Comparing the cancer potencies of emissions from vapourised nicotine products including e-cigarettes with those of tobacco smoke

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WHAT THIS PAPER ADDS

- Emissions from vapourised nicotine products (VNP) including e-cigarettes contain carcinogens but generally in lower concentrations than tobacco smoke.
- Each carcinogen contributes quantifiably to the overall cancer potency and risk
- Users and policy makers need quantitative evidence on the relative risks of cancer from the use of VNPs compared with smoking tobacco.
- Previous studies considered the individual carcinogens in an emission; here a method is developed that models the aggregate cancer potencies of all carcinogens and overcomes incompatibilities in data reporting conventions for direct comparison of the potencies and risks of tobacco smoke with VNP emissions.
- Cancer potencies span five orders of magnitude creating a spectrum ranging from uncontaminated air through VNPs to tobacco smoke.
- Most e-cigarette analyses indicate cancer potencies <1% that of tobacco smoke and <10% that of a heat-not-burn prototype although a minority of analyses indicate higher potencies.
- Highly carcinogenic emissions from e-cigarettes are avoidable, being due largely to user choice of device setting, liquid formulation and vaping behaviour, highlighting a need for increased user awareness and personal involvement in reducing risk.

ABSTRACT

Background: Quantifying relative harm caused by inhaling the aerosol emissions of vapourised nicotine products compared with smoking combustible tobacco is an important issue for public health.

Methods: The cancer potencies of various nicotine-delivering aerosols are modelled using published chemical analyses of emissions and their associated inhalation unit risks. Potencies are compared using a conversion procedure for expressing smoke and e-cigarette vapours in common units. Lifetime cancer risks are calculated from potencies using daily consumption estimates.

Results: The aerosols form a spectrum of cancer potencies spanning five orders of magnitude from uncontaminated air to tobacco smoke. E-cigarette emissions span most of this range with the preponderance of products having potencies <1% of tobacco smoke and falling within two orders of magnitude of a medicinal nicotine inhaler, however a small minority have much higher potencies. These high-risk results tend to be associated with high levels of carbonyls generated when excessive power is delivered to the atomiser coil. Samples of a prototype heat-not-burn device have lower cancer potencies than tobacco smoke by at least one order of magnitude, but higher potencies than most e-cigarettes. Mean lifetime risks decline in the sequence: combustible cigarettes ≫ heat-not-burn ≫ e-cigarettes (normal power) ≥ nicotine inhaler.

Conclusions: Optimal combinations of device settings, liquid formulation and vaping behaviour normally result in e-cigarette emissions with much less carcinogenic potency than tobacco smoke, notwithstanding there are circumstances in which the cancer risks of e-cigarette emissions can escalate, sometimes very substantially. These circumstances are usually avoidable when the causes are known.

INTRODUCTION

Opinion is divided about vapourised nicotine products (VNPs), such as e-cigarettes (ECs) which heat a nicotine-containing liquid and Heat-not-Burn (HnB) products which heat tobacco.[1, 2] The general public tends to view "vaping" ECs as equally or more harmful than smoking tobacco.[3, 4] More scientific evidence addressing the factors that lead to harm and, where possible, quantification of their effects is needed to inform scientific debates and address public uncertainty.[5]

Both cigarette smoke and VNP emissions (or "vapour") are aerosol mixtures of particulates and droplets in gaseous matrices, but each is generated by a different process from different precursors within different temperature ranges. Nicotine is not a carcinogen but VNPs transfer a burden of known carcinogens in the aerosol. Assessing the relative harmfulness of different VNPs requires meaningful indicators of harm. Common approaches include the analysis of biomarkers in bodily fluids, *in vivo* and *in vitro* toxicity studies, and evaluating the numbers of individual chemicals that exceed a specified threshold of safety.[6-10] The health risks of smoking were established half a century ago by epidemiological evidence of associated morbidity and mortality but it may be several years before such approaches can definitively assess the harm of VNPs.[11] Meanwhile, public health advice depends on quantitative chemical and toxicological approaches to model VNP risks.

