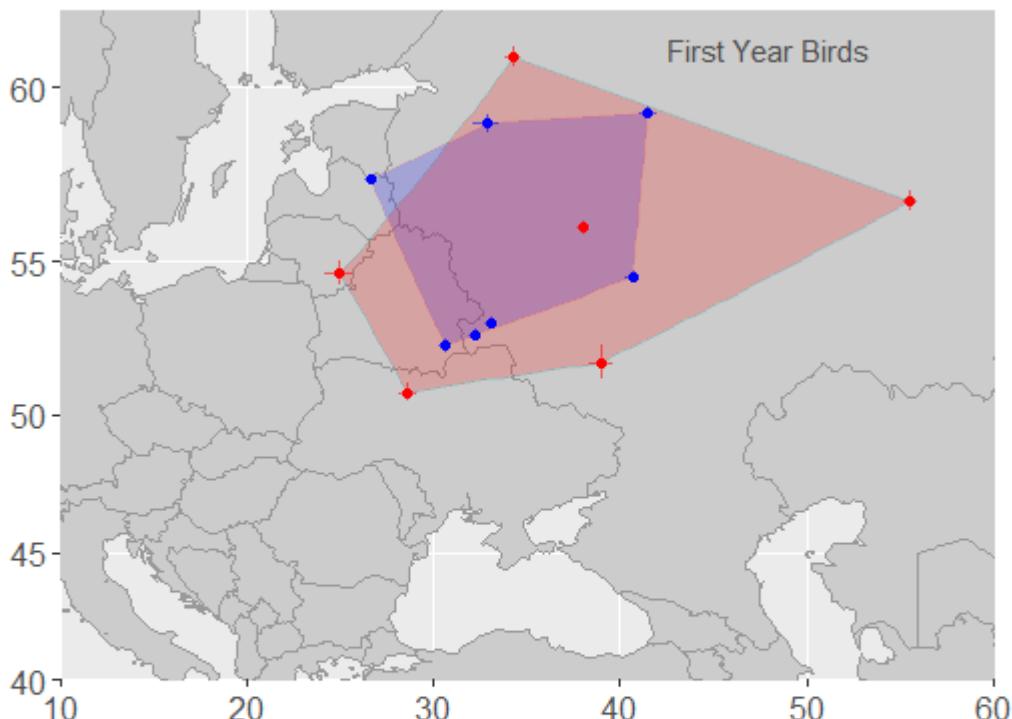
901 2013 0.487; 2014 1.145 million square km
902 620 (+/- 43) vs 1032 (+/- 73) km ($t_{12} = 3.6$, $P = 0.003$)
903 Area diff 0.658 mskm (96th percentile)
904 Random -0.706 to 0.590
905 Overall 1.462 mskm = 14.3%
906 18.9% (76th percentile)
907 Random 4.8% - 58.6%



908

909 **Figure S16:** First year birds winter adjusted (*WINTER SEA* Table S2.b.2). Blue 2013, Red 2014; +/- 2
910 standard errors in latitude and longitude are plotted for each location.

