RESEARCH ARTICLE

A time geographic approach for delineating areas of sustained wildlife use

Trisalyn A. Nelson^{a*}, Jed A. Long^a, Karen Laberee^a and Benjamin P. Stewart^a

^aSpatial Pattern Analysis and Research Lab, Department of Geography, University of

Victoria, Victoria, BC, Canada

(Received ***; accepted***)

*Corresponding author. Email: trisalyn@uvic.ca

Pre-print of published version.

Reference:

Nelson, TA, Long, JA, Laberee, K, Stewart, BP. 2015. A time geographic approach for delineating areas of sustained wildlife use. Annals of GIS. 21(1): 81-90.

DOI:

10.1080/19475683.2014.992366

Disclaimer:

The PDF document is a copy of the final version of this manuscript that was subsequently accepted by the journal for publication. The paper has been through peer review, but it has not been subject to any additional copy-editing or journal specific formatting (so will look different from the final version of record, which may be accessed following the DOI above depending on your access situation).

Abstract

2 Geographic information systems (GIS) are widely used for mapping wildlife movement patterns, and observed wildlife locations are surrogates for inferring on wildlife 3 4 movement and habitat selection. We present a new approach to mapping areas where 5 wildlife exhibit sustained use, which we term slow movement areas (SMAs). Nested 6 within the habitat selection concepts of home range and core areas, SMAs are an 7 additional approach to identifying areas important for wildlife. Our method for 8 delineating SMAs is demonstrated on a grizzly bear (*Ursus arctos*) case study examining 9 road density. Our results showed that subadult females had significantly higher road 10 densities within SMAs than in their PPA home ranges. The lowest road density was 11 found in the SMAs of adult male grizzly bears. Given increased mortality risks associated 12 with roads, female encampment near roads may have negative conservation implications. 13 The methods presented in this manuscript compliment recent developments to identify 14 movement suspension and intensively exploited areas defined from wildlife telemetry 15 data. SMA delineation is sensitive to missing data and best applied to telemetry data 16 collected with a consistent resolution. **Keywords:** time geography; stopover ecology; GPS telemetry; potential path area; 17 18 grizzly bear

1. Introduction

Due to improved GPS technology there has been an increase in availability of telemetry
data that has led to growth in movement analysis methods development (e.g., Thériault et
al. 1999, Dodge et al. 2008, Long and Nelson 2013). At the frontier of movement
research are wildlife studies that use movement as a surrogate for understanding
behaviour. Taking a classic spatial statistics perspective, the spatial pattern of observed
wildlife locations is an expression of spatial processes that are difficult to measure
directly (Nelson and Boots 2008). In this case, the spatial processes are biological and
originate from dynamic wildlife behaviour. Given that we cannot observe behaviour
continuously in space and time, patterns of movement are a surrogate measure for
behavioural states (Morales et al. 2004). For instance, Hunter (2007) determined that
foraging behaviours occurred when grizzly bears (Ursus arctos) were moving at a
velocity of less than 52m/minute. Food searching was associated with movement
velocities of 52 m/minute to 223 m/minute and active walking occurred at velocities of
greater than 223 m/minute.
Related to wildlife movement research is the use of telemetry data for
understanding spatial and temporal patterns of habitat selection (e.g., Berland et al. 2008)
Telemetry data represent discrete locations of an individual animal in space and time and
have been used extensively to study habitat selection by wildlife (Smulders et al. 2010).
Most habitat selection research employs the concept of home range or core area
(Smulders et al. 2012). A home range is typically defined as the area to which an animal
confines its normal movements (Burt 1943) and the core area is an intensively utilised
subset of the home range (Samuel and Green 1988). Recently, other concepts such as

43 intensively selected areas are also being employed (Benhamou and Roitte-Lambert 2012). 44 There are many ways to define a home range and core area, but arguably the most 45 common is by applying kernel density estimation to telemetry data to generate an 46 utilisation distribution. The 95% contour of the utilisation distribution is associated with 47 the home range while the 50% contour is associated with the core (Worton 1987). Spatial 48 units, like home range and core area, represented as discrete polygons, are integrated with 49 spatially continuous data on the physical environment to characterise habitat conditions. 50 The end goal is often to characterise the environmental elements (e.g., land cover types) 51 that can be best managed for the purpose of wildlife conservation (Bourbonnais et al. 52 2013). 53 Even with the availability of detailed wildlife location data and more continuous 54 and timely landscape data from remotely sensed imagery, it has proven difficult to 55 quantify links between movement, behaviour, and habitat. As typical with spatial pattern 56 analysis, assumptions are required to make a linkage between spatial patterns of wildlife 57 occurrence and processes of wildlife behaviour (Getis and Boots 1978). These 58 assumptions pose difficulty due to subjective thresholds that are applied to patterns to 59 categorise behaviour (Hunter 2007). Concepts of home range and core area have persisted 60 in the literature, in part, because they are conceptually tidy and do not require inference 61 on behaviour. Using home ranges and related concepts, habitats are associated with 62 locations where wildlife are observed and the more often they are in observed in a 63 location the greater the resource utilisation. Though home range and core areas are 64 important for identifying habitats selected by wildlife, they cannot be employed to

identify wildlife "use" areas, as the activity of wildlife at any given location is not known.