This study aims to 1) derive a procedure that overcomes the problem of incompatible emissions datasets; 2) reduce the information on carcinogenic risk represented by multiple individual compounds to a single latent variable (potency) that reflects the cancer risk; and 3) calculate cancer potencies from published emissions data and compare various nicotine product risks by factoring in exposure estimates.

METHODOLOGY

Smoke and vapour emissions

Compounds from mainstream smoke analyses are typically reported in mass units (mg, μ g, ng as appropriate) per cigarette. Fewer analyses of hazardous compounds in VNPs have been published and the units reported vary widely, usually as mass of compound per volume of vapour (often as μ g per N puffs where N ranges from 1 to several hundred puffs of fixed volume). A few studies report mass units of toxicant/volume of precursor e-liquid but these are not included in this study.

Different machine smoking protocols purported to simulate human smoking adds another complication. The ISO protocol draws 35 mL per puff of smoke whereas the MDPH protocol draws 45 mL and the HCI (intense) protocol draws 55 mL (see Supplementary file).[12] Only volume is considered here, not the withdrawal rate or interval between puffs. The ISO protocol leaves cigarette filter ventilation holes unblocked whereas MDPH blocks 50% by taping over half the filter circumference, and HCI blocks these holes entirely. Ventilation blocking reduces dilution by air and increases emissions, thus smoking protocol and filter ventilation are fundamental considerations when comparing toxicant concentrations in tobacco smoke with vapour.[13]

Smoke toxicants conversion

Taking N to be the number of puffs in a smoking experiment, S the puff volume, and D the filter ventilation (i.e. fraction of air flow through filter vents) then

$$V_{k,i}^* = N_{k,i} S_k (1 - D_k) \tag{1}$$

where $V_{k,i}^*$ represents the undiluted smoke volume, and k is the machine smoking protocol for the i^{th} sample (cigarette). For the ISO protocol S=35 mL and D ranges from 0 to 1, for MDPH S=45 mL and D is half the filter ventilation, and for the HCI protocol S=55 mL and D=0.

The usefulness of the V^* parameter is illustrated in Supplementary figure S1 in which nicotine concentrations from the various smoking machine experiments are plotted against $V_{k,i}^*$.[14] The graph shows strong correlation between undiluted smoke volume V^* and nicotine concentration per cigarette (r^2 =0.93, p<0.01) with a linear regression intercept close to the origin. Similar co-linearity

of V^* over the major carcinogens under consideration provides a means of expressing smoke carcinogens as concentrations in a form independent of machine smoking protocol.

$$E_{k,i,j} = C_{k,i,j}^{tob} / V_{k,i}^* \tag{2}$$

where $C_{k,i,j}^{tob}$ (µg/cigarette) is the mass per cigarette of the j^{th} carcinogen in the i^{th} product that has been machine-smoked using protocol k. $E_{k,i,j}$ is the carcinogen concentration in units of µg/mL in undiluted smoke. This transformation expresses tobacco smoke toxicants as concentrations in the product, in common with ECs. Furthermore the denominator V* is a good proxy for nicotine (Supplementary figure 1). HnB emissions are treated the same as tobacco smoke. Toxicants in EC vapours ($C_{i,j}^{vap}$), usually reported as mass/volume concentrations, only need conversion to µg/mL, as appropriate.

Aerosol cancer potency

Environmental protection agencies assess toxicological, epidemiological and other data relevant to the long-term risks of inhaling chemicals.[15] Unit risk values for cancer have been published where data are deemed adequate.[16] The inhalation unit risk (U_j) for the j^{th} carcinogen is defined as the excess lifetime cancer risk from continuous inhalation exposure to 1 µg of the carcinogen per m³ of air and is expressed in (µg/m³)⁻¹.[15]

The cancer potency from the mixture of carcinogens in tobacco smoke may be estimated using an aggregate model that weights each inhalation unit risk by its concentration in undiluted smoke

$$P_{k,i}^{tob} = \sum_{j=1}^{m} E_{k,i,j} U_j$$
 (3).

Similarly, the cancer potency of EC vapour can be formulated as

$$P_i^{vap} = \sum_{j=1}^m C_{i,j}^{vap} U_j$$
 (4).