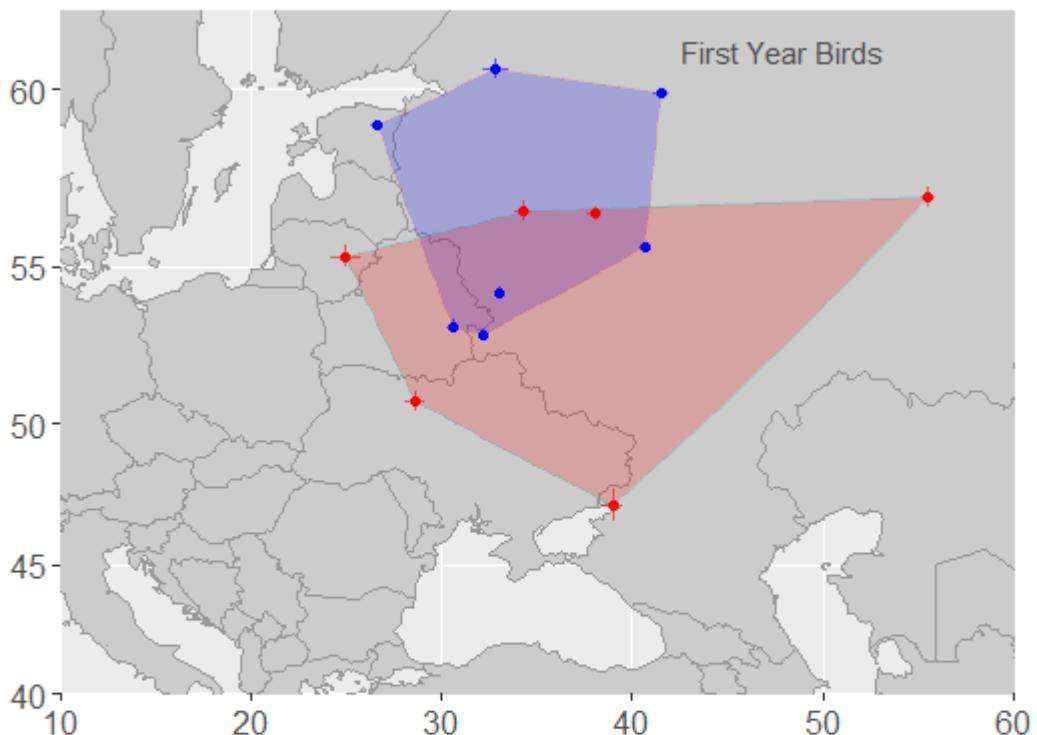
911 2013 0.487; 2014 1.176 million square km
912 620 (+/- 43) vs 1028 (+/- 66) km ($t_{12} = 3.8$, $P = 0.002$)
913 Area diff 0.689 mskm (99th percentile)
914 Random to -0.605 to 0.442
915 Overall 1.220 mskm = 12.0%
916 39.4% (29th percentile)
917 Random 3.6% - 58.8%



918

919 **Figure S17:** First year birds average unadjusted (*AVERAGE SEA* Table S1.b.3). Blue 2013, Red 2014;
920 +/- 2 standard errors in latitude and longitude are plotted for each location.

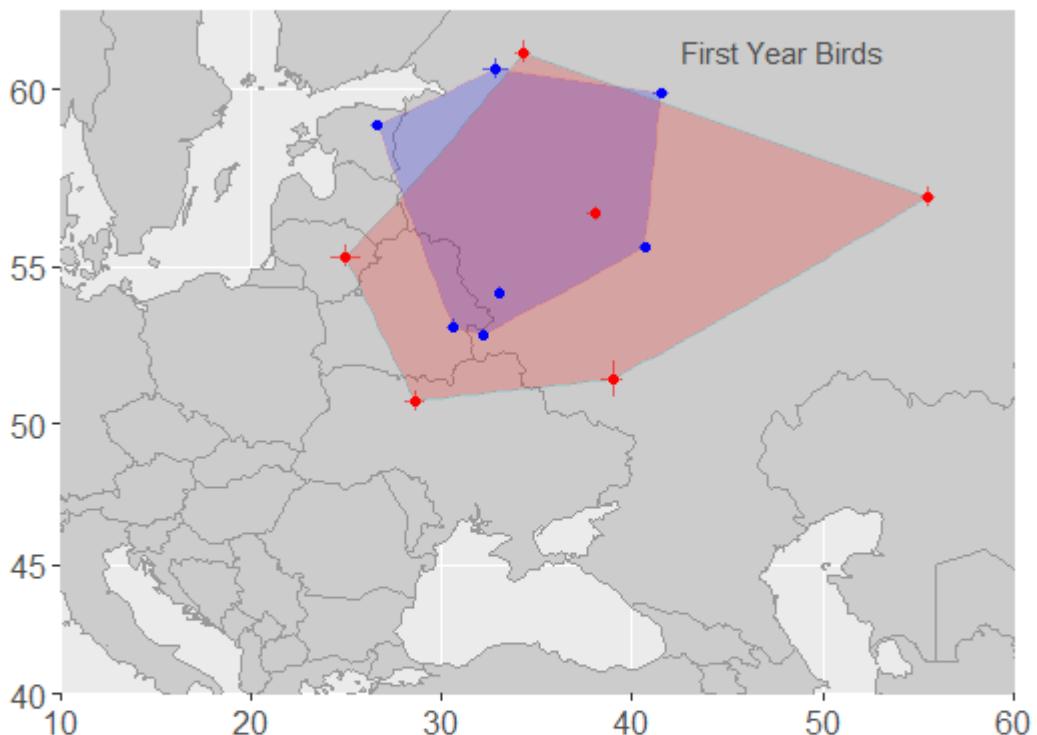
921 2013 0.518; 2014 1.199 million square km
922 640 (+/- 41) vs 1060 (+/- 72) km ($t_{12} = 3.7$, $P = 0.003$)
923 Area diff 0.681 mskm (95th percentile)
924 Random -0.832 to 0.622
925 Overall 1.689 mskm = 16.6%
926 11.6% (90th percentile)
927 Random 3.9% - 60.2%