In this paper we revisit the notion of spatial-temporal patterns of resource
utilisation, selection, and use by defining slow movement areas (SMAs). SMAs are
defined at the finest spatial and temporal scale afforded by given telemetry data sets. We
argue that when animals are moving the slowest there is a high likelihood that the
selected or available habitat is being actively utilised for a specific biological function.
We do not classify use behaviour in SMAs, however likely behaviours are resting and
feeding or stopovers, depending on the scale of data.
We propose a new method for delineating SMAs by modifying an existing
technique for quantifying animal home ranges: the potential path area (PPA) home range
(Long and Nelson 2012). This method builds upon an existing analytical framework,
termed time geography (Hägerstrand 1970), useful for quantifying and examining the role
of spatial-temporal constraints on movement (Baer and Butler 1999). The PPA method
takes a pragmatic approach to movement analysis, focused on defining areas that are
spatially and temporally accessible. Given the importance of relating movement to habitat
conditions, the PPA polygons provide a simple approach for characterising habitats
associated with locations likely utilised. The benefit of using the existing PPA to define
new SMAs is demonstrated through a case study on grizzly bears in Alberta, Canada. We
show how within PPA home ranges, which represent habitat selection, and SMAs, which
represent likely habitat use, the density of roads varies and trends for males and females
are also opposite. Given that bears in Alberta are most likely to die near roads (Benn and
Herrero 2002), intensive habitat use near roads may negatively impact survival
(Bourbonnais et al. 2013).

2. Slow Movement Areas (SMAs)

While there are several methods for computing a home range, we based our approach for computing SMAs on Long and Nelson (2012) and calculated a grizzly bear home range using the PPA home range method. The PPA home range delineates the area accessible to the animal given its sequence of telemetry fixes and a movement parameter (v_{max} – defined as an animal's maximum travelling speed). The spatial area of accessibility between two fixes can be defined based on the single parameter – v_{max} , and the time difference between the two fixes, and is easily computed as a perfect ellipse shape. (Figure 1 – upper panel). The PPA ellipse encompasses the entire area the animal could have traversed based on its maximum travel speed and the location and time duration between consecutive fixes. By combining the n-1 ellipses, from a dataset of n telemetry fixes, the PPA home range is delineated (Long and Nelson 2012).

In order to compute SMAs we first define a statistic m_i (i = 1...n-1) representing the number of consecutive telemetry fixes that fall within each PPA ellipse (Figure 1 – lower panel). By definition, each ellipse will include, at a minimum, two fixes ($m_i \ge 2$; i.e., the current fix, and the next telemetry fix). The m_i with the highest scores can then be used to represent SMAs on a map, taking the highest score(s), or based on some threshold. Mapping of SMAs involves taking the union of the m_j -1 PPA ellipses of the m_j fixes beginning with index j, where an SMA can be defined as:

$$SMA = \bigcup PPA_{j..(j+mj-1)})$$

for each high scoring m_i . Spatially, the SMAs are sub-regions of the individual's home range and represent the local accessibility space while encamped or slow moving. Once delineated, SMA polygons can be treated much like home range polygons for analysing underlying environmental characteristics. Similarly, because of how they are defined, the

SMAs also represent a temporal sub-interval of the telemetry dataset, and this temporal information can be used to further assess the timing of encamped and slow movement behaviour.

The calculation of the PPA home range and SMAs will be impacted by the selection of the v_{max} parameter. Parameterising v_{max} is subject to similar issues as have been discussed for the bandwidth selection when using kernel density estimation to define home ranges (Seamann *et al.* 1999, Gitzen *et al.* 2006, Nelson and Boots 2008). Higher values of v_{max} will lead to the delineation of larger PPA home range and SMAs. As v_{max} increases the animal is represented as being able to move more quickly and therefore has more accessible habitat. Like kernel density bandwidth selection, selecting v_{max} will always be prone to some subjectivity (see Nelson and Boots 2008 for discussion of bandwidth selection). We suggest that analysts use multiple confirmatory sources when determining the most appropriate v_{max} parameter. Biological information on maximum or typical speeds of travel can be compared to estimates generated from observed data to build confidence in the v_{max} value selected.

The spatial-temporal extent and resolution of telemetry data will also impact the interpretation, and indeed appropriateness, of home range and SMAs defined using PPA approaches (Figure 2). Ideally, PPA approaches are applied when the spatial-temporal resolutions and extents of telemetry fixes are similar throughout a dataset (see Wiens 1989 for discussion of scale). For instance, if wildlife data are collected every 20 minutes in a 10 by 10 km area the data are relatively fine and SMAs are likely representation of sleeping or feeding. In contrast, landscape scale trends, such as migratory stop-over locations, could be identified when the SMA is defined for data collected once a day over

a broad area. Interpreting a SMA will be problematic if the resolution of the data is coarse and the study area fine, as the areas delineated as SMA will be overgeneralised.

Similarly, if the study area is very large and the resolution of telemetry data fine, the SMAs defined will likely be too small. Partitioning the telemetry data into smaller subsets prior to analysis may lead to more meaningful results.

Users should also be cautioned against defining SMAs in datasets that have variable spatial-temporal resolutions and missing fixes. If some fixes are taken at both one and four hour intervals, the longer intervals will have larger PPA ellipses and be biased towards higher counts of consecutive points within the ellipse. When data are sampled at varying resolutions the data can be partitioned by resolution, for separate analysis and SMA delineation, or all the data downgraded to the coarsest resolution.

Missing fixes are also problematic. If dropped fixes are not accounted for, PPA ellipses could be artificially large and/or counts of consecutive points within the ellipse low. If many fixes are dropped we recommend excluding that portion of the telemetry data from SMA calculations to preserve analysis integrity. However, if only a few fixes are missing it may be possible to clean the data by interpolating fixes. Given the sensitivity of many methods to missing data (Frair *et al.* 2010) corrections, such as linear interpolation based on curvilinear interpolation, have been demonstrated to improve data quality (Tremblay *et al.* 2006). Another technical issue will arise when fixes are dropped if an individual is denning or resting in an area that has poor signal coverage. No method can pick up habitat selection or use in locations that for reasons of terrain or vegetation do not record telemetry fixes and many movement metrics are sensitive to missing data (Laube and Purves 2011). However, since SMAs are intended to pick up slow movement,

which may be associated with resting, missed resting locations is an important omission to consider. There is no systematic way to identify omitted resting area due to missing data. Rather, when fix frequency becomes low or is missed for an extended period we recommend manual assessment.