 P^{tob} and P^{vap} are aggregates of individual compound potencies and are measures of the carcinogenic potential of a given aerosol. The normalised ratio

$$P_i^N = P_i / \overline{P}^{tob} \tag{5}$$

is the cancer potency of smoke or vapour relative to average tobacco smoke and is the dimensionless parameter used to discuss relative potency below.

Modelling exposure and risk

Metrics of exposure that reflect human use patterns are needed to translate potencies into risk estimates for each nicotine product.[17] Studies of smoking usually report the number of cigarettes per day. Analogous reports of EC vaping topographies include estimates of vapour volume inhaled per day. Vaping is associated with greater volumes than tobacco smoking, possibly resulting in greater relative risks than implicit in the potency spectrum.[17]

The Lifetime Cancer Risk (L_i^{tob}) for the i^{th} cigarette based on daily smoke exposure (assuming HCI conditions) takes into account the mean daily cigarettes in stick units (d) and the daily breathing rate of air (b), typically 20m^3 ,

$$L_i^{tob} = \frac{\sum_{j=1}^{m} C_{HCI,i,j}^{tob} U_j}{b} d$$
 (6).

The same procedure is applied to HnB tobacco sticks.

For EC, the Lifetime Cancer Risk (L_i^{vap}) for the i^{th} product is the aerosol potency scaled by the average daily volume of vapour inhaled \bar{V}_{day}^{vap} ,

$$L_i^{vap} = P_i^{vap} \frac{\overline{V}_{day}^{vap}}{b} \tag{7}.$$

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The mean daily number of cigarettes smoked was 13 in 22 countries during 2008-14.[18] In 2016, the average daily consumption in the UK was 12 cigarettes whereas in the US in 2015 the comparable figure was 14. [19, 20] For this study, cigarettes per day were rounded to 15 reflecting the low daily smoking in these countries. The same number is applied to HnB modelling, consistent with exposure studies in Japan (average 10 sticks/day) and Poland (average 17).[21, 22]

The average daily volume of EC vapour inhaled (\bar{V}_{day}^{vap}) has been reported as 29.7 L/day based on measurements using 22 subjects over a 24 hour period but this result is associated with considerable variability (standard deviation 29.6).[23] A daily volume of 30 L is used here for risk modelling while recognising the large uncertainty involved. It has also been estimated that about four times greater volume of vapour is inhaled than tobacco smoke and this is broadly consistent with the adopted values.[24] This volume falls within the range of daily nicotine inhaler use.[25]

Data sources

Eleven compounds commonly reported in tobacco smoke are classified by the International Agency for Research on Cancer (IARC) as human carcinogens (Type 1) and a further 7 as possible human carcinogens (Type 2B). The inhalation unit risks for 14 of these carcinogens were obtained from the Office of Environmental Health Hazard Assessment database (OEHHA, State of California, USA) supplemented with a recommended value for the carcinogenic nitrosamine NNK (Table 1).[26]

Concentrations of several major carcinogens in cigarette smoke and in VNP vapour were obtained from studies of cigarettes[14, 27, 28] a prototype HnB device,[29] and EC products, ranging from early generation disposables though second-generation clearomisers and cartomisers to third generation mods and tanks.[7, 30-37] EC coil resistance and battery voltage were collated where reported. This compilation has many gaps where some carcinogens were not analysed. After removing duplicates across studies and omitting non-steady state emissions (where stated) the resulting dataset contained 93 analyses divided into three subsets.[34] The Goniewicz subset is the benchmark with 12 EC samples analysed for seven carcinogens including carbonyls, volatile

organic compounds (VOCs), nitrosamines and metals.[7] The remaining 81 analyses include concentrations for some organic compounds (formaldehyde, acetaldehyde and occasionally VOCs) and thus yield minimum estimates of potency, of which 32 involved coil heating experiments applying multiple powers to the same liquid in the same device (variable power subset). The remainder (n=48) are termed the organics-only subset.