928

929 **Figure S18:** First year birds average adjusted (AVERAGE SEA Table S2.b.3). Blue 2013, Red 2014; +/-
930 2 standard errors in latitude and longitude are plotted for each location.

931 2013 0.518; 2014 1.227 million square km
932 640 (+/- 41) vs 1040 (+/- 63) km ($t_{12} = 3.8$, $P = 0.002$)
933 Area diff 0.709 mskm (99th percentile)
934 Random -0.660 to 0.449
935 Overall 1.327 mskm = 10.0%
936 33.4% (48th percentile)
937 Random 2.2% - 60.1%



938

939 **Figure S19:** First year birds summer unadjusted; all June locations (*SUMMER SEA JUNE* Table S1.b.4).

940 Blue 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

941 2013 0.429; 2014 1.153 million square km

942 567 (+/- 40) vs 1013 (+/- 70) km ($t_{12} = 3.3$, $P = 0.007$)

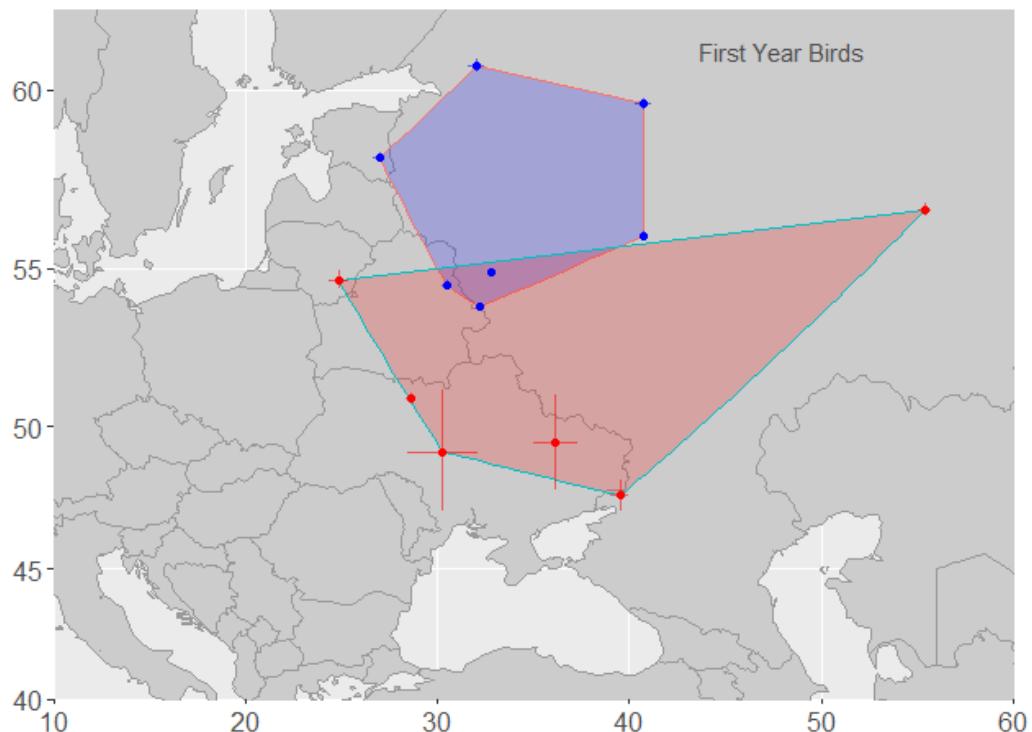
943 Area diff 0.681 mskm (97th percentile)