3. Case study

Our methodology to define SMAs is demonstrated with a case study of grizzly bears in Alberta, Canada. In the Kakwa region of west-central Alberta, grizzly bears share their habitat with many anthropogenic disturbances that are affecting the bears' traditional use and selection of habitat. To illustrate the utility of our methodology, we examined the density of roads within the SMAs compared to the PPA home range. Road density has been found to correlate with mortality risk and reduced survival in grizzly bears (Benn and Herrero 2002, Nielsen *et al.* 2004), yet areas with high road densities are often selected as habitat (Roever *et al.* 2008, Graham *et al.* 2010, Stewart *et al.* 2013). It is not fully understood why bears appear to select habitat with roads, but it has been speculated that roadside areas offer bear food (Roever *et al.* 2008, Graham *et al.* 2010, Stewart *et al.* 2013).

Differences in habitat selection by male and female grizzly bears are becoming increasingly documented as the body of grizzly bear research grows. The much larger males are known to have larger home ranges (Proctor *et al.* 2004, Roever *et al.* 2008, Graham *et al.* 2010) and greater daily movement rates when compared to females (Boulanger *et al.* 2013). In contrast, females have been found to select habitat containing roads more than males (Roever *et al.* 2008, Graham *et al.* 2010, Stewart *et al.* 2013). Female selection of roads is of concern, especially for a threatened population, given that

female survival is paramount for population viability (Eberhardt *et al.* 1994, Stewart *et al.* 2013). Understanding the behaviour of female grizzly bears associated with roads will provide important conservation information for those tasked with land-use decision making.

3.1 Study area and data

The study area for this research is an 8308 km² landscape in the Kakwa region of west-central Alberta, Canada. The elevation ranges from 549 m to 2446 m, and the area comprises a diverse and multi-use landscape. Resource extraction industries have been active in the area for a number of decades (White *et al.* 2011), with most disturbances in the area arising from the forest industry and oil and gas exploration (Schneider 2002).

A dataset of GPS locations collected over 2005–2010 from 25 grizzly bears in the study area were provided by the Foothills Research Institute Grizzly Bear Program (Hinton, AB). The FRI researchers followed the accepted protocols of the Canadian council of animal care for the safe handling of bears (animal use protocol number 20010016) (Stenhouse and Munro 2000). Bears were fitted with Televilt/Followit brand GPS collars (Lindesburg, Sweden). We obtained road data for the study area from Alberta Environment and Sustainable Resource Development and updated it through heads up digitising of medium and high resolution satellite imagery (SPOT and air photos).

3.2 Analysis

By definition, SMAs are delineated using both the spatial and the temporal structure of telemetry data. Thus, we began our analysis by correcting for missing fixes that are inevitable with GPS-based telemetry systems (Rempel *et al.* 1995). An interpolation

algorithm was used to accommodate for missing fixes in order to generate trajectories with consistent sampling intervals (regular trajectories—Calenge *et al.* 2009). When the GPS-signal was disrupted for an extended period of time (i.e., > 4 fixes) we analysed the bear trajectory separately on either side of the disruption.

For each individual bear we calculated the PPA home range and SMA (Figure 3). We took a simple approach to SMA analysis here using only the longest encamped period (i.e., the m_i with the highest score) to generate the SMA. Road density within individual bear home ranges and SMAs were then calculated and summarised by age and sex. We excluded the SMA from the home range when calculating road density to compare between the home range and SMA. Grizzly bears less than five years old are considered subadults and their selection of habitat has been shown to be different from adult bears (Mueller *et al.* 2004). We partitioned bears by age (subadult or adult) and by sex. For each age-sex class (adult females, adult males, subadult females, subadult males) we assessed the statistical differences in the road density within home ranges and SMAs by comparing frequency distributions using a Mann-Whitney U statistical test.

3.3 Results

For adult females, the average PPA home range was 466.83 km², whereas the average SMA size was 117.24 km². The average PPA home range for sudadult females was 540.00 km² and the average SMA was calculated to be 153.91 km². When all female data were combined, the average PPA home range was 479.44 km² and the average SMA was 131.72 km².

Adult male grizzly bears were found to have an average PPA home range that was 674.58 km² and an average SMA that was 149.21 km². The average PPA home range for

subadult males was 560.70 km² whereas their average SMA was found to be 136.81 km². When all males were considered together, the average PPA home range was 651.49 km² and the SMA was calculated to be 144.53 km².

Average road density was calculated for all groups in both the PPA home range and the SMAs (Table 1, Figure 4 and Figure 5). Road density in the SMAs for adult females was very similar to the road density in the PPA home range (0.60 km/km 2 and 0.59 km/km 2 , respectively) (Table 1 and Figure 5). However, for subadult females, the road density was significantly higher (p = 0.0209) in the SMA compared to the PPA home range (0.66 km/km 2 and 0.50 km/km 2 , respectively)(Table 1 and Figure 5). In general, male grizzly bears were found to have lower road densities in their

SMAs compared to their PPA home ranges (Table 1 and Figure 4). The lowest road density was found in the SMAs of adult males (0.43 km/km²). Males also generally had lower road densities in both their SMAs and PPA home ranges (0.46 km/km² and 0.52 km/km², respectively) compared to their female counterparts (0.63 km/km² and 0.57 km/km², respectively).