Carcinogen emissions from an unheated nicotine inhaler represent a device accepted for medical use.[7] Ambient air is represented by a suite of analyses that establish a relatively uncontaminated reference baseline for studies of smoke and vapour exposure.[38]

CARCINOGENS			TOBACCO SMOKE n=309	HEAT-NOT-BURN EMISSION n=44	E-CIGARETTE VAPOUR n=44	NICOTINE INHALER n=1
Compound	IARC type	OEHHA unit risk U _j (μg/m³) ⁻¹	Mean concentration E _j (μg/mL)	Mean concentration E _j (μg/mL)	$\begin{array}{c} \text{Mean concentration in } 1^{\text{st}} \\ \& \ 2^{\text{nd}} \ \text{generation e-cigs} \\ C_j \ (\mu\text{g/mL}) \end{array}$	Mean concentration C _j (μg/mL)
Acetaldehyde	2B	2.7 x 10 ⁻⁶	2.55 x 10 ⁻⁰	3.33 x 10 ⁻¹	4.41 x 10 ⁻³	1.05 x 10 ⁻⁴
Formaldehyde	1	6.0×10^{-6}	1.54 x 10 ⁻¹	1.06 x 10 ⁻²	8.07 x 10 ⁻³	1.90 x 10 ⁻⁴
Acrylonitrile	2B	2.9 x 10 ⁻⁴	4.59 x 10 ⁻²	2.96 x 10 ⁻⁴	NR	NR
Benzene	1	2.9 x 10 ⁻⁵	1.57 x 10 ⁻¹	9.32×10^{-4}	NR	NR
1,3-Butadiene	1	1.7 x 10 ⁻⁴	1.83 x 10 ⁻¹	3.94 x 10 ⁻⁴	NR	NR
2-Amino-naphthalene	1	5.14 x 10 ⁻⁴	4.13 x 10 ⁻⁵	4.82×10^{-8}	NR	NR
4-Amino-biphenyl	1	6.0 x 10 ⁻³	8.68 x 10 ⁻⁶	1.80 x 10 ⁻⁸	NR	NR
Benzo[a]pyrene	1	1.1 x 10 ⁻³	3.67×10^{-5}	2.12 x 10 ⁻⁶	NR	NR
NNN	1	4.0 x 10 ⁻⁴	4.63×10^{-4}	2.57×10^{-5}	1.94 x 10 ⁻⁷	BDL
NNK	1	4.0 x 10 ⁻⁴	2.88×10^{-4}	1.64 x 10 ⁻⁵	8.39×10^{-7}	BDL
Cadmium	1	4.2 x 10 ⁻³	1.99 x 10 ⁻⁴	BDL	1.01 x 10 ⁻⁵	9.52 x 10 ⁻⁷
Lead	2B	1.2 x 10 ⁻⁵	7.52 x 10 ⁻⁵	4.09 x 10 ⁻⁶	7.06 x 10 ⁻⁶	1.90 x 10 ⁻⁶
Chromium	1	1.5 x 10 ⁻¹	BDL	BDL	NR	NR
Nickel	2B	2.6 x 10 ⁻⁴	BDL	BDL	6.98 x 10 ⁻⁶	1.90 x 10 ⁻⁶
Arsenic	1	3.3×10^{-3}	2.20 x 10 ⁻⁵	2.14 x 10 ⁻⁶	NR	NR
MEAN CANCER POTENCY RATIO (equation 5)			1.0 x 10 ⁻⁰	2.01 x 10 ⁻²	1.81 x 10 ⁻³	1.02 x 10 ⁻⁴
MEAN LIFETIME CANCER		consumption	15 cigarettes/day	15 sticks/day	30 L vapour/day	30 L vapour/day
RISK (equations 6 & 7)		risk	2.4 x 10 ⁻²	5.7 x 10 ⁻⁴	9.5 x 10 ⁻⁵	8.9 x 10 ⁻⁶
ratio to tobacco smoke			1.0	0.024	0.004	0.0004
ratio to nicotine inhaler			2697	64	10.7	1.0

Table 1. Unit risks and average concentrations for IARC type 1 and 2 carcinogens measured in tobacco smoke and other forms of nicotine delivery. Average concentrations for e-cigarettes for 1st and 2nd generation devices only (where identifiable). See text for sources of data. Concentrations in bold type represent compounds that contribute more than 5% of the modelled potency for a particular form of delivery. Abbreviations: BDL Below Detection Limit, NR Not Reported, NNN N'-nitrosonornicotine, NNK 4-(methylnitrosamino)-1-(3-pyridyl)-1-butanone.