944 Random -0.780 to 0.577

945 Overall 1.672 mskm = 16.4%

946 2.7% (96th percentile)

947 Random 4.6% to 59.3%

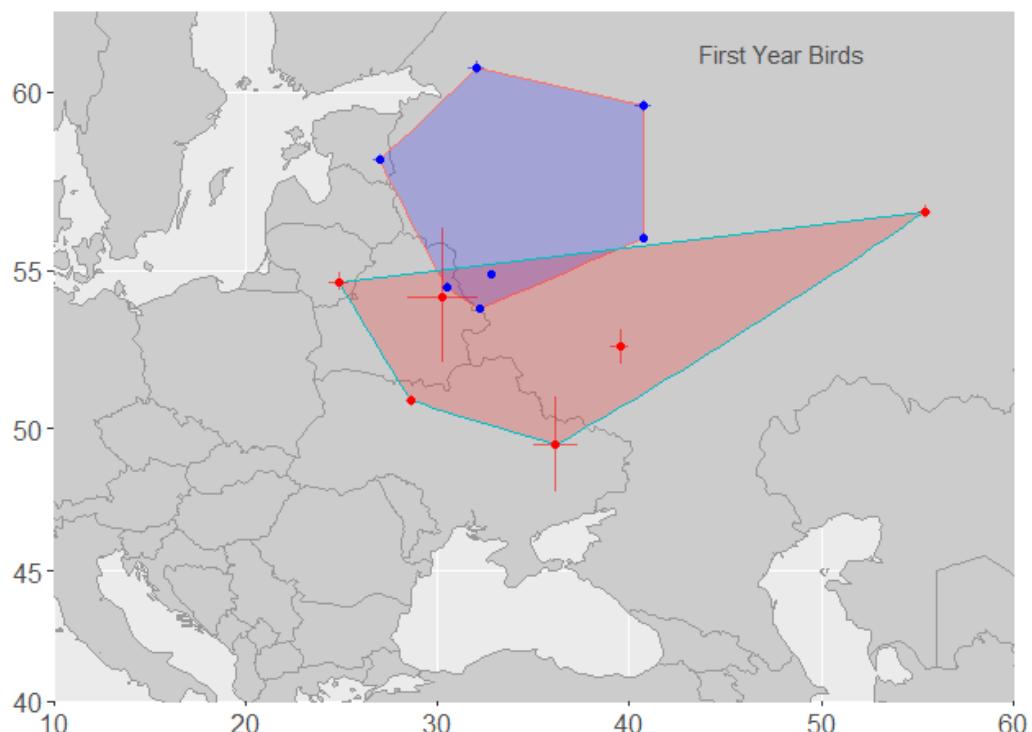


948

949

950 **Figure S20:** First year birds summer adjusted; all June locations (*SUMMER SEA JUNE* Table 2.b.4). Blue
951 2013, Red 2014; +/- 2 standard errors in latitude and longitude are plotted for each location.

952 2013 0.429; 2014 0.829 million square km
953 567 (+/- 40) vs 950 (+/- 65) km ($t_{12} = 3.1$, $P = 0.01$)
954 Area diff 0.400 mskm (88th percentile)
955 Random -0.652 to 0.521
956 Overall 1.348 mskm = 13.2%
957 3.3% (96th percentile)
958 Random 4.1% to 58.5%



959
960