4. Discussion

Grizzly bears often rest adjacent to sites recently used for feeding (Phillips 1987) and it is reasonable to assume that the low mobility activities in the SMAs consisted primarily of feeding/foraging and resting. Previous research has demonstrated the selection of roads by grizzly bears (Chruszcz *et al.* 2003, Graham *et al.* 2010, Stewart *et al.* 2013), yet were not able to provide movement details. Roads have been associated with increased mortality in grizzly bears (Benn and Herrero 2002) and it is important to fully recognise their attraction to bears when making land-use decisions that support conservation (see

Stewart *et al.* 2013 for a more in-depth discussion) and aid in population recovery efforts. It is concerning that the subadult females in our study had a significantly greater concentration of roads in their SMAs compared to the remainder of their home ranges. The survival of vulnerable subadult females into the adult breeding stage is essential for population viability (Mueller *et al.* 2004). While previous studies have observed the selection of roads by subadult females (Mueller *et al.* 2004), the results of our case study provide insights into the behaviour associated with roads.

It is interesting that the male grizzly bears had fewer roads in their SMAs compared to the remainder of their home range and also to female bears. A previous study using the same database had found male grizzly bears to select natural edge habitats over anthropogenic edges (Stewart *et al.* 2013). Our case study has enabled us to determine that slow movement behaviours in males are associated with areas with fewer roads. A road density of 0.6 km/km² has been previously postulated as the limit for naturally functioning landscapes containing sustained populations of large predators including grizzly bears (Forman and Alexander 1998). Our study suggests that areas with lower road densities appear to be most desirable for adult males' encampment and it is possible that the female bears are being competitively excluded from these areas by more dominant conspecifics (Mattson *et al.* 1987; Edwards *et al.* 2011).

Methods for analysing spatial-temporal data have been touted as an opportunity area for spatial science development (Nelson 2012). Movement data are inherently spatial and temporal and there are many examples of recent developments in methods for quantifying movement in people (Jankowski *et al.* 2010), wildlife (Langrock 2012), and traffic (Andrienko and Andrienko 2013). SMA delineation compliments recent progress

in movement science, such as the development of methods to identify suspension in human movement (Orellana and Wachowicz 2011), stopover ecology (Sawyer and Kauffman 2011), and areas intensively exploited by wildlife (Benhamou and Riotte-Lambert 2012).

Wildlife researchers require a range of methods to characterise different types of movement patterns. The most simple movement pattern measure is velocity obtained by dividing the spatial distance by the time difference of two consecutive fixes (Brillinger *et al.* 2004, Chapman *et al.* 2007). Velocity (speed, time lag, or step length) (Brillinger *et al.* 2004, Calenge *et al.* 2009) and other metrics (turning angle and bearing) (Turchin 1986, Calenge *et al.* 2009) are related to behaviour by defining arbitrary thresholds. A one-to-one relationship between spatial patterns of movement and behaviour is difficult to define making it problematic to relate behaviour and habitat. As well, it is often desirable to associate movement behaviour with an area, as is often the case in habitat selection studies, point based representations are limited. Another potential limitation of basic velocity measures is that they are computed based on only two consecutive fixes, ignoring potentially useful information from larger consecutive intervals within the telemetry data. Turning angle is typically computed on three points but has similar limitations.

The theory of SMAs links conceptually with existing notions of home range and core area delineation (Worton 1987). Without requiring a classification of wildlife behaviour, SMAs allow us to define areas that have a high probability of resource use. The nature of that resources use will vary depending on species and scales of data. Like home range and core area delineation the strength of SMAs lies in the assumption that

spatial patterns are expressions of spatial processes (Getis and Boots 1978). A single spatial pattern can be related to many different processes making behaviour difficult to infer. Therefore, methods that can identify utilisation, without requiring one-to-one relationships with behaviour, are important for wildlife research and support assessment of utilised habitats and wildlife conservation. However, a unique component of both the PPA and SMA methods is the utilisation of the temporal component of the data. The selection of SMAs is consecutive in time. With increasingly available high resolution telemetry data, the SMA approach to identifying habitats associated with sub-regions of the home range associated with encamped or slow movement behaviours. In the case of the grizzly bear, identifying SMAs may indicate critical foraging regions that are important in conservation management efforts since grizzly bears require almost continuous feeding to meet their nutritional needs (Rode *et al.* 2001).

As SMAs are an extension of the PPA home range approach there are also strong ties to other recently developed path-based measures of animal home range, namely the Brownian bridge home range (Bullard 1999, Horne *et al.* 2007), and time geographic kernel density estimation (Downs 2010, Downs *et al.* 2011). The PPA home range represents a direct measure of spatial range (as a spatial polygon) while both the Brownian bridge and time geographic kernel density estimation methods first estimate a utilisation distribution, followed by extracting a home range polygon. The definition of SMAs within a PPA home range is another mechanism for understanding utilisation, though it is a measure of encampment rather than percentage time spent at a location. A benefit of the SMA delineation is that areas are defined using maximum speed as the only subjective parameter. Our approach specifically does not identify frequently revisited

areas and alternative approaches (e.g., Benhamou and Riotte-Lambert 2012) are more appropriately designed specifically for discovering these revisitation areas, for example associated with important movement corridors.

To calculate SMAs, every telemetry location is assigned both an ellipse and a value (m_i) identifying the number of consecutive points that fall within that ellipse. As such, it is possible to map how long an animal was in an area for all PPA ellipses or telemetry locations. Different lengths of utilisation could be linked with different behaviours (i.e., foraging, resting, and travelling); however, as with velocities, linkages to behaviour require that subjective thresholds be defined. In this analysis we defined the SMA using only the largest – $\max(m_i)$ – value, identifying a SMA, however in many applications it will be advantageous incorporate, for example, the 10 largest values of m_i . This may be especially important with larger telemetry datasets covering long temporal durations, where multiple SMAs could identify recurring behaviour associated with sustained use and low mobility rates.