RESULTS

Distribution of cancer potencies

Cancer potencies were computed for the datasets using equations 3 & 4 following conversion into compatible units of concentrations (equations 1-2) and matched with unit risks for cancer. The calculation sequence is summarised in Table S1 (Supplementary file). Average concentrations, unit risks and cancer potencies for various nicotine products are given in Table 1 and the potencies of individual samples are graphically presented in figure 1. All results are normalised to the average value for tobacco smoke and plotting the results on a logarithmic axis creates a relative cancer potency spectrum for the various nicotine-bearing aerosols.

Smoke collected under the ISO protocol and converted to undiluted smoke volume V^* has essentially the same potency as that collected under HCI (figure 1). Without factoring for filter ventilation the ISO samples would be widely dispersed ranging down to much lower apparent cancer potencies for the most ventilated brands. HCI and ISO tobacco smoke have the same distributions after applying the V^* conversion (figure 1) which indicates that raw undiluted tobacco smoke defines a narrow band on the potency spectrum regardless of protocol and defines the upper boundary of cancer potencies among common nicotine products in high income countries.

Ambient air, essentially a non-hazardous aerosol, defines the lower end of the cancer potency spectrum, clustering around a carcinogenic potency of slightly less than 10⁻⁵ that of tobacco smoke. The potency of the nicotine inhaler is about 10⁻⁴ that of tobacco smoke (figure 1), with metals contributing more than 70% of its carcinogenic potency.

The cancer potencies of EC are more difficult to define and are highly dispersed compared with cigarettes. The Goniewicz subset, which analyses the majority of anticipated carcinogens in EC vapour, clusters around $\sim 10^{-3}$ of the potency of tobacco smoke (figure 1). The preponderance of potencies in the organics-only and variable power subsets also fall in this range although higher

levels of carbonyls found in these two subsets extend the range as far as the potency of tobacco smoke.

The cancer potencies of the HnB prototype device that heated various tobacco blends, lie between one and two orders of magnitude less than tobacco smoke but higher than the preponderance of EC emissions (figure 1).

Lifetime cancer risk

Equations 6 and 7 estimate the lifetime cancer risks from daily exposure (Table 1). The values for ECs in this table are based on a compilation of 1st and 2nd generation devices because most 3rd generation tanks and mods were used for variable voltage experiments. The average for EC includes the Goniewicz subset supplemented by carbonyl analyses from those members of the organic-only subset for which the device generation could be identified. The potency ratio relative to tobacco smoke for the average of these ECs is 1.8 x10⁻³ while HnB had an order of magnitude higher relative potency at 2.1 x 10⁻². The excess cancer risk for a lifetime of smoking 15 cigarettes a day is 2.4 x 10⁻², 5.7 x 10⁻⁴ for inhaling 15 HnB sticks, 9.5 x10⁻⁵ for vaping 30 L from ECs at normal power, and 8.9 x10⁻⁶ for inhaling 30 L from the nicotine inhaler (Table 1). Compared with a nicotine inhaler the relative risks are 11 for EC, 64 for HnB and ~2700 for cigarettes, whereas compared with cigarettes the relative risks are 0.024 for HnB, 0.004 for EC and 0.0004 for the nicotine inhaler (Table 1).

Ranking carcinogens

The highest-ranking carcinogens in Table 1 for cigarettes are 1,3-butadiene and acrylonitrile, accounting for more than three-quarters of the cancer potency, whereas for HnB, acetaldehyde is the dominant carcinogen, which along with 1,3-butadiene accounts for almost three-quarters of the aerosol's potency. For ECs, only the Goniewicz subset includes nitrosamines and metals as well as carbonyls and VOCs. The highest ranked carcinogens in this subset are cadmium and formaldehyde. Even very small levels of cadmium can have a major effect given its extremely high unit risk (Table

1) but cadmium is not detected in all samples and in others it is present in concentrations very close to blank level.[7] The carbonyls formaldehyde and acetaldehyde account for over 95% of the contributions of organic compounds to cancer potencies in this EC subset.

The role of carbonyls in the cancer potency of EC vapour

Given the importance of the carbonyls as carcinogens in EC vapour the relationship between battery voltage and carbonyl generation is explored further in figure 2. Several studies have addressed this issue by measuring carbonyl concentrations under various experimental conditions.[31, 32, 34, 37, 39] These were omitted from the modelling of potency and risk of Table 1 because they often involve applying voltages beyond the normal range of use. This figure shows that the partial cancer potencies attributable to both formaldehyde and acetaldehyde are highly variable. The Goniewicz subset and some other samples from both of the other subsets represent the lower range of potencies attributable to carbonyls. Other data indicate considerably higher potencies due to both carbonyls, sometimes approaching and, for formaldehyde, even exceeding the partial potencies of carbonyls in conventional tobacco smoke (figure 2). Most of the high-potency data points in figure 2 were from studies investigating carbonyl formation by varying the atomiser coil heat within the same device while varying the power (watts) at the fluid-coil interface. Solid lines in figure 2 connect the various powers achieved in the same experiment with the direction of increasing power indicated by an arrow. The highest cancer potencies are commonly associated with coils subjected to the highest applied voltages in any particular experiment, however there is no consistent relationship between devices for carbonyl potency and the rate of heat energy transfer (watts) at the coil. It has been suggested that such high power operating conditions produce "dry puffs", i.e. when the supply of eliquid to the heating coil is inadequate producing an aerosol which is unpleasant to inhale although an exclusive causal relationship between dry puffs and high aldehyde concentrations has been questioned.[40, 41]

DISCUSSION

Strengths and limitations of emissions modelling

The accuracy of emissions modelling can be evaluated by comparing the lifetime risk model for tobacco smoking with epidemiological data for smoking-related mortality. An early application of this model to tobacco smoke predicted an excess risk of 1.6×10^{-2} for a daily consumption of 20 cigarettes measured with the ISO protocol.[26] This differed from the estimated excess cancer risk of 7.9×10^{-2} derived from mortality data for smokers in the USA in 1995.[14] Applying the same model to the Centers for Disease Control (CDC) tobacco smoke dataset measured with the HCI protocol weighted by current OEHHA unit risk values doubles the estimated risk to 3.3×10^{-2} averaged over 48 cigarette products (standard deviation 3.5×10^{-3}) for the same daily consumption rate. This closer agreement is largely due to the use of the HCI machine smoking protocol for emissions suggesting that these independent estimates reflect the same fundamental causes of mortality.

Using emissions can under- and over-estimate risk Aggregation of risks for individual carcinogens assumes that each inhaled carcinogen is wholly absorbed and no account is taken of differential mechanisms and rates of clearance. Unit risks are upper bound toxicological estimates that may overestimate true potency.[42] If large volumes of carbonyls are generated by EC use under circumstances not normally tolerated by humans (e.g. "dry puffs") then the apparent high risk levels may not be reached.[40]

A simple aggregation does not account for enhancement by synergystic interactions between components (but there may also be antagonistic interactions). A few synergystic effects are well known, often between tobacco smoke and particles such as asbestos and silica but these are largely restricted to industrial exposure (although silica particles have been recognised in EC vapour).[43, 44] Unit risk values are not available for all known carcinogens and are thus not involved in the aggregation leading to underestimation. Furthermore, only carcinogens formally classified by IARC are included. Although cigarette smoke has been thoroughly investigated for toxicants over several

decades EC have not received the same level of scrutiny and some carcinogens are possibly yet to be identified, although these "missing carcinogens" would need to be present in high concentrations and/or have high unit risks to make a noticeable difference to EC potency and risk. Other sources of uncertainties are differences in unit risks between environmental protection authorities, and the assumption that risk is linearly proportional to dose with no threshold.