Future research could develop techniques for more objective definitions of movement pattern thresholds. For instance, using theory from spatial statistics it may be possible to begin teasing apart when various movement patterns are most likely realisations of different processes (Getis and Boots 1978, Smulders *et al.* 2010). Similar to variability in the nature of home ranges, how the SMA is utilised is related to the spatial and temporal scales of the telemetry data. As in the grizzly bear example, when data are hourly or finer, the behaviours most likely associated with SMAs are feeding or resting. If telemetry data are collected at coarse temporal resolutions and extents the SMAs will reflect broader scale processes such as migratory stopovers.

5. Conclusions

New methods for characterising wildlife movement patterns will give researchers greater flexibility in the types of hypotheses investigated. We present a new approach to delineating areas where an animal exhibits sustained use. Similar to home ranges and core areas, SMAs are areas where spatial patterns indicate habitat selection and do not require explicit categorisation of behaviour. However, SMAs are related to encampment and will represent a range of short-term behaviours such as foraging or resting when telemetry data are collected frequently, and migratory stopover locations for data sets with a long temporal extent. Regardless of scale, the areas defined by the SMA have a high likelihood of wildlife resource use. SMA delineation methods require consistent spatial-temporal resolutions and minimal missing data. Future research should investigate how a time geographic framework, such as the PPA ellipses presented here, can be used to map a range of habitat utilisation behaviours based on length of time spent in each area.

Acknowledgements

This research was funded by NSERC Discovery Grant number 327664 to Dr. Trisalyn Nelson. The authors thank researchers with the Foothills Research Institute (FRI) Grizzly Bear Program for the use of the grizzly bear dataset and the many funding partners of this research program. Gordon Stenhouse (FRI) provided helpful insights into the interpretation of the results. Tracy McKay (FRI) is thanked for her feedback during the development of this methodology. Gillian Harvey is thanked for her assistance with the analysis.

References

365	Andrienko, N., and Andrienko, G., 2013. Visual analytics of movement: An overview of						
366	methods, tools and procedures. Information Visualization, 12(1), 3-24.						
367	Berland, A., Nelson, T., Stenhouse, G., Graham, K., & Cranston, J. (2008). The impact of						
368	landscape disturbance on grizzly bear habitat use in the Foothills Model Forest,						
369	Alberta, Canada. Forest ecology and management, 256(11), 1875–1883.						
370	Benhamou, S. and Riotte-Lambert, L., 2012. Beyond the Utilisation Distribution:						
371	identifying home range areas that are intensively exploited or repeatedly visited.						
372	Ecological Modelling, 227, 112–116.						
373	Benn, B. and Herrero, S., 2002. Grizzly bear mortality and human access in Banff and						
374	Yoho National Parks, 1971–98. Ursus, 13, 213–221.						
375	Boulanger, J., Cattet, M., Nielsen, S.E., Stenhouse, G., and Cranston, J., 2013. Use of						
376	multi-state models to explore relationships between changes in body condition,						
377	habitat and survival of grizzly bears Ursus arctos horribilis. Wildlife Biology,						
378	19(3), 274–288.						
379	Bourbonnais, M.L., Nelson, T.A., Cattet, M.R., Darimont, C.T., and Stenhouse, G.B.,						
380	2013. Spatial analysis of factors influencing long-term stress in the grizzly bear						
381	(Ursus arctos) population of Alberta, Canada. PloS one, 8(12), e83768.						
382	Brillinger, D.R., Preisler, H.K., Ager, A.A. and Kie, J.G., 2004. An exploratory data						
383	analysis (EDA) of the paths of moving animals, Journal of Statistical Planning						
384	and Inference, 122(1-2), 43–63.						
385	Bullard, F., 1999. Estimating the home range of an animal: A Brownian bridge approach						
386	Master of Science, University of North Carolina, Chapel Hill.						

387	Burt, W. H., 1943. Territoriality and nome range concepts as applied to mammals.
388	Journal of Mammalogy, 24, 346-352.
389	Calenge, C., Dray, S. and Royer-Carenzi, M., 2009. The concept of animals' trajectories
390	from a data analysis perspective. <i>Ecological Informatics</i> , 4, 34–41.
391	Chapman, D.S., Dytham, C. and Oxford, G.S., 2007. Landscape and fine-scale
392	movements of a leaf beetle: the importance of boundary behaviour. Oecologia,
393	154(1), 55–64.
394	Chruszcz, B. Clevenger, A.P., Gunson, K.E. and Gibeau, M.L., 2003. Relationships
395	among grizzly bears, highways, and habitat in the Banff-Bow Valley, Alberta,
396	Canada. Canadian Journal of Zoology, 81, 1378–1391.
397	Dodge, S., Weibel, R. and Lautenschütz, A.K., 2008. Towards a taxonomy of movement
398	patterns. Information Visualization, 7(3-4), 240–252.
399	Downs, J.A., 2010. Time-geographic density estimation for moving point objects.
400	Lecture Notes in Computer Science, 6292, 16-26.
401	Downs, J.A., Horner, M.W. and Tucker, A.D., 2011. Time-geographic density estimation
402	for home range analysis. Annals of GIS, 17, 163–171.
403	Eberhardt L.L., Blanchard, B.M. and Knight, R.R., 1994. Population trend of the
404	Yellowstone grizzly bear as estimated from reproductive and survival rates.
405	Canadian Journal of Zoology, 72, 360–363
406	Edwards, M.A., Derocher, A.E., Hobson, K.A., Branigan, M., and Nagy, J.A., 2011. Fast
407	carnivores and slow herbivores: differential foraging strategies among grizzly
408	bears in the Canadian Arctic. Oecologia, 165, 877–889.