Another limitation is the absence of carcinogen analyses in most published studies of EC vapours. Only the Goniewicz subset presents data for most of the relevant carcinogens and these indicate very important roles for metals (especially cadmium) and formaldehyde. The importance of cadmium as a carcinogen in vapour needs further investigation, as it has not been detected at hazardous levels in other studies of vapour.[44-46] Also missing from most studies are analyses of the nitrosamine carcinogens, notably (methylnitrosamino)-1-(3-pyridyl)-1-butanone (NNK) and N'-nitrosonornicotine (NNN), although potencies calculated from the Goniewicz subset suggest these are minor contributors to the risk (Table 1). Indeed missing data for carcinogens at low concentrations or with low unit risks may have little or no significant effect on modelled potency and risk. Note that particle size effects have not been taken into account in this or previous chemical studies of this type but are potentially significant.[47, 48]

VNP and cigarette emissions have been compared in this study using carcinogen concentrations whereas normalisation to unit nicotine levels is sometimes preferred as an estimate of human exposure.[49] Nicotine has an unprotonated (free-base) and two protonated forms, unprotonated nicotine being of greatest concern in terms of blood transfer. The fraction of total nicotine in free-base form varies widely in both tobacco smoke and EC vapour largely as a function of pH.[50, 51] Whether free-base nicotine could be a useful proxy for exposure in tobacco-based aerosols requires further research as does the relationship between the concentrations of the various nicotine forms and exposure to toxicants in EC emissions.

Biomarkers for 1,3-butadiene and acrylonitrile in urine samples from EC-only users indicate levels

similar to nicotine replacement therapy users, in contrast to the high levels found in smokers.[10] Biomarkers for the same compounds declined rapidly in smokers switching to ECs at rates similar to abstinence from smoking over five days.[52] Neither compound is normally detectable in EC vapour yet both have been shown above to account collectively for more than three-quarters of the cancer potency of tobacco smoke.[53] This close correlation between reduced emissions in EC vapours and their associated biomarkers for two of the most important carcinogens in tobacco smoke supports the use of emissions in estimating potency and risk.

Carbonyl generation

The preponderance of EC analyses conducted at recommended voltages indicates relatively low risks of exposure related to carbonyls (figure 2, table 1), notwithstanding, additional carbonyl generation at higher powers has a major influence on cancer potency (figure 2). While factors such as the propylene glycol/glycerol balance and other e-liquid ingredients as well as puff number may influence aldehyde formation, overheating appears to dominate.[34, 54-56] Indeed, the connected arrows on figure 2 indicate that up to half the cancer potency range for ECs can be generated from the same liquid in a single device by varying only the power. This implies that the way a device is used might be more important than the device itself. The unpleasant sensory experience of a "dry puff" may mean that these laboratory-derived data points are replicated less frequently in normal human exposure than implied by the machine smoking data presented in figure 2.[40]

Heat-not-Burn devices

Data from experiments on one HnB device indicates cancer potencies and risks lower than tobacco smoke by more than one order of magnitude greater than the preponderance of ECs operating under normal conditions (Table 1). This finding is provisional as it is based on a single prototype design from a single manufacturer.

Caveats

Unit risk values (Table 1) indicate a potentially major role for metals even at low concentrations. ECs are typically metallic devices, not least the atomiser which is normally a metallic coil usually made of nichrome (NiCr alloy) or kanthal (FeCrAl alloy). Other metallic components can include electrical conductors, crimps, connectors, solder, cartridges and structural features of the device, many of which show some evidence of corrosion over time in pre-filled devices. Several papers document elevated levels of metals in liquids and occasionally the aerosol but no consistency is yet evident in emissions.[7, 44, 45, 57] The lack of metals data for samples other than the Goniewicz subset mean that some cancer potency values for ECs in this study may be underestimated. Even less is known about the speciation of metals in EC vapour despite the importance of different valence states and molecular speciation in tobacco smoke toxicity.[58, 59] More detailed study of metals, especially Ni, Cr and Pb and their speciation in a wider of range of products may lead to new perspectives on the cancer potential of metals in EC vapours.

The presence of fine and nanosize particulates in EC vapour was recently demonstrated.[48] Given the greater surface area and potential reactivity of nanoparticles, if their presence in VNP vapours is confirmed as substantial then safety estimates for these products may require re-evaluation, especially in relation to cancer, respiratory and cardiovascular diseases including stroke.[60]

The cancer potency of second hand vapour was not specifically addressed in this study. The few published experiments of secondhand vaping suggest low cancer potencies although none appeared to involve products that emitted relatively high levels of carbonyls in the mainstream vapour.[61, 62]

Implications for policy and practice

The data imply a considerable range of cancer risks. Many EC emissions have cancer potencies within an order of magnitude of a nicotine inhaler, a product generally regarded as safe. Notwithstanding, some EC emissions tended towards much higher cancer potencies and risks, a few possibly approaching those of tobacco smoke. Indeed, the cancer potency of formaldehyde, the

most important EC carcinogen, can exceed that of tobacco smoke especially in the highest power settings of some multi-power experiments but the full range of relative potency spans nearly 4 orders of magnitude with the vast majority of potencies being much lower than combustible cigarettes. It is likely that 3rd-4th generation EC devices with adjustable coil power, are implicated in these higher risks. With regard to involuntary exposure, better understanding of potential effects of second-hand EC exposure is needed to determine if their use in indoor public spaces should be banned, as is currently the case in 25 countries.[63]

The conclusions of this study refer only to the chemical risks of cancer and do not account for any other carcinogenic effects such as those attributable to small particle sizes. There is some evidence that the large differentials found here between the cancer potencies of most ECs and those of tobacco smoke may be less for other medical conditions involving the cardiovascular and respiratory systems.[37] Future research should address these gaps.

The cancer potency calculation for VNPs was applied to a range of devices covering different generations purchased in different countries but no claim is made that these samples are representative of all products in the current marketplace. As the VNP market continues to expand and diversify a better understanding of the origins and aerosol transfer of carbonyls and metals is required to achieve EC emissions with consistently low carcinogenic potency. Such understanding may prompt calls for constraints on device design and e-liquid formulations but it is likely that significant reduction in cancer risk could also be achieved by effective advice to EC users on generic aspects of devices, coil selection and refill liquids as well as vaping behaviour and product manipulation.

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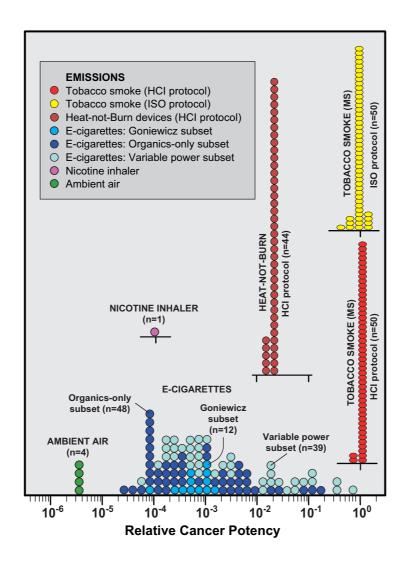


Figure 1. Relative cancer potency modelled for common forms of nicotine delivery. Potency is the aggregate estimated using equations 3 & 4 and normalised to average tobacco smoke with cancer risk set to 1.0 (equation 5). Each circle or ellipse on the graph represents the emission from a single sample or experiment in the peer reviewed literature. In all emissions from 14 peer-reviewed studies are summarised in the diagram (see text for data sources). MS=Mainstream smoke.

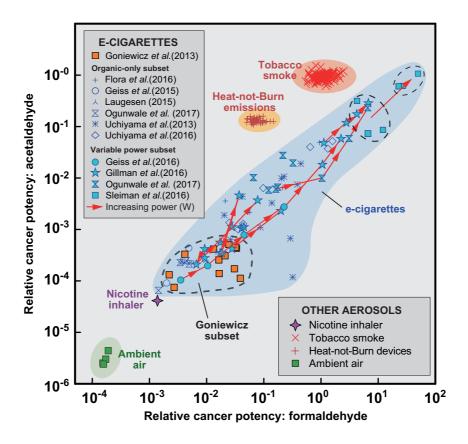


Figure 2. Relative cancer potencies of formaldehyde and acetaldehyde in vapour from e-cigarettes compared with tobacco smoke, heat-not-burn devices, a nicotine inhaler and ambient air. Values are normalised to the mean potency of the same carcinogens in tobacco smoke. Data sources for e-cigarettes are identified with different symbols (see key for sources). The variable power subset includes all devices used in experiments to create vapour at more than one atomiser power (watts). Solid red lines connect the potencies of emissions for the same device run at different powers with arrows indicating the direction of increasing power. The lines show that the cancer potencies from the same device can vary by more than two orders of magnitude for both formaldehyde and acetaldehyde.