109	Forman, R.T.T., and Alexander, L.E., 1998. Roads and their major ecological effects.
410	Annual Review of Ecology and Systematics, 29, 207–231.
411	Frair, J.L., Fieberg, J., Hebblewhite, M., Cagnacci, F., DeCesare, N.J., and Pedrotti, L.,
412	2010. Resolving issues of imprecise and habitat-biased locations in ecological
413	analyses using GPS telemetry data. Philosophical Transactions of the Royal
414	Society B: Biological Sciences 365.1550 (2010): 2187–2200.
415	Getis, A. and Boots, B., 1978. Models of Spatial Processes. Cambridge: Cambridge
416	University Press.
417	Gitzen, R.A., Millspaugh, J.J. and Kernohan, B.J., 2006. Bandwidth selection for fixed-
418	kernel analysis of animal utilization distributions. Journal of Wildlife
419	Management, 70, 1334–1344.
420	Graham, K., Boulanger, J., Duval, J. and Stenhouse, G., 2010. Spatial and temporal use
421	of roads by grizzly bears in west-central Alberta. Ursus, 21, 43-56.
122	Hägerstrand, T., 1970. What about people in regional science? Papers of the Regional
123	Science Association, 24, 7–21.
124	Harris, S., Cresswell, W.J., Forde, P.G., Trewhella, W.J., Woollard, T., and Wray, S.,
125	1990. Home range analysis using radio tracking data- a review of problems and
126	techniques as applied to the study of mammals. Mammal Review, 20(2-3), 97-
127	123.
128	Horne, J.S., Garton, E.O., Krone, S.M. and Lewis, J.S., 2007. Analyzing animal
129	movements using brownian bridges. Ecology, 88, 2354–2363.
430	Hunter, A., 2007. Sensor-Based Animal Tracking. PhD Diss., Department of Geomatics
431	Engineering, University of Calgary.

132	Kie, J.G., Matthiopoulos, J., Fieberg, J., Powell, R.A., Cagnacci, F., Mitchell, M.S.,
133	Gaillard, JM. and Moorcroft, P.R., 2010. The home-range concept: are
134	traditional estimators still relevant with modern telemetry technology?
135	Philosophical Transactions of the Royal Society B, 365, 2221–2231.
136	Jankowski, P., Andrienko, N., Andrienko, G., and Kisilevich, S., 2010. Discovering
137	landmark preferences and movement patterns from photo postings. Transactions
138	in GIS, 14(6), 833–852.
139	Johnson, D.S., London, J.M., Lea, M.A., and Durban, J.W., 2008. Continuous-time
140	correlated random walk model for animal telemetry data. Ecology, 89(5), 1208-
141	1215.
142	Langrock, R., King, R., Matthiopoulos, J., Thomas, L., Fortin, D. and Morales, J.M.,
143	2012. Flexible and practical modeling of animal telemetry data: Hidden Markov
144	models and extensions. <i>Ecology</i> , 93(11), 2336–2342.
145	Laube, P. and Purves, R., 2011. How fast is a cow? Cross-scale analysis of movement
146	data. Transactions in GIS, 15, 401–418.
147	Laver, P.N. and Kelly, M. J. 2008. A critical review of home range studies. <i>The Journal</i>
148	of Wildlife Management, 290–298.
149	Long, J.A. and Nelson, T.A., 2012. Time geography and wildlife home range delineation.
450	Journal of Wildlife Management, 76(2), 407–413.
451	Long, J.A. and Nelson, T.A., 2013. A review of quantitative methods for movement data.
452	International Journal of Geographical Information Science, 27(2), 292–318.
453	Mattson, D.J., Knight, R.R., and Blanchard, B.M., 1987. The effects of developments and
154	primary roads on grizzly bear habitat use in Yellowstone National Park,

455	Wyoming. International Conference of Bear Research and Management, 7, 259–
456	273.
457	Morales, J.M., Haydon, D.T., Frair, J., Holsinger, K.E. and Fryxell, J.M., 2004.
458	Extracting more out of relocation data: building movement models as mixtures of
459	random walks. <i>Ecology</i> , 85(9), 2436–2445.
460	Mueller, C., Herrero, S., and Gibeau, M.L., 2004. Distribution of subadult grizzly bears
461	in relation to human development in the Bow River Watershed, Alberta. Ursus,
462	15(1), 35–47.
463	Nielsen, S.E., Herrero, S., Boyce, M.S., Benn, B., Mace, R.D., Gibeau, M.L. and Jevons,
464	S., 2004. Modelling the spatial distribution of human-caused grizzly bear
465	mortalities in the Central Rockies Ecosystem of Canada. Biological Conservation
466	120, 101–113.
467	Orellana, D. and Wachowicz, M., 2011. Exploring patterns of movement suspension in
468	pedestrian mobility. Geographical Analysis, 43, 241–260.
469	Phillips, M.K., 1987. Behavior and habitat use of grizzly bears in northeastern Alaska.
470	International Conference on Bear Research and Management, 7, 159–167.
471	Proctor, M.F., McLellan, B.N., Strobeck, C. and Barclay, R.M.R., 2004. Gender-specific
472	dispersal distances of grizzly bears estimated by genetic analysis. Canadian
473	Journal of Zoology, 82, 1108–1118.
474	Rempel, R.S. Rodgers, A.R. and Abraham, K.F., 1995. Performance of a GPS animal
475	location system under boreal forest canopy. Journal of Wildlife Management, 59,
476	543–551.

477	Rode, K.D., Robbins, C.T. and Shipley, L.A., 2001. Constraints on herbivory by grizzly
478	bears. <i>Oecologia</i> , 128, 62–71.
479	Roever, C.L., Boyce, M.S. and Stenhouse, G.B., 2008. Grizzly bears and forestry II:
480	Grizzly bear habitat selection and conflicts with road placement. Forest Ecology
481	and Management, 256: 1262–1269.
482	Samuel, M.D. and Green, R.E., 1988. A revised test procedure for identifying core areas
483	within the home range. Journal of Animal Ecology, 57(3), 1067–1068.
484	Seaman, D.E., Millspaugh, J.J., Kernohan, B.J., Brundige, G.C., Raedeke, K.J. and
485	Gitzen, R.A., 1999. Effects of sample size on kernel home range estimates. The
486	Journal of Wildlife Management, 739-747.
487	Sawyer, H. and Kauffman, M.J., 2011. Stopover ecology of a migratory ungulate.
488	Journal of Animal Ecology, 80, 1078–1087.
489	Schneider, R.R., 2002. Alternative Futures: Alberta's Boreal Forest at the Crossroads.
490	Edmonton: The Federation of Alberta Naturalists.
491	Smulders, M., Nelson, T.A., Jelinski, D.E., Nielsen, S.E., and Stenhouse, G.B., 2010. A
492	spatially explicit method for evaluating accuracy of species distribution models.
493	Diversity and Distributions, 16(6), 996–1008.
494	Smulders, M., Nelson, T.A., Jelinski, D.E., Nielsen, S.E., Stenhouse, G.B., and Laberee,
495	K., 2012. Quantifying spatial-temporal patterns in wildlife ranges using STAMP:
496	A grizzly bear example. <i>Applied Geography</i> , 35(1), 124–131.
497	Stenhouse, G.B. and Munro, R.H.M. 2000. Foothills Model Forest Grizzly Bear Research
498	Program — 1999 Annual Report. Hinton: Foothills Model Forest.

499	Stewart, B.P., Nelson, T.A., Laberee, K., Nielsen, S.E., Wulder, M.A., and Stenhouse,
500	G.B., 2013. Quantifying grizzly bear selection of natural and anthropogenic
501	edges. Journal of Wildlife Management, doi: 10.1002/jwmg.535.
502	Thériault, M., Claramunt, C. and Villeneuve, P., 1999, A spatio-temporal taxonomy for
503	the representation of spatial set behaviours, in Spatio-temporal Database
504	Management, Springer-Verlag, LNCS 1678, pp. 1–19.
505	Tremblay, Y., Shaffer, S.A., Fowler, S.L., Kuhn, C.E., McDonald, B.I., Weise, M.J.,
506	Bost, CA., et al. (2006). Interpolation of animal tracking data in a fluid
507	environment. Journal of Experimental Biology, 209(1), 128–140.
508	Turchin, P., 1986. Modelling the effect of host patch size on Mexican bean beetle
509	emigration, <i>Ecology</i> , 67(1), 124–132.
510	White, J.C., Wulder, M.A., Gomez, C. and Stenhouse, G.B., 2011. A history of habitat
511	dynamics: characterising 35 years of stand replacing disturbance. Canadian
512	Journal of Remote Sensing, 37(2), 234–25.
513	Wiens, J.A. 1989. Spatial scaling in ecology. Functional ecology, 3, 385-397.
514	Worton, B.J., 1987. A review of models of home range for animal movement. <i>Ecological</i>
515	Modelling, 38(3), 277–298.
516	

Road Density (km/km²)				P-value	Ν	
	in PP	in PPA HR in SMA				
	Mean	CoV	Mean	CoV		
adult female	0.59	0.24	0.60	0.36	0.7401	53
subadult female	0.50	0.19	0.66	0.42	0.0209	15
adult male	0.53	0.43	0.43	0.70	0.8373	33
subadult male	0.60	0.28	0.50	0.63	0.2732	20
female	0.57	0.24	0.63	0.22	0.1503	68
male	0.52	0.39	0.46	0.30	0.4634	53

Table 1. A comparison between the road density in the SMAs and the HR. Statistical significance was determined using a Mann Whitney U test. Subadult females were found to have significantly different road density in their SMAs compared to their HRs. N represents the sum of individual bears by season for each year of the study (2005-2010).

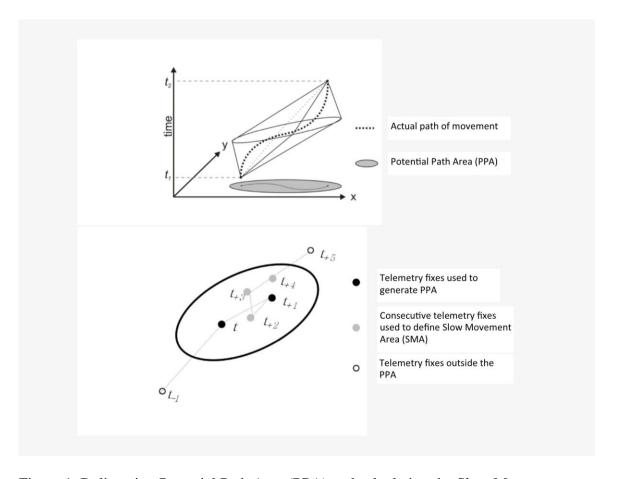


Figure 1. Delineating Potential Path Area (PPA) and calculating the Slow Movement Area (SMA). The upper panel shows how the space-time prism contains all sets of accessible locations given two telmetry fixes, t_1 and t_2 . By combining several PPA ellipses a PPA home range is defined (see Long and Nelson 2012). In the lower panel, all consecutive telemetry locations within a PPA ellipse are counted in the calculation of the SMA. The PPA ellipse containing the largest number of consecutive telemetry fixes is used as the basis for the SMA.

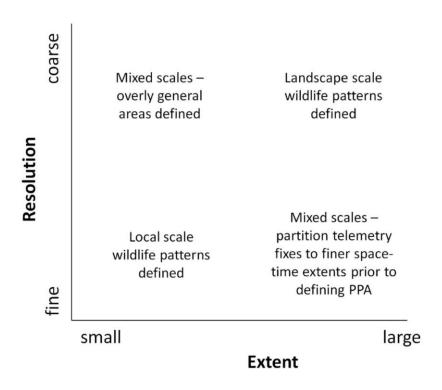


Figure 2. Data resolutions and extents most appropriate for use with PPA home range and SMA delineation do not mix scale. Grey areas indicate appropriate combinations of data resolutions and extents for applying SMA delineation. When the scales are mixed the SMA defined will be overly general and likely too large (upper left) or so small relative to the space-time extent that it is not useful.

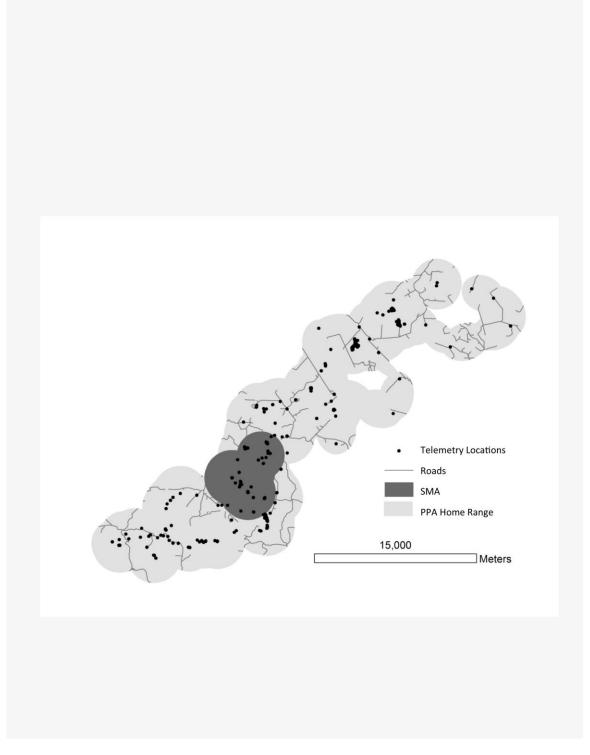


Figure 3. Defining the PPA home range and SMA for one male bear.

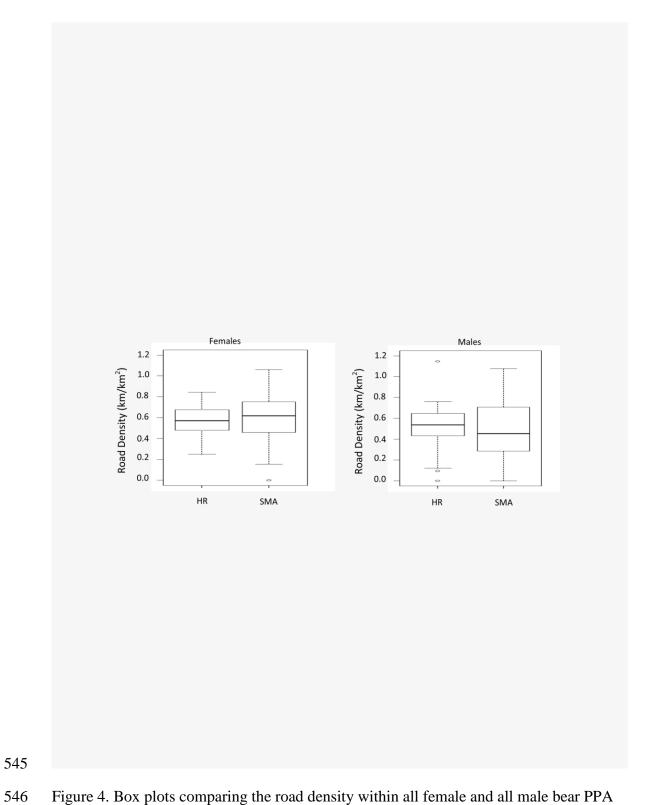


Figure 4. Box plots comparing the road density within all female and all male bear PPA home ranges to their SMAs.

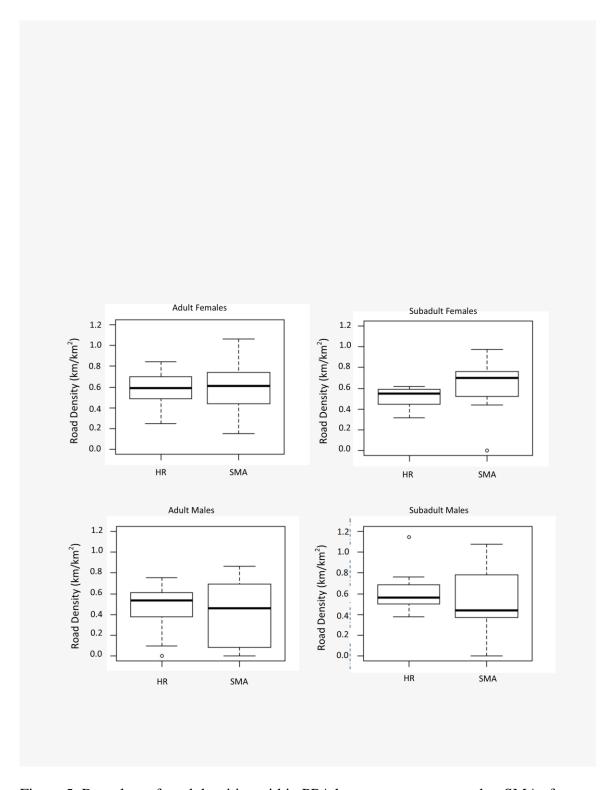


Figure 5. Box plots of road densities within PPA home ranges compared to SMAs for adult females, subadult females, adult males and subadult